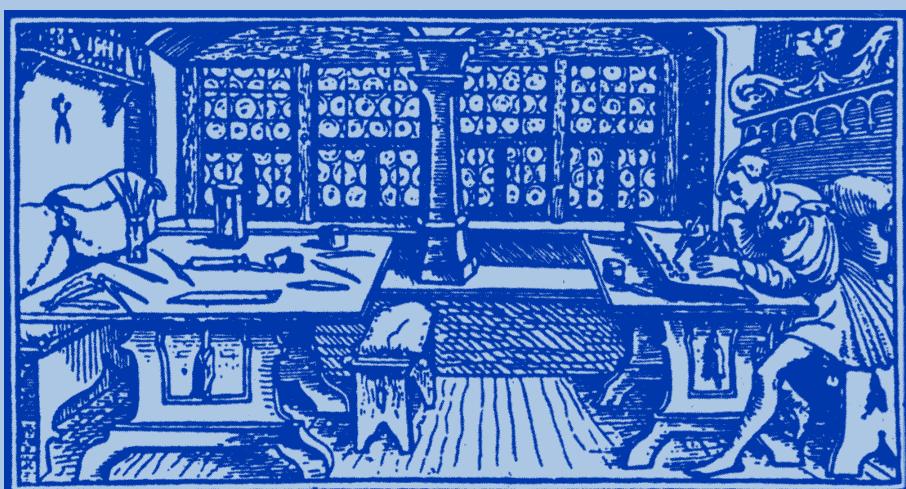


STUDIA

UNIVERSITATIS
BABEŞ-BOLYAI

i a
g o i o i
B i o i
C L U J - N A P O C A 2 0 0 4



STUDIA
UNIVERSITATIS BABEŞ-BOLYAI
BIOLOGIA
1

EDITORIAL OFFICE: Republicii no. 24, 400015 Cluj-Napoca, Romania. Phone: 0264-40.53.52

SUMAR – CONTENTS – SOMMAIRE – INHALT

L. RÁKOSY-TICAN, L. MOMEU, F. LŐRINCZI, Data on the Microbiota in the Vadu Crișului Cave.....	3
P. -V. BELDEAN, Analiza faunistică, ecologică și zoogeografică a suprafamiliei <i>Pentatomoidea</i> (<i>Insecta, Heteroptera</i>) din sectorul inferior al bazinului râului Arieş * Faunistical, Ecological and Zoogeographic Analysis of the Suprafamily <i>Pentatomoidea</i> (<i>Insecta, Heteroptera</i>) from the Inferior Area of the Arieş River Basin	9
A. DAVID, I. COROIU, Breeding Biology of the Golden Eagle (<i>Aquila chrysaetos chrysaetos</i> L. 1758) in the Gorge of Turda Nature Reserve	23
S. ONAC, V. CRISTEA, M. KEUL, D. BÁTHORY, M. TRIFU, Peroxidase Activity, Stomatal Conductance and Carbohydrate Metabolism in Soybean Plants under Heavy Metal Stress	37
D. PODAR, C. DOBROTĂ, M. TRIFU, Uptake of Heavy Metals by Maize (<i>Zea mays</i>) Plants Cultivated on Mine Spoils.....	47
A. FLOREA, C. CRĂCIUN, C. PUICĂ, Aspecte histologice și ultrastructurale ale timusului în intoxicație acută cu venin de albine * Histological and Ultrastructural Aspects of the Thymus in the Acute Intoxication with Honey Bee Venom	61
C. PAŞCA, V.-D. SANDU, Aspecte histopatologice ale toxicității pancreatică a unui agent alchilant antitumoral (Cisplatin) la șobolanii albi Wistar de vârste diferite * Histopathological Aspects of the Toxicity of an Antitumour Alkylating Agent (Cisplatin) on the Pancreas of White Wistar Rats of Different Ages.....	71

C. TARBA, F. SUĂRĂŞAN, Functional Correlations between the Permeability Transition, Membrane Potential Collapse and Calcium Release by Liver Mitochondria of Ethanol-Fed Rats	83
C. TARBA, F. SUĂRĂŞAN, Effect of Acute Endotoxin Administration to Rats on Membrane Permeability and Related Functional Parameters of Liver Mitochondria.....	105
M. DRĂGAN-BULARDA, C. E. GRIGORE, D. ȚURA, Utilizarea indicatorului enzimatic al calității nămolului în scopul valorificării și protecției lacurilor saline * Utilisation of the Enzymatic Indicator of Mud Quality for Exploitation and Protection of Salt Lakes	129
L. OPREAN, Contribuții la studierea drojdiilor industriale *Contributions to the Study of Industrial Yeasts.....	141
S. KISS, M. DRĂGAN-BULARDA, D. PAŞCA, Additional Bibliography on Enzymology of Disturbed Soils	149

IN MEMORIAM

Prof. Univ. dr. Traian Ceuca (1921-2003) (D. CEUCA, I. COROIU).....	169
--	------------

RECENZII - BOOK REVIEWS - COMPTES RENDUS - BUCHBESPRECHUNGEN

Universität Trier. Abteilung Bodenkunde, Skript zur Rekultivierung (M. DRĂGAN-BULARDA, S. KISS).....	171
T. Szili-Kovács (Editor), Proceedings of the Workshop on Managing Soil Quality - Using Microbial Resources (D. PAŞCA, S. KISS, M. DRĂGAN-BULARDA).....	172
I. Kiss, Az erodált talajok enzimológiaja (K. BARTÓK).....	173
D. Bradshaw, Vertebrate Ecophysiology. An Introduction to Its Principles and Applications (P. GHERGHEL)	174

DATA ON THE MICROBIOTA IN THE VADU CRIŞULUI CAVE

LENUȚA RÁKOSY-TICAN*, LAURA MOMEU**
and FRANCISC LŐRINCZI***

SUMMARY. - The Vadu Crișului cave, located in the Pădurea Craiului Mountains (Romanian Western Carpathians) intensely visited by tourists, has been investigated from a microfloristic point of view. After a brief topographic and topo-climatic presentation of the cave, the paper provides a taxonomic approach to the troglophil and troglobiontic micromycetes and algae occurring in the cave. The investigation revealed the presence of 6 strains of micromycetes, a bacterial strain, 5 blue-green algae (*Cyanoprokaryota*), 4 taxa of xanthophytes (*Xanthophyta*) and 12 green algae (*Chlorophyta*) in the soil, on calcareous cave walls, in condensed water and limestone pool water.

Located at an altitude of 304 m a.s.l., on the left side of the Crișul Repede River, between the localities Şuncuiuş and Vadu Crișului, the Vadu Crișului cave is the main element of the “Defileul Crișului Repede” (Crișul Repede Strait) Nature Reserve and one of the most important habitats of the underground fauna in the Apuseni Mountains. The cave gallery has formed along a diaclase and it is penetrated by an underground water flow, except for an upper cavity, *i.e.* the balcony hall, which is the only part exhibiting outstanding concretions (Fig. 1). The first siphon occurs at 655 m from the entrance and may be penetrated only during dry periods, while the second siphon located at the end of the cave is always closed.

Although the cave is 1000 m long, it is not subjected to external meteorological influences except for the first 100 m. Racoțiu [10] described the cave as a cavity exhibiting permanent bidirectional ventilation. This topo-climatic type is in perfect agreement with the general topography of the cave, which is a practically horizontal one-entrance cavity. This sole opening concentrating all the aerodynamic exchanges with the outer world is the site of two air currents of opposite directions. The lower part of the gallery is affected by an ascending current during winter and a descending one during summer. The limits in time of the two ventilation types correspond to October and May, so that winter lasts for seven months and summer for five.

* Babeş-Bolyai University, Department of Experimental Biology, 400006 Cluj, Romania.

E-mail: lrakosy@hasdeu.ubbcluj.ro

** Babeş-Bolyai University, Department of Taxonomy and Ecology, 400006 Cluj, Romania

*** Deceased

The annual ventilation-dependent variations in air temperature have two different characteristics: at floor level the thermic values are directly influenced by outer temperatures only during winter, while in summertime the mean thermic value remains constantly around 10.2° C. Annual relative humidity ranges between 98 and 100% [10].

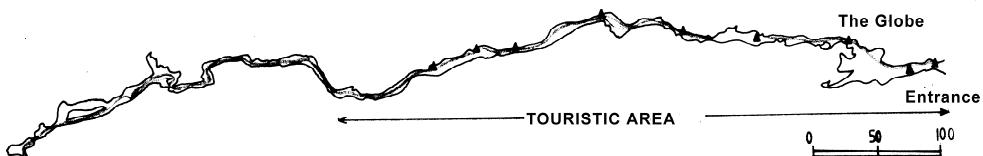


Fig. 1. *The map of the Vadu Crișului cave.*
The sampling sites are indicated by ▲.

Ventilation speed varies according to season; yet the speed of ascending winter current resembles that of the descending summer one. Evapocondensation is strictly dependent upon ventilation: condensation typically occurs during summer, while the winter current change brings about an increasing evaporation intensity with ventilation speed.

Light intensity in the lower half of the gallery drops sharply along the first 20 m from 3200 lux (outside) to 0.1 lux (October 31, 1980).

Few data have been reported as regarding microbiota of the Vadu Crișului cave [5]; consequently, the aim of our study was to investigate from a taxonomic point of view the trogophil and troglobiontic micromycetes and algae occurring in the cave.

Materials and methods. The first studies were carried out in the cave (October 31, 1980) and comprised temperature, water pH and light intensity measurements (by means of field thermometer, pH-paper and luxmeter) and underground microscopic investigation (portable light microscope). As the resulting data were not conclusive, due to poor magnifying possibilities and scarce occurrence of organisms, samples were collected for further laboratory investigation (December 8, 1980) either in sterile test tubes or in tubes with culture media. The microclimatic measurements were repeated.

Micromycetes were isolated from samples obtained from cave sediments, limestone pool water and calcareous walls (pelicles, crusts or spots displaying different colours and consistency against the bearing rocks) and preserved in sterile test tubes. Sterility of tubes was achieved by maintenance in drying oven at 105° C for 48 h. Czapek medium solidified with agar was used for inoculation. Incubation took place in laboratory lockers in the absence of light at room temperature (19-22° C). The cultures were macro- and microscopically examined after 14 and 28 days of incubation.

Algae were collected from limestone pool water samples and the underground water flow, and subsequently transferred in sterile test tubes. These samples were immediately fixed in 4% formol. Other samples collected from limestone pools, soil and calcareous walls were inoculated on mineral culture media, such as Knop-Pringsheim (KP), Benecke (BK) or Watanabe (W) [9]. Incubation was carried out in light at room temperature (19-22° C) for about 3 months, until colonies were clearly visible. The taxonomic identification of algae was carried out by light microscopy.

Results and discussion. The topoclimatic measurements were similar to those obtained by R a c o v i t ă [10]: the recorded temperature by the end of October was 10° C at the entrance, subjected to relative variations, and 11° C in the stable meroclimatic part. As mentioned before, light intensity drops below 1 lux after the first 20 m. The underground water has been found to be slightly acid (pH=5) without any variations.

Table 1
Micromycetes and bacteria isolated from the Vadu Crişului cave

Isolated strains	Sampling sites
<i>Cladosporium herbarum</i> (Person) Link	Aphotic area, calcareous wall
<i>Scopulariopsis brevicaulis</i> Bainier var. <i>glabra</i> Thom	Aphotic area, condensed water
<i>Torula</i> sp.	Aphotic area, calcareous wall
<i>Verticillium lateritium</i> Berkeley	Aphotic area, calcareous wall
<i>Verticillium lateritium</i> Berkeley, white mutant	Aphotic area, calcareous wall
<i>Verticillium terrestris</i> (Link) Lindau	Aphotic area, condensed water
Unidentified bacterium	Aphotic area, calcareous wall

The taxonomic data reveal the presence of bacteria, micromycetes and algae in the cave microflora. The micromycetes (Table 1) have been assigned to the *Fungi imperfecti* and most of them were previously isolated from other caves of the Crişul Repede Strait [4, 6, 11]. Most of the taxa isolated from the Vadu Crişului cave also occur in the soils of the Apuseni Mountains [6, 7], thus revealing their origin and high adaptability. While this is the first account of *Torula* sp. in the Apuseni Mountains, the species *Scopulariopsis brevicaulis* Bainier has been recorded in all the caves studied in the area [5]. Noteworthy is the high morphological variability of the micromycete strains which makes their identification extremely difficult. For instance, *Verticillium lateritium* Berkeley, usually developing brownish colonies, produced a white mutant under laboratory conditions. All micromycete and bacterium strains were isolated from the aphotic part of the cave (Table 1).

The water samples collected in the sterile test tubes and fixed with formol were found to contain only the resistant forms. An unidentified viable blue-green alga also occurred. It seems [3] that formol is not suitable for fixing underground samples, as it easily destroys the few forms present in this biotope.

Table 2
Algal species isolated from the Vadu Crișului cave

Algal taxa	Sampling sites	Culture media*
CYANOPROKARYOTA		
<i>Aphanethece saxicola</i> Naegeli	□ ▲ ▼ • □ ○	KP, BK, W
<i>Gloeocapsa montana</i> (Kützing) Hollerbach	□ ▼ •	KP, BK
<i>Gloeocapsa</i> sp.	△ •	KP
<i>Nostoc</i> sp.	△ ▲ • □ ○	KP, W
<i>Symploca</i> sp.	□ ▲ ▼ • ○	KP, BK
XANTHOPHYTA		
<i>Monodus dactylococcoides</i> Pascher	□ ▲ • □	BK
<i>Monodus subterranea</i> Boye-Petersen	△ □ ▲ • □	KP, BK
<i>Monodus</i> sp.	▲ ○	KP
<i>Pleurochloris</i> sp.	□ •	BK
CHLOROPHYTA		
<i>Chlorella vulgaris</i> Beijerinck	□ ▲ ▼ • □ ○	KP, BK
<i>Chlorella luteo-viridis</i> Chodat	△ •	KP
<i>Chlorella zofingiensis</i> Dönz	□ ▲ • □	KP, BK
<i>Chlorococcum</i> sp.	△ □ ▲ ▼ • □	KP, BK, W
<i>Chlamydomonas rosae</i> H. et O. Ettl	△ ▼ •	KP
<i>Chlamydomonas</i> sp.	□ ▲ ▼ • □	KP, BK, W
<i>Gloeotila</i> sp. Kützing	□ •	BK
<i>Gloeococcus mucosus</i> Braun	▲ □	W
<i>Hormidium flaccidum</i> Braun	▲ ○	KP
<i>Stichococcus bacillaris</i> Naegeli	△ □ ▲ ▼ • □ ○	KP, BK, W
<i>Stichococcus minor</i> Naegeli	▲ ▼ • □	KP
<i>Tetraecystis dissociata</i> Brown et Bold	△ •	KP

▲ - Calcareous wall. □ - Soil. ▼ - Lignified plant. △ - Water from cave pool.

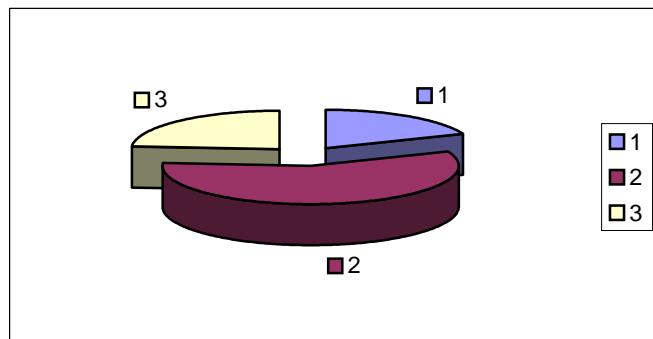
● - Aphotic area. ▨ - Entrance area (low light). ○ - Photic area (external area).

* For the composition of culture media see [9].

The algae living in the cave (Table 2) were identified after incubating the sample on mineral media in the light. Being highly resistant, the algae can survive latently in a cave and recover under adequate conditions (e.g. in light). Such phenomenon also occurs in electrified caves and leads to the formation of the so called „lamp flora“. This was the case in the Vadu Crișului cave too, where electricity was installed for a certain period of time.

The algal forms isolated are very similar to those recorded most often in underground environment [2, 8]. In the Vadu Crișului cave the most frequent ones are the green algae (12 taxa) as compared to the blue-green ones (*Cyanoprokaryota* – 5 taxa) (Fig. 2). The former ones are supposed to exist latently in the aphotic area, considering their presence in different cave microbiotopes as well as in the entrance

area and outside the cave. Of the 7 outer species isolated from the calcareous wall above the entrance, only two could not be identified inside the cave; all the others were found in different parts of the aphotic area suggesting a possible external source and dissemination through the cave *e.g.* water flux (Table 2).



F i g. 2. *The proportion of different algal groups in the Vadu Crișului cave.*
1 – *Xanthophyta*. 2 – *Chlorophyta*. 3 – *Cyanoprokaryota*.

Other 10 taxa were identified only in the aphotic area. These species are usually present in restricted microbiotopes, being sometimes isolated from a single site. This account for their ability to live in an underground environment, either by heterotrophy or by other mechanisms. Of course, these suppositions require experimental support.

Conclusions. The microbiota, which is considered to play a leading role in the caves [1], should be thoroughly investigated for a better understanding of the underground hollow genesis and ecology. Therefore, knowledge of the microbiota in the Vadu Crișului cave should also be completed by further investigations.

R E F E R E N C E S

1. Co m a n, D., *Essai sur une interprétation écologique de l'origine des grottes*, „Trav. Inst. Spéol. «E. Racovitza»”, **18**, 1979, 191-200.
2. D r a g a n o v, S. I., D i m i t r o v a – B u r i n, E. A., *Speleoalgorologicheskie issledovaniya v Bolgarii*, „Int. Union Speleol.” (Praha), **5**, 1977, 11-17.
3. H a j d ú, L., *Algological studies in the cave of Mátyás Mount, Budapest, Hungary*, „Int. J. Speleol.”, **2** (1-2), 1966, 137-150.
4. L ó r i n c z i, F., B e n k e, P., *Noi contribuții la cunoașterea micromicetelor din peșterile de la Rezervația Naturală „Defileul Crișului Repede”*, „Contrib. Bot.” (Cluj-Napoca), 1979, 31-36.

5. Lőrinczi, F., Ticăni, L., *Contribuții la cunoașterea micromicetelor din peșteri din Munții Apuseni*, „Contrib. Bot.” (Cluj-Napoca), 1983, 29-32.
6. Lőrinczi, F., Turec, L., *Micromicete izolate din sol și peșteri. Flora și vegetația rezervației naturale Defileul Crișului Repede*, „Contrib. Bot.” (Cluj), 1966, 69-82.
7. Lőrinczi, F., Turec, L., *Contributions to the study of soil micromycetes in the natural reservation „The Defile of the Crișul Repede”*, „Symp. Soil Biol.” (Cluj, 1966), 1966, 203-211.
8. Mason-Williams, M. A., Benson-Evans, K., *Summary of results obtained during a preliminary investigation into bacteria and botanical flora of caves in South Wales*, „Int. J. Speleol.”, 2 (4), 1967, 397-402.
9. Petrif, Ş., Nagy-Tóth, F., *Cultura algelor. Microalge dulcicole*, in Petrif, Ş., Ionescu, A. (Eds.), *Tratat de algologie*, Vol. I, pp. 453-472, Ed. Acad. Române, București, 1976.
10. Raciotești, G., *Étude écologique sur les coleoptères bathysciinae cavernicoles*, „Mém. Biospéol.”, 6, 1980, 1-200.
11. Szász, E., Benkő, P., *Cercetări microfloristice la gura peșterilor din Defileul Crișului Repede*, „Contrib. Bot.” (Cluj-Napoca), 1979, 21-36.
12. Van Landingham, L. S., *A new species of Gomphonema (Bacillariophyta) from Mammoth Cave, Kentucky*, „Int. J. Speleol.”, 2 (4), 1967, 405-406.

ANALIZA FAUNISTICĂ, ECOLOGICĂ ȘI ZOOGEOGRAFICĂ A
SUPRAFAMILIEI *PENTATOMOIDEA* (*INSECTA, HETEROPTERA*)
DIN SECTORUL INFERIOR AL BAZINULUI RÂULUI ARIEŞ

PAUL-VASILE BELDEAN*

SUMMARY. - Faunistic, Ecological and Zoogeographic Analysis of the Suprafamily Pentatomoidae (Insecta, Heteroptera) from the Inferior Area of the Arieş River Basin. The biological material was collected during 2001, 2002 and 2003. In the studied area we have revealed a number of 565 *Pentatomoidae*, belonging to 26 species, 19 genera and 4 family. The majority of the species (20) belongs to the family *Pentatomidae*. The researches were made in nine different types of ecosystems: mixed forest (beech and hornbeam), forest of hornbeam, cluster of osier, hay field with high humidity, hay field with medium humidity, rocky region, grassland, crop of alfalfa and roadside vegetation. Most species (17) live in the hay field with medium humidity. In the mixed forest 10 species of *Pentatomoidae* live (*Dolycoris baccarum*, *Pentatoma rufipes* and *Palomena prasina* have the highest values of numerical and relative abundance). In the hay field, the highest values of abundance were shown by *Eurydema oleraceum*, *Coptosoma scutellatum* and *Aelia acuminata*. In the crop of alfalfa *Piezodorus lituratus* and in the roadside vegetation *Coptosoma scutellatum* and *Graphosoma lineatum* are the most abundant species. The highest number of adults was recorded in June. A proportion of 69% of the species is polyphagous. From all species 50% belong to the Palearctic Region. *Sciocoris (Sciocoris) deltocephalus* is a new record for Transilvania.

Publicațiile anterioare referitoare la fauna de *Pentatomoidae* din Transilvania au caracter faunistic, cuprinzând liste de specii din diferite zone [3, 5, 9, 11]. Unele lucrări prezintă speciile de *Pentatomoidae* considerate dăunătoare în diverse culturi agricole, în acest caz apărând și informații sumare legate de mediul de viață al acestora [6, 7, 10].

Nu se cunosc publicații anterioare referitoare la întreg sectorul inferior al Arieșului, semnalări ale cătorva specii din această zonă fiind cuprinse în volumul VIII din Fauna R.S.R. [4]. În perioada 1974-1995 au fost colectate din Cheile Turenilor 13 specii de *Pentatomoidae* aparținând la trei familii (*Cydnidae*, *Scutelleridae* și *Pentatomidae*), datele nefiind însă publicate [14].

* Universitatea "Babeş-Bolyai", Catedra de Taxonomie și Ecologie, Cluj-Napoca, 400006, România.
E-mail: pbeldean@hasdeu.ubbcluj.ro

Material și metode. Materialul biologic analizat a fost colectat pe parcursul anilor 2001, 2002 și 2003 în diferite perioade ale anului. Au fost colectate 4 probe din fiecare ecosistem prin metode specifice tipului de vegetație existent: cu fileul entomologic din vegetația ierboasă (50 cosiri/probă) și prin scuturarea coronamentului arborilor și arbustilor în plasa umbrelă. Pentru fiecare probă a fost notată data și punctul de colectare. Materialul colectat a fost omorât în alcool 70% și păstrat în alcool cu aceeași concentrație în tuburi de plastic, separat pe probe.

În sectorul inferior al Bazinului Arieșului au fost studiate următoarele tipuri de ecosisteme:

I. Făgeto-căpinet în stațiunea Moldovenești situat pe malul stâng al Arieșului, cu expoziție sud-estică, între 350-400 m altitudine. Alături de *Carpinus betulus* și *Fagus sylvatica*, se mai întâlnesc: *Acer campestre*, *Corylus avellana*, *Betula pendula*. Stratul ierbos este bine dezvoltat, mai ales de-a lungul pârâului care trece prin pădure, fiind frecvente: *Mentha*, *Medicago*, *Trifolium pratense*, *Urtica dioica*, *Equisetum arvense*, *Carduus spinosus* și *Carduus echinatus*.

II. Căpinet situat pe versantul drept al Cheilor Turenilor, cu expoziție nordică, între 550-600 m altitudine, în care alături de carpen (*Carpinus betulus*) este frecvent alunul (*Corylus avellana*).

III. Salicete în stațiunile Luna și Gura Arieșului. Sunt formate predominant din specii de salcie (*Salix* sp.), la care se adaugă plopi și anini (*Populus* sp. și *Alnus* sp.). Stratul ierbos este format din *Ononis* sp., *Centaurea* sp., *Typha* sp., *Urtica dioica*, *Coronilla varia*, *Plantago lanceolata* etc.

IV. Fânațe higrofile în Luna și Gura Arieșului.

Fânațul studiat la Luna este situat lângă zăvoi, pe malul drept al Arieșului, la 300 m altitudine. Alături de *Agrostis* și *Festuca* se întâlnesc *Trifolium pratense*, *Lotus corniculatus*, *Coronilla varia*, *Ononis spinosa* și *Plantago lanceolata*. Există și portiuni cu *Sambucus ebulus*.

Fânațul de la Gura Arieșului este situat pe malul drept al Arieșului, în zonă de luncă, la 200 m altitudine. Este puternic higrofil, alături de poacee fiind bine reprezentate speciile de: *Juncus*, *Typha*, *Trifolium*, *Ononis*, *Linaria*, *Cirsium* și *Polygonum*. Se întâlnește și *Caltha palustris*.

V. Fânațe mezofile în Moldovenești și Mărtinești. Sunt formate predominant din specii de *Agrostis* (*Agrostis tenuis*, *Agrostis alba* etc.) și *Festuca* (*Festuca rubra*, *Festuca pratensis* etc.) alături de care, la Moldovenești se întâlnesc *Trifolium pratense*, *Lotus corniculatus*, *Coronilla varia*, *Ononis spinosa* și *Plantago lanceolata*. La Mărtinești stratul ierbos este format predominant din *Trifolium pratense*, *Trifolium repens*, *Coronilla varia*, *Medicago* sp. și *Cirsium* sp.

VI. Stâncării în Cheile Turenilor. Vegetația acestor stâncării se încadreză în asociațiile *Seslerietum rigidae* și *Spireetum ulmifolie* [1, 8].

VII. Pășuni în Moldovenești, Mărtinești, Cheile Turenilor, Luna și Gura Arieșului. Predomină speciile de *Festuca* și *Agrostis*. Pășunea din Moldovenești este higrofilă, bogată în vegetație ierboasă, fiind întâlnite specii de *Trifolium*, *Linaria* și *Polygonum*.

În Mărtinești pășunea studiată are expoziție sud-vestică, fiind predominant *Festuca sulcata* și *Agrostis tenuis*. Sporadic se întâlnesc tufe de *Prunus spinosa*, *Crataegus monogyna* și *Rosa canina*.

În Cheile Turenilor a fost studiată o pășune puternic xerofilă și intens pășunată, fiind încadrată în asociația *Festucetum sulcatae calcophilum* [1,8].

În pășunea din Luna, situată în lunca Arieșului și cu un grad relativ scăzut de exploatare, alături de poacee sunt întâlnite *Trifolium pratense* și *Althea rosea*, precum și specii de *Ononis*, *Medicago*, *Euphorbia* și *Artemisia*.

La Gura Arieșului pășunea studiată este intens exploataată, situată pe malul drept al Arieșului, fiind reprezentată în general de poacee și specii de *Medicago*, *Euphorbia* și *Artemisia*. Izolat se întâlnesc tufe de *Crataegus monogyna*, *Syringa vulgaris*, *Prunus spinosa* și *Rosa canina*.

VIII. Cultură de lucernă situată în stațiunea Moldovenești, înainte de intrarea în localitate, la aproximativ 350 m altitudine.

IX. Vegetație ruderălă în 4 stațiuni: Moldovenești, Mărtinești, Luna și Gura Arieșului. Este formată predominant din *Artemisia* sp., *Matricaria* sp., *Trifolium* sp., *Linaria* sp., *Verbascum* sp., *Raphanus* sp., *Echium vulgare*, *Coronilla varia*, *Equisetum arvense*, *Urtica dioica*, *Lycopus europaeus*, *Capsella bursa-pastoris*, *Mentha longifolia*, *Carduus nutans* și *Carduus acanthoides*.

În stațiunea Moldovenești au fost efectuate colectări lunare din mai până în octombrie, în vederea efectuării studiilor de dinamică a populațiilor de Pentatomoidea.

Determinările au fost efectuate în laborator cu ajutorul stereo-microscopului, folosind diverse surse bibliografice [4, 12, 13]. Materialul determinat a fost ordonat pe familii, subfamilii și specii conform sistemului taxonomic actual [2, 4, 12].

Rezultate și discuții. Analiza taxonomică și faunistică. În urma studiilor efectuate în sectorul inferior al Arieșului am colectat 565 indivizi de *Pentatomidae*, aparținând la 26 de specii, 19 genuri și 4 familii. Atât din punct de vedere al numărului de specii cât și al numărului de indivizi, cel mai bine reprezentată este familia *Pentatomidae* (481 de indivizi din 20 de specii). Din familia *Scutelleridae* au fost colectați 36 de indivizi din 3 specii, familia *Plataspidae* este prezentă printr-o singură specie cu 45 de indivizi, iar familia *Cydnidae* este reprezentată prin 2 specii cu 3 indivizi (Tabel 1).

La nivelul teritoriului României, din suprafamilia *Pentatomidae* cel mai bine reprezentată este familia *Pentatomidae*, urmată de familiile *Cydnidae*, *Scutelleridae*, *Acanthosomidae* (prezentă în fauna României, dar nesemnalată în sectorul inferior al Arieșului) și *Plataspidae* [4].

Ambele specii de *Cydnidae* colectate sunt încadrate în subfamilia *Sehirinae*. Cele trei specii ale familiei *Scutelleridae* sunt distribuite în două subfamilii: *Odontotarsinae* cu o singură specie și *Eurygasterinae* cu 2 specii. Din familia *Pentatomidae* cel mai bine reprezentată este subfamilia *Pentatominae* cu 16 specii, în timp ce subfamiliile *Podopinae* și *Amyotinae* sunt prezente fiecare prin câte 2 specii.

Tabel 1

**Speciile de Pentatomoidă colectate din
sectorul inferior al Bazinului Arieșului**

Familia, subfamilia, specia	Numărul de indivizi			Stațiuni cercetate					NTr. S.
	T	♂	♀	1	2	3	4	5	
Familia Plataspidae	45	19	26	23	9		8	5	
<i>Coptosoma scutellatum</i> (Geoffroy, 1785)	45	19	26	23	9		8	5	
Familia Cydnidae	3	1	2	1			2		
<i>Subfam. Sehirinae</i>	3	1	2	1			2		
<i>Tritomegas bicolor</i> (Linné, 1758)	2	1	1	1			1		
<i>Tritomegas sexmaculatus</i> (Rambur, 1842)	1		1				1		
Familia Scutelleridae	36	18	18	25	7	4			
<i>Subfam. Odontotarsinae</i>	1		1				1		
<i>Odontotarsus purpureolineatus</i> (Rossi, 1790)	1		1				1		
<i>Subfam. Eurygasterinae</i>	35	18	17	25	7	3			
<i>Eurygaster maura</i> (Linné, 1758)	21	13	8	13	5	3			
<i>Eurygaster testudinaria</i> (Geoffroy, 1785)	14	5	9	12	2				
Familia Pentatomidae	481	231	250	251	52	56	105	17	
<i>Subfam. Podopinae</i>	25	13	12	12	5	4	1	3	
<i>Graphosoma lineatum</i> (Linné, 1758)	24	12	12	12	5	4		3	
<i>Podops inuncta</i> (Fabricius, 1775)	1	1					1		S.
<i>Subfam. Pentatominae</i>	432	206	226	227	37	52	102	14	
<i>Sciocoris (Sciocoris) deltocephalus</i> Fieber, 1861	1		1	1					NTr.
<i>Sciocoris (Aposciocoris) microphthalmus</i> Flor, 1860	2	1	1		2				
<i>Aelia acuminata</i> (Linné, 1758)	54	28	26	28	9	7	8	2	
<i>Aelia rostrata</i> Boheman, 1852	2		2		2				
<i>Eusarcoris aeneus</i> (Scopoli, 1763)	11	6	5	8			3		
<i>Stagnomus (Dalleria) pusillus</i> (Herrich-Schäffer, 1830)	3	1	2		3				S.
<i>Holcostethus (Holcostethus) vernalis</i> (Wolff, 1904)	31	15	16	16	2	10		3	
<i>Carpocoris purpureipennis</i> (De Geer, 1773)	22	10	12	15	1	6			
<i>Carpocoris pudicus</i> (Poda, 1761)	10	1	9	3	1	6			
<i>Dolycoris baccarum</i> (Linné, 1758)	101	49	52	57	11	19	11	3	
<i>Palomena prasina</i> (Linné, 1761)	11	5	6	9		2			
<i>Piezodorus lituratus</i> (Fabricius, 1794)	41	22	19	39			2		
<i>Pentatoma rufipes</i> (Linné, 1758)	12	7	5	12					
<i>Eurydema ventrale</i> Kolenati, 1846	2		2	1			1		S.
<i>Eurydema ornatum</i> (Linné, 1758)	3	1	2	2		1			
<i>Eurydema oleraceum</i> (Linné, 1758)	126	60	66	36	9	4	71	6	
<i>Subfam. Amyotinae</i>	24	12	12	12	10		2		
<i>Picomerus bidens</i> (Linné, 1758)	19	11	8	9	10				
<i>Zicrona coerulea</i> (Linné, 1758)	5	1	4	3			2		
Numărul total de indivizi	565	269	296	300	68	60	115	22	
Numărul total de specii			26		20	13	11	12	6

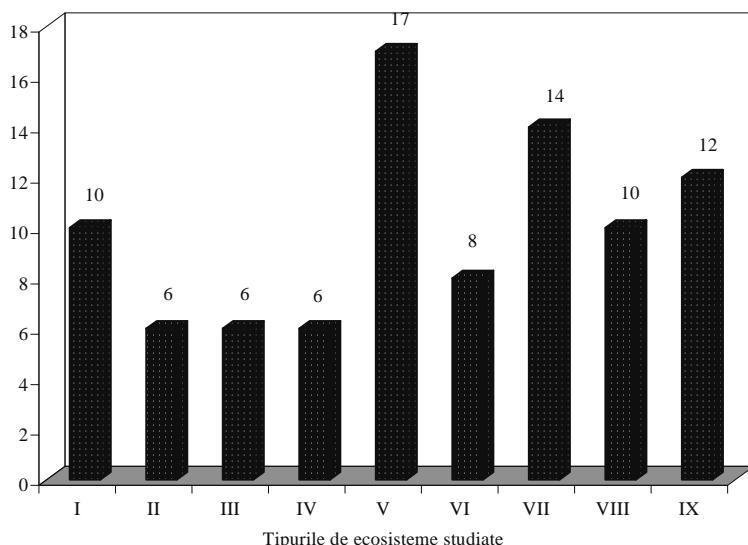
Abrevieri: 1 - Moldoveneşti. 2 - Mărtineşti. 3 - Cheile Turenilor. 4 - Luna. 5 - Gura Arieşului.

NTr. - Specie semnalată pentru prima dată în fauna Transilvaniei. S - Specie sporadică în fauna României.

Sciocoris (Sciocoris) deltocephalus este semnalată de noi pentru prima dată în fauna Transilvaniei, semnalările anterioare fiind din județele Dolj, Giurgiu, Galați, Constanța și Tulcea. Este specie termofilă ponto-mediteraneană, întâlnită pe diferite plante ierboase (*Digitaria*, *Eragrostis*, *Filago*, *Salvia* etc.) în biotopuri nisipoase [4]. În urma studiilor noastre *Sciocoris (Sciocoris) deltocephalus* a fost colectată din stațiunea Moldovenesci, de pe vegetația ierboasă din ecosistemul reprezentat de pășune situată pe malul stâng al Arieșului, pe substrat nisipos.

Podops inuncta, *Stagnomus (Dalleria) pusillus* și *Eurydema ventrale* sunt considerate sporadice în fauna României [4].

Distribuția numărului de specii pe tipuri de ecosisteme. Din făgeto-cărpinetul studiat au fost colectate 10 specii (Tabel 2), fiind semnalate atât specii tipice pentru acest tip de ecosistem (*Palomena prasina*, *Pentatoma rufipes*) cât și specii caracteristice vegetației ierboase (*Eusarcoris aeneus*, *Piezodorus lituratus*). Numărul cel mai mare de specii (17) a fost colectat din ecosistemele mezofile exploataate ca fânaț (Tabel 3), caracterizate printr-un număr mare de specii vegetale ierboase, urmate de pășuni cu 14 specii (Tabel 4), vegetație ruderălă cu 12 specii și cultura de lucernă cu 10 specii (Tabel 5). Din celelalte ecosisteme caracterizate prin vegetație lemnosă (cărpinet, salicete) dar cu un număr de specii vegetale mai redus comparativ cu făgeto-cărpinetul studiat, au fost colectate câte 6 specii de *Pentatomidae*. Tot 6 specii au fost colectate și din fânațele higrofile cu suprafață redusă (aproximativ 500 m² fiecare) situate în vecinătatea culturilor agricole tratate cu ierbicide și insecticide, în timp ce din ecosistemul reprezentat de stâncării au fost colectate 8 specii (Fig. 1).



F i g. 1. Numărul de specii înregistrat în ecosistemele studiate.

Numărul de specii scade odată cu scăderea cantității de vegetație ierboasă, reducerea numărului de plante prezente și creșterea influențelor antropice în ecosistemelor studiate.

Analiza ecologică. În făgeto-cărpinetul din stațiunea Moldovenești (Tabel 2), din punct de vedere al abundenței numerice și relative se remarcă *Dolycoris baccarum* ($A = 0,5$, $Ar = 21,87$) - specie polifagă larg răspândită în România, *Pentatoma rufipes* ($A = 0,43$, $Ar = 18,75$) și *Palomena prasina* ($A = 0,32$, $Ar = 14,06$) - specii caracteristice ecosistemelor de tip forestier. Valorile abundenței numerice și relative înregistrate în cărpinetul din stațiunea Cheile Turenilor sunt asemănătoare pentru majoritatea speciilor colectate, arboretul fiind slab încheiat și prezentând numeroase poteci și luminișuri păsunate. Majoritatea speciilor colectate din salicetele situate în stațiunile Luna și Gura Arieșului prezintă valori reduse ale abundenței numerice și relative.

Tabel 2

Abundența numerică și abundența relativă a speciilor de *Pentatomoidea* colectate din ecosistemele lemonase din sectorul inferior al Bazinului Arieșului

Familia, specia	I			II			III		
	N	A	Ar%	N	A	Ar%	N	A	Ar%
Familia Pentatomidae									
<i>Graphosoma lineatum</i> (Linné, 1758)	6	0,21	9,37	3	0,25	15			
<i>Aelia acuminata</i> (Linné, 1758)							1	0,08	33,33
<i>Eusarcoris aeneus</i> (Scopoli, 1763)	3	0,11	4,68				2	0,16	16,65
<i>Stagnomus (Dalleria) pustium</i> (Herrich-Schäffer, 1830)				3	0,25	15			
<i>Holcostethus (Holcostethus) vernalis</i> (Wolff, 1904)	5	0,17	7,81	5	0,41	25			
<i>Carpocoris purpureipennis</i> (De Geer, 1773)	2	0,07	3,12	4	0,33	20			
<i>Carpocoris pudicus</i> (Poda, 1761)							2	0,16	16,65
<i>Dolycoris baccarum</i> (Linné, 1758)	14	0,5	21,87	3	0,25	15	2	0,16	16,65
<i>Palomena prasina</i> (Linné, 1761)	9	0,32	14,06	2	0,16	10			
<i>Piezodorus lituratus</i> (Fabricius, 1794)	3	0,11	4,68						
<i>Pentatoma rufipes</i> (Linné, 1758)	12	0,43	18,75						
<i>Eurydema oleraceum</i> (Linné, 1758)	5	0,17	7,81				6	0,5	50
<i>Picomerus bidens</i> (Linné, 1758)	5	0,17	7,81				1	0,08	33,33
Numărul total de indivizi/stațiune	64			20			12		3
Numărul total de indivizi/ tip de ecosistem	64			20			15		
Numărul total de specii/stațiune	10			6			4		3
Numărul total de specii/ tip de ecosistem	10			6			6		

Abrevieri: I - Făgeto-cărpinet. II - Cărpinet. III - Salicet. 1 - Moldovenești. 3 - Cheile Turenilor. 4 - Luna. 5 - Gura Arieșului. A - Abundența numerică. Ar - Abundența relativă. N - Numărul de indivizi.

SUPRAFAMILIA PENTATOMOIDEA ÎN SECTORUL INFERIOR AL ARIEȘULUI

În fânațul higrofil din stațiunea Luna, *Eurydema oleraceum* prezintă cea mai mare abundență numerică pentru acest tip de ecosistem (A= 1,75) (Tabel 3). În ambele fânațe higofile studiate (Luna și Gura Arieșului) este bine reprezentată *Coptosoma scutellatum*, singura specie din familia *Plataspidae* semnalată în fauna Transilvaniei, dar foarte frecventă în ecosistemele ierboase.

Tabel 3
Abundența numerică și abundența relativă a speciilor de Pentatomidea colectate din ecosistemele ierboase cu intervenție antropică moderată (fânațe, stâncării) din sectorul inferior al Bazinului Arieșului

Familia, specie	IV						V						VI					
	4		5		1		2		3		4		5		6		7	
	N	A	Ar %	N	A	Ar %	N	A	Ar %	N	A	Ar %	N	A	Ar %	N	A	Ar %
Familia Plataspidae																		
<i>Coptosoma scutellatum</i> (Geoffroy, 1785)	6	0,5	15,38	5	0,62	55,55	9	0,32	8,1									
Familia Cydnidae																		
<i>Tritomegas bicolor</i> (Linné, 1758)							1	0,03	0,9									
Familia Scutelleridae																		
<i>Odontotarsus purpureolineatus</i> (Rossi, 1970)															1	0,08	3,57	
<i>Eurygaster maura</i> (Linné, 1758)							6	0,21	5,4	5	0,31	20	3	0,25	10,71			
<i>Eurygaster testudinaria</i> (Geoffroy, 1785)							7	0,25	6,3	1	0,06	4						
Familia Pentatomidae																		
<i>Graphosoma lineatum</i> (Linné, 1758)							6	0,21	5,4									
<i>Sciocoris (Aposciocoris) microphthalmus</i> Flor, 1860														2	0,12	8		
<i>Aelia acuminata</i> (Linné, 1758)	4	0,33	10,25	1	0,12	11,11	22	0,78	19,82	7	0,43	28	6	0,5	21,42			
<i>Eusarcoris aeneus</i> (Scopoli, 1763)							5	0,17	4,5									
<i>Holcostethus (Holcostethus) vernalis</i> (Wolff, 1904)							6	0,21	5,4	1	0,06	4	4	0,33	14,28			
<i>Carpocoris purpureipennis</i> (De Geer, 1773)							9	0,32	8,1					2	0,16	7,14		
<i>Carpocoris pudicus</i> (Poda, 1761)	2	0,16	5,12				1	0,03	0,9									
<i>Dolycoris baccarum</i> (Linné, 1758)	4	0,33	10,25	1	0,12	11,11	17	0,6	15,31	5	0,31	20	9	0,75	32,14			
<i>Piezodorus lituratus</i> (Fabricius, 1794)	2	0,16	5,12				8	0,28	7,2									
<i>Eurydema ornatum</i> (Linné, 1758)							2	0,07	1,8					1	0,08	3,57		
<i>Eurydema oleraceum</i> (Linné, 1758)	21	1,75	53,84	2	0,25	22,22	5	0,17	4,5	4	0,25	16	2	0,16	7,14			
<i>Picomerus bidens</i> (Linné, 1758)							4	0,14	3,6									
<i>Zicrona coerulea</i> (Linné, 1758)							3	0,1	2,7									
Numărul total de indivizi/stațiune	39		9			111			25			28						
Numărul total de indivizi/ tip de ecosistem		48				136						28						
Numărul total de specii/stațiune	6		4			16			7			8						
Numărul total de specii/ tip de ecosistem		6				17						8						

Abrevieri: IV - Fânaț higrofil. V - Fânaț mezofil. VI - Stâncării. 1 - Moldovenești. 2 - Mărtinești. 3 - Cheile Turenilor. 4 - Luna. 5 - Gura Arieșului. A - Abundența numerică. Ar - Abundența relativă. N - Numărul de indivizi.

Aelia acuminata, specie polifagă caracteristică ecosistemelor ierboase însorite, a înregistrat valorile cele mai ridicate ale abundenței numerice și absolute în fânațele mezofile studiate ($A = 0,78$, $Ar = 19,82$ în stațiunea Moldovenești, $A = 0,43$, $Ar = 28$ în stațiunea Mărtinești). Valori ridicate ale acestor indici în fânațele mezofie au înregistrat și speciile *Coptosoma scutellatum*, *Eurygaster maura*, *Carpocoris purpureipennis* și *Dolycoris baccarum* (Tabel 3).

Pe stâncăriile din Cheile Turenilor, cel mai bine reprezentată este *Dolycoris baccarum* ($A = 0,75$, $Ar = 32,14$) urmată de *Aelia acuminata* ($A = 0,5$, $Ar = 21,42$) și *Holcostethus vernalis* ($A = 0,33$, $Ar = 14,28$) (Tabel 3).

În păsunile studiate, valorile cele mai ridicate ale abundenței numerice și relative au fost înregistrate de către *Eurydema oleraceum* în stațiunea Luna ($A = 2,31$, $Ar = 80,43$) (Tabel 4). Prinț-un număr mare de indivizi au fost prezente și speciile *Picomerus bidens* ($A = 0,62$, $Ar = 50$ în Mărtinești) - una dintre puținele specii de pentatomoide zoofage, adulții apărând în lunile de toamnă, *Dolycoris baccarum* ($A = 0,58$, $Ar = 58,33$ în Cheile Turenilor) și *Coptosoma scutellatum* ($A = 0,5$, $Ar = 35,29$ în Moldovenești).

Din cultura de lucernă, a fost colectat un număr mare de indivizi de *Piezodorus lituratus* ($A = 1,16$, $Ar = 40$), specie frecventă în culturile de leguminoase din Transilvania [6]. Atât din cultura de lucernă, cât și de pe vegetația ruderălă au fost colectate în număr relativ mare *Dolycoris baccarum* și *Eurydema oleraceum*. În vegetația ruderălă din stațiunea Mărtinești, valori ridicate ale abundenței numerice și relative au prezentat și *Coptosoma scutellatum* ($A = 0,75$, $Ar = 39,13$) și *Graphosoma lineatum* ($A = 0,41$, $Ar = 21,74$) (Tabel 5).

Diversitatea cea mai ridicată s-a înregistrat în fânațul mezofil din stațiunea Moldovenești urmat de făgeto-cărpinetul studiat în aceeași stațiune și de stâncăriile studiate în Cheile Turenilor. Diversitatea cea mai scăzută s-a înregistrat în păsunile din Gura Arieșului și Luna (Tabel 6).

Dinamica populațiilor. În vederea efectuării studiilor de dinamică a populațiilor de Pentatomoidea, în stațiunea Moldovenești au fost efectuate colectări lunare. Atât numărul de indivizi cât și numărul de specii variază pe parcursul anului (Tabel 7). Numărul maxim este atins în luna iunie, când sunt prezenti atât indivizi ai unor specii la care generația ieșită din diapauză se menține până în această perioadă, cât și indivizi ai unor specii la care a apărut noua generație. În iulie, atât datorită atingerii unui grad ridicat de exploatare al păsunii studiate, cât și datorită cosirii fânațului și culturii de lucernă, numărul de specii și de indivizi a scăzut. În luna august, când vegetația ierboasă din ecosistemele antropizate s-a refăcut parțial, numărul de indivizi și de specii de Pentatomoidea a înregistrat o ușoară creștere comparativ cu luna iulie. Valorile înregistrate în luna mai (situația populațiilor la ieșirea din diapauză) sunt asemănătoare cu valorile înregistrate în lunile septembrie și octombrie (situată populațiilor la intrarea în diapauză) (Fig. 2).

Spectrul trofic. În urma analizei spectrului trofic al pentatomoideelor semnalate se remarcă speciile polifage (18), acestea reprezentând 69%. Nu a fost semnalată nici o specie monofagă. Dintre cele 26 de specii colectate 8 sunt oligofage, 4 specii fiind întâlnite pe graminee, 3 pe crucifere, iar una pe leguminoase (Fig. 3).

SUPRAFAMILIA PENTATOMOIDEA ÎN SECTORUL INFERIOR AL ARIEȘULUI

Tabel 4

**Abundență numerică și abundență relativă a speciilor de
Pentatomoidea colectate din ecosistemele ierboase intens
exploataste (pășuni) din sectorul inferior al Bazinului Arieșului**

Familia, specie	VII														
	1			2			3			4			5		
	N	A	Ar %	N	A	Ar %	N	A	Ar %	N	A	Ar %	N	A	Ar %
Familia Plataspidae															
<i>Coptosoma scutellatum</i> (Geoffroy, 1785)	12	0,5	35,29							2	0,12	4,34			
Familia Cydnidae															
<i>Tritomegas bicolor</i> (Linné, 1758)										1	0,06	2,17			
Familia Scutelleridae															
<i>Eurygaster testudinaria</i> (Geoffroy, 1785)	4	0,16	11,76	1	0,06	5									
Familia Pentatomidae															
<i>Graphosoma lineatum</i> (Linné, 1758)							1	0,08	8,33						
<i>Podops inuncta</i> (Fabricius, 1775)										1	0,06	2,17			
<i>Sciocoris (Sciocoris) deltocephalus</i> Fieber, 1861	1	0,04	2,94												
<i>Aelia acuminata</i> (Linné, 1758)	2	0,08	5,88	1	0,06	5	1	0,08	8,33						
<i>Aelia rostrata</i> Bohemian, 1852				2	0,12	10									
<i>Holcostethus (Holcostethus) vernalis</i> (Wolff, 1904)	1	0,04	2,94				1	0,08	8,33				2	0,25	66,67
<i>Carpocoris pudicus</i> (Poda, 1761)										1	0,06	2,17			
<i>Dolycoris baccarum</i> (Linné, 1758)	10	0,41	29,41	5	0,31	25	7	0,58	58,33	3	0,18	6,52			
<i>Eurydema ventrale</i> Kolenati, 1846										1	0,06	2,17			
<i>Eurydema oleraceum</i> (Linné, 1758)	4	0,16	11,76	1	0,06	5	2	0,16	16,65	37	2,31	80,43	1	0,12	33,33
<i>Picomerus bidens</i> (Linné, 1758)					10	0,62	50								
Numărul total de indivizi/ stațiune	34			20			12			46		3			
Numărul total de indivizi/ tip de ecosistem							115								
Numărul total de specii/ stațiune	7			6			5			7		2			
Numărul total de specii/ tip de ecosistem							14								

Abrevieri: VII - Pășune. 1 - Moldoveniști. 2 - Mărtinești. 3 - Cheile Turenilor. 4 - Luna. 5 - Gura Arieșului. A - Abundență numerică. Ar - Abundență relativă. N - Numărul de indivizi.

Tabel 5

**Abundență numerică și abundență relativă a speciilor de *Pentatomoidae*
colectate din ecosisteme ierboase (cultură de lucernă, vegetație ruderală)
situate în zone puternic antropizate din sectorul inferior al Bazinului Arieșului**

Familia, specia	VIII						IX					
	1		1		2		4		5			
	N	A	Ar %	N	A	Ar %	N	A	Ar %	N	A	Ar %
Familia <i>Plataspidae</i>												
<i>Coptosoma scutellatum</i> (Geoffroy, 1785)	2	0,08	2,85				9	0,75	39,13			
Familia <i>Cydnidae</i>												
<i>Tritomegas sexmaculatus</i> (Rambur, 1842)										1	0,08	5,56
Familia <i>Scutelleridae</i>												
<i>Eurygaster maura</i> (Linné, 1758)	7	0,29	10									
<i>Eurygaster testudinaria</i> (Geoffroy, 1785)	1	0,04	1,42									
Familia <i>Pentatomidae</i>												
<i>Graphosoma lineatum</i> (Linné, 1758)							5	0,41	21,74			3 0,37 42,85
<i>Aelia acuminata</i> (Linné, 1758)	1	0,04	1,42	3	0,12	14,28	1	0,08	4,35	4	0,33	22,22
<i>Eusarcoris aeneus</i> (Scopoli, 1763)										1	0,08	5,56
<i>Holcostethus</i> (<i>Holcostethus</i>) <i>vernalis</i> (Wolff, 1904)	3	0,12	4,28	1	0,04	4,76	1	0,08	4,35			
<i>Carpocoris purpureipennis</i> (De Geer, 1773)	2	0,08	2,85	2	0,08	9,52	1	0,08	4,35			
<i>Carpocoris pudicus</i> (Poda, 1761)	2	0,08	2,85				1	0,08	4,35	1	0,08	5,56
<i>Dolycoris baccarum</i> (Linné, 1758)	11	0,45	15,71	5	0,2	23,8	1	0,08	4,35	2	0,16	11,11
<i>Piezodorus lituratus</i> (Fabricius, 1794)	28	1,16	40									
<i>Eurydema ventrale</i> Kolenati, 1846				1	0,04	4,76						
<i>Eurydema oleraceum</i> (Linné, 1758)	13	0,54	18,57	9	0,37	42,85	4	0,33	17,39	7	0,58	38,88
<i>Zicrona coerulea</i> (Linné, 1758)										2	0,16	11,11
Numărul de indivizi/ stațiune	70		21		23		18		7			
Numărul de indivizi/ tip de ecosistem	70						69					
Numărul de specii	10			6		8		7		3		
Numărul de specii/ tip de ecosistem	10						12					

Abrevieri: VIII - Cultură de lucernă. IX - Vegetație ruderală. 1 - Moldovenești. 2 - Mărtinești.
4 - Luna. 5 - Gura Arieșului. A - Abundență numerică. Ar - Abundență relativă. N - Numărul de indivizi.

Tabel 6

**Diversitatea și echitabilitatea ecosistemelor studiate
în funcție de populațiile de *Pentatomoidea***

Ecosistemele și stațiunile studiate																			
I	II	III	IV	V	VI	VII	VIII	IX											
1	3	4	5	4	5	1	2	3	1	2	3	4	5	1	1	2	4	5	
H'	0,92	0,76	0,53	0,47	0,60	0,49	1	0,76	0,79	0,69	0,59	0,53	0,35	0,27	0,76	0,65	0,73	0,72	0,46
E	0,92	0,97	0,89	0,99	0,77	0,82	0,83	0,9	0,87	0,82	0,76	0,76	0,42	0,91	0,76	0,83	0,80	0,85	0,98

Abrevieri: E - Indicele de echitabilitate. H' - Indicele de diversitate Shanon-Wiener. I-IX - Tipurile de ecosisteme studiate. 1 - Moldovenești. 2 - Mărtinești. 3 - Cheile Turenilor. 4 - Luna. 5 - Gura Arieșului.

Tabel 7

**Dinamica populațiilor de *Pentatomoidea* în ecosistemele
studiate în zona localității Moldovenești**

Perioada colectării	Făgeto-cărpinet		Fânaț mezofil		Pășune		Cultură de lucernă		Vegetație ruderală		Numărul total	
	nr. ind.	nr. sp.	nr. ind.	nr. sp.	nr. ind.	nr. sp.	nr. ind.	nr. sp.	nr. ind.	nr. sp.	ind.	sp.
Mai	3	3	40	11	2	1	13	7	1	1	59	13
Iunie	19	6	16	7	15	4	16	4	9	4	75	15
Iulie	15	4	11	4	2	2	5	3	4	2	37	9
August	13	4	20	5	5	3	19	5	3	2	60	12
Septembrie	4	2	9	4	8	3	7	3	1	1	29	8
Octombrie	10	3	15	7	2	2	10	4	3	2	40	8

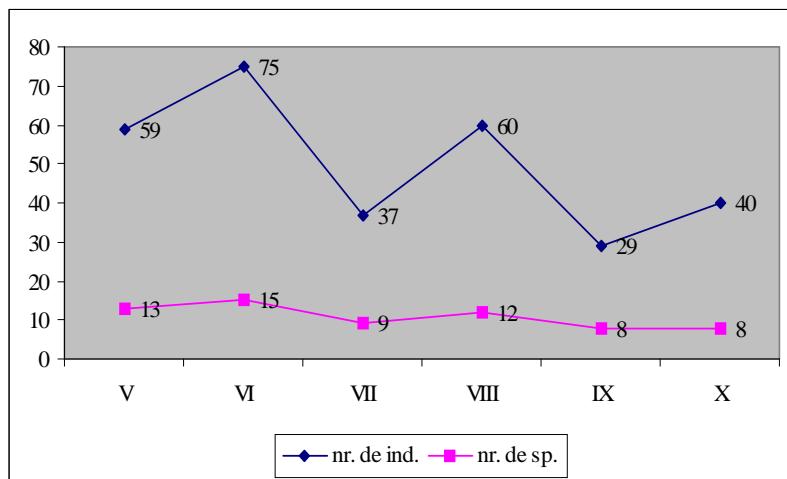
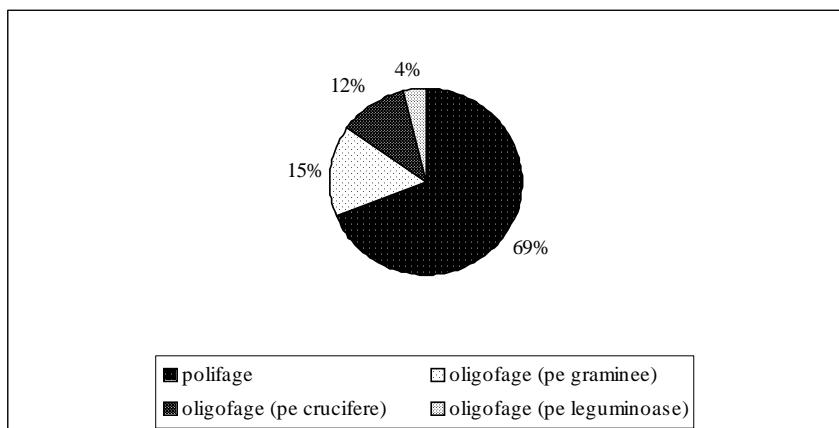
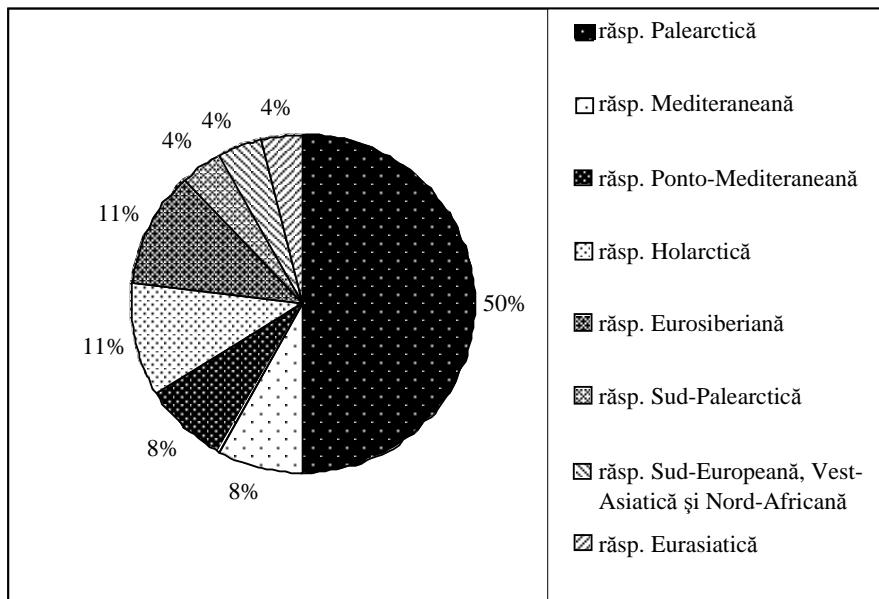


Fig. 2. Dinamica suprafamiliei Pentatomidae în zona localității Moldovenești.



F i g. 3. Spectrul trofic al pentatomoideelor semnalate în sectorul inferior al Bazinului Arieșului.

Analiza zoogeografică. Din punct de vedere al spectrului zoogeografic cel mai bine reprezentate sunt speciile cu răspândire Palearctică (13). Au fost semnalate și specii cu răspândire Mediteraneană (2), Ponto-Mediteraneană (2), Holarctică (3), Eurosiberiană (3), Sud Palearctică (1), Sud Europeană, Vest Asiatică și Nord Africană (1) și Eurasiatrică (1) (Fig. 4).



F i g. 4. Spectrul zoogeografic al pentatomoideelor semnalate în sectorul inferior al Bazinului Arieșului.

Concluzii. 1. Din sectorul inferior al Bazinului Arieșului am colectat 26 specii de *Pentatomoidae* aparținând la 19 genuri și 4 familii. Atât din punct de vedere al numărului de specii, cât și al numărului de indivizi, cel mai bine reprezentată este familia *Pentatomidae*.

2. Specia *Sciocoris (Sciocoris) deltocephalus* este citată pentru prima dată în fauna Transilvaniei, iar trei specii citate sunt sporadice în fauna României: *Podops inuncta*, *Stagnomus (Dalleria) pussilus* și *Eurydema ventrale*.

3. Numărul cel mai mare de specii s-a înregistrat în ecosistemele ierboase și scade odată cu scăderea cantității de vegetație ierboasă, reducerea numărului de plante prezente și creșterea influențelor antropice în ecosistemele studiate.

4. Cele mai ridicate valori ale abundenței numerice și relative au fost înregistrate în ecosistemele ierboase de către speciile *Coptosoma scutellatum*, *Aelia acuminata*, *Dolycoris baccarum* și *Eurydema oleraceum*.

5. Diversitatea cea mai ridicată s-a înregistrat în fânațul mezofil din stațiunea Moldoveniști, iar cea mai scăzută în păsunile din Gura Arieșului și Luna.

6. În dinamica populațiilor de *Pentatomoidae* s-a înregistrat o variație a numărului de indivizi și de specii pe parcursul anului, atingând maximul în luna iunie.

7. Din cele 26 de specii, 69% sunt polifage, iar 31% sunt oligofage. Nu a fost semnalată nici o specie monofagă.

8. Din punct de vedere zoogeografic, predomină speciile cu răspândire Palearctică, urmate de speciile cu răspândire Holarctică și de cele cu răspândire Eurosiberiană.

B I B L I O G R A F I E

1. C s ū r ö s – K á p t a l a n, M., *Aspecte din vegetația Cheii Turului*, “Stud. Univ. Babeș-Bolyai, Biol.”, (1), 1962, 17-31.
2. D a v i d o v a – V i l i m o v a, J., M c P h e r s o n, J. E., *History of the higher classification of the subfamily Podopinae (Heteroptera: Pentatomidae), a historical review*, “Acta Univ. Carol., Biol.”, **38** (2), 1994 (1995), 99-124.
3. H o r v á t h, G., *Fauna Regni Hungariae, Hemiptera*, Budapest, 1897.
4. K i s, B., *Insecta, Heteroptera: Partea generală și suprafamilia Pentatomidea*, Fauna R. S. R., **VIII** (8), București, 1984.
5. P a s c u, M., *Studiul ecologic al artropodelor dintr-o pajiște de luncă din județul Sibiu*, “Muz. Brukenthal, Stud. Comun., Ști. Nat.”, **21**, 1983, 279-307.
6. P e r j u, T., S c h n e i d e r, E., *Contribuții la cunoașterea faunei de Heteroptere din culturile de leguminoase perene din Transilvania de Nord-Vest*, “Muz. Brukenthal, Stud. Comun., Ști. Nat.”, **17**, 1972, 277-289.
7. P o p o v, C., R o š c a, I., *Recent researches regarding cereal bugs (Eurigaster ssp.) in Romania*, “Rev. Roum. Biol., Sér. Biol. Anim.”, **36** (1-2), 1991, 51-56.

8. Rákosy, L., Véhmann, I., *Argumente în favoarea unei rezervații naturale în Cheile Turului*, "Ocrot. Nat. Med. Înconj.", **35** (1-2), 1991, 15-25.
9. Roșca, I., *A zoogeographic analysis of the species of the superfamily Pentatomidae, Reuter 1910, as recorded in Romania*, "Rev. Roum. Biol., Sér. Biol. Anim.", **28** (1), 1983, 55-58.
10. Spunar, T., Voicu, T., *Insecte dăunătoare și folositoare din culturile de grâu și aspecte de combatere integrată*, "Lucr., III-a Conf. de Entomol." (Iași, 1983), 459-467.
11. Schneide, E., *Suprafamilia Pentatomidea Reut. 1910 (Heteroptera) în colecțiile Muzeului de Istorie Naturală din Sibiu*, "Muz. Brukenthal, Stud. Comun., Ști. Nat.", **15**, 1970, 225-245.
12. Schuh, R. T., Slater, J. A., *True Bugs of the World* (Hemiptera: Heteroptera). *Classification and Natural History*, Cornell Univ. Press, New York, London, 1995.
13. Southwood, T. R. E., Leston, D., *Land and Water Bugs of the British Isles*, Frederick Warne, London, 1959.
14. *** Kis, B., *Fauna de dermaptere, mantodee și heteroptere din Cheile Turenilor*, Comun. pers., 1995.

BREEDING BIOLOGY OF THE GOLDEN EAGLE (*AQUILA CHRYSÆTOS CHRYSÆTOS* L. 1758) IN THE GORGE OF TURDA NATURE RESERVE

ALIN DAVID* and IOAN COROIU*

SUMMARY. - The paper presents the results of the research regarding the biology and ethology of the Golden Eagle (*Aquila chrysaetos chrysaetos*) during the breeding season. The Golden Eagle is an extremely rare species in Romania as well as in its whole spreading area, due to the loss of its specific habitat and excessive hunting. Protecting this species requires, apart from the legislative measures, the knowledge of its biological features. The present study took place between 2000 and 2003 in the Gorge of Turda Nature Reserve and its main objective was to analyse breeding biology of this species. The observations were carried out in 140 days (70 days in the year 2000, 50 days in 2001 and 10 days in 2002 and in 2003). In 2000, the behaviour of the adults as well as that of the hatchlings were monitored during the whole breeding season. In 2001, great attention was paid to the brooding biology and in 2002 and 2003, the main objective of the investigations was the breeding success. In each of these 4 years, two eggs were laid at 3-day intervals. Both partners take part in the brooding, but in different proportions, the main role belonging to the female (67.53% in 2000 and 70.7% in 2001). The incubation lasted for 42 days. In 2000, in the fledging period, the male provided food for the family in the proportion of 85.24%. As a result of the competition between the hatchlings, in three of the four seasons the younger hatchling died about one month after hatching. The duration of fledging period was 65 days in 2000 and 64 days in 2001. Totally, the breeding success was of 62.5%. Knowledge of the ecological requirements and of the reproductive features of the Golden Eagle is a very important step for working out the long-term conservation strategies that are so important for the survival of this species.

The Golden Eagle is the biggest *Aquila* species and has the largest spreading area among the species of this genus. This area comprises the whole Holarctic, where the species is represented by 6 subspecies. The nominate *Aquila chrysaetos chrysaetos* (Linnaeus, 1758) is spread in the whole Europe, except for the Iberian Peninsula (where it is replaced by *A. c. hommeyery* (Svertzov, 1888), which also reaches into North Africa, Asia Minor and Iran) and it reaches, in the East, well into Siberia and the Altai Mountains. In Turkestan, Manchuria, south-western China, northern India and Pakistan *A. c. daphneia* (Svertzov, 1888) can be found, and in the western part of Siberia, of the Altai Mountains and in Kamtchatka the

* Babeş-Bolyai University, Department of Taxonomy and Ecology, 400006 Cluj-Napoca, Romania.
E-mail: adavid@biolog.ubbcluj.ro and icoroiu@biolog.ubbcluj.ro

subspecies *A. c. kamtchatica* (Severtzov, 1888) is present. The smallest of the subspecies, *A. c. japonica* (Severtzov, 1888) lives in the Korean Peninsula and the Archipelago of Japan, while *A. c. canadensis* (Linnaeus, 1758) is present on the North American continent^{**}.

The mountainous regions represent the most typical habitat of this species; it can rarely be found in depressionary forests and in wet lowlands. For nestling, it needs inaccessible places and uses cliffs and sea walls for this purpose. In Great Britain, this species was identified brooding between the altitudes of 16 and 900 m [20]. In the Alps, the nestling territories are situated between 1000 and 2460 m, exceptionally at 3000-m altitude [7] and in the Caucasus Mountains, nests were seen at altitudes between 1200 and 2400 m [8].

The trophic range is very varied, the diet of the Golden Eagle containing small and middle-sized mammals (*Lepus*, *Capreolus*, *Sus scrofa* or *Alces alces* youngs), as well as middle-sized and large birds (*Tetrao*, *Lyrurus*, *Phasianus*) and even reptiles (tortoises). Studies carried out in captivity have shown that the food requirements for an adult Golden Eagle is around 309 g of food/day for a female and 260 g of food/day for a male [11]. Such studies have also been done in Scotland, involving free individuals. These consume, on the average, 200 g of food/day, resulting that the annual necessity is around 84 kg/individual. By considering the losses too, it has been estimated that an adult Golden Eagle captures 214 kg of prey each year [5].

So, the Golden Eagle needs a large amount of food, when the efficiency of prey capturing does not exceed 30% [1]. Any qualitative and quantitative changes in the structure of their specific habitat can lead to the decrease of their chance of survival and breeding success.

In the first seven decades of the last century the number of individuals of this species has decreased at an alarming rate in the whole spreading area, mainly due to anthropic causes, such as habitat loss, excessive hunting and the use of intended carcasses poisoned with strychnine for the control of carnivorous mammals. In France, in this period, the number of Golden Eagle has decreased by about 70%, only 60 breeding pairs remaining [31]. A similar situation was recorded in Germany, where at the beginning of the seventies only 15 pairs were still breeding [12] and also in Great Britain, where in the sixties about 3000 pairs were breeding [9].

As a result, this species was put under protection. At this moment, due to the Appendix no. 3 of the Bern Convention, the Golden Eagle is a strictly protected species and the Appendix no. 1 of the Birds Directive [32] shows that the Golden Eagle is a priority species for conservation in the whole Europe. Under these conditions in some of the countries of Europe, the species effective became stable and even started to grow. For example, in France, the present number of Golden Eagle is estimated to be around 288 breeding pairs, in Germany 50 breeding pairs and 422 pairs breed in Great Britain [15]. The most important effectives are presently in Turkey, where 3000 pairs are breeding, in Spain (1100 breeding pairs) and in Norway, where the effective of this species is estimated to be about 1000 pairs [15].

^{**} Linnaeus (1758) and Severtzov (1888) are cited based mainly on the review published by [7].

At the beginning of the last century, in Romania, the Golden Eagle was a widespread species. D o m b r o w s k y used to say "in the Carpathians of Romania the Golden Eagle is a frequent species" [10], but in the 50's the situations appeared to have changed: "the king of the birds and of the sky has become such a rare bird that I would be optimistic to estimate to about 50 the number of breeding pairs of Golden Eagle on the territory of R.P.R." [17]. The present data regarding the situation of this species in Romania are relatively few. At the beginning of the 90's, the whole effective of this species was estimated to be around 10 pairs [6], and 10 years later, the effective was estimated to be around 30-40 breeding pairs [21].

By joining the Bern Convention [33] and by issuing Law no. 462 [34], Romania made important steps towards the conservation of this species in its wild fauna. Although the anthropic factors have played and are still playing a decisive role, the decline of the Golden Eagle can not be explained only from this point of view. It is possible that the behavioural particularities of the species combined with these anthropic factors could have lead to the present situation. For the application of protective measures, it is necessary to know the ecology and ethology of the species involved.

The Golden Eagle, heraldic symbol of the Romanian State, has not yet been studied in Romania from a behavioural point of view, mostly due to the difficulties raised by this kind of study. By taking advantage of the specific conditions of the Gorge of Turda, favourable for this purpose, the present study is a first step in this direction. The data about the presence of the Golden Eagle in this region are sporadic. N y á r á d y mentioned it for the first time in 1937 [23]. One year later, S e n c h e a reports once again the Golden Eagle [26], then K o h l in 1969 [16] and R ă s i n a r u in 1995 [24]. None of these authors dealt with the behavioural aspects of the life of Golden Eagle.

Methods. The observations were carried out in The Gorge of Turda Nature Reserve between the years 2000 and 2003. The one Golden Eagle (*Aquila chrysaetos chrysaetos*) pair that living here, was monitored by the direct observation method using for this goal a Nikon 26X45 field scope.

In 2000, the observations covered the entire breeding season. They began on 18 of February and continued weekly until 1 of July, totalling 70 days. In 2001, the observations began on the 15 of February and followed the same pattern until 28 of May (31 days) focusing upon the brooding biology. Also, during the 19 days, observations were done upon fledging period and the relationship between hatchlings until the death of the younger one. In the years 2002 and 2003, the breeding success was the main aspect studied.

Results and discussions. The territory. L o r e n t z defined territory as "the space protected by an individual, into which other individuals of the same species are not allowed to enter" [18]. The existence of the territory confers maximal advantages to the species. The chance for the congeners to meet is reduced to the

minimum, so that the intraspecific competition is attenuated, the efficiency of space usage is increased, the reproductive chance is also increased and the energy of the couple is economised in order to be channeled strictly towards breeding success.

The surface of the territory varies much among the different species. The existing trophic resources condition it. Some observations carried out during 2000 and 2001 allow the estimation of the territory surface of the studied Golden Eagle pair. Thus, on the day of 19th of April 2000, in full breeding season, the male was seen in a tree at the skirt of a forest close to the village of Micesti. Eight days later, on the 27th of April 2000 the same male was seen flying above the grazing grounds near the village of Apahida. Characteristic of this male in the mentioned period was the lack of the third secondary remige. On the 20th of June 2000 the female was seen on the Ciucăş Hill, near Corneşti village (about 12 km downstream on the Hăşdate, towards southeast). The last observation of this kind was done on 18th of May 2001 at Ocna-Mureş, where the male was sighted flying above the moorland of the Mureş, which accounts for a surface of about 300 km² (Fig. 1).

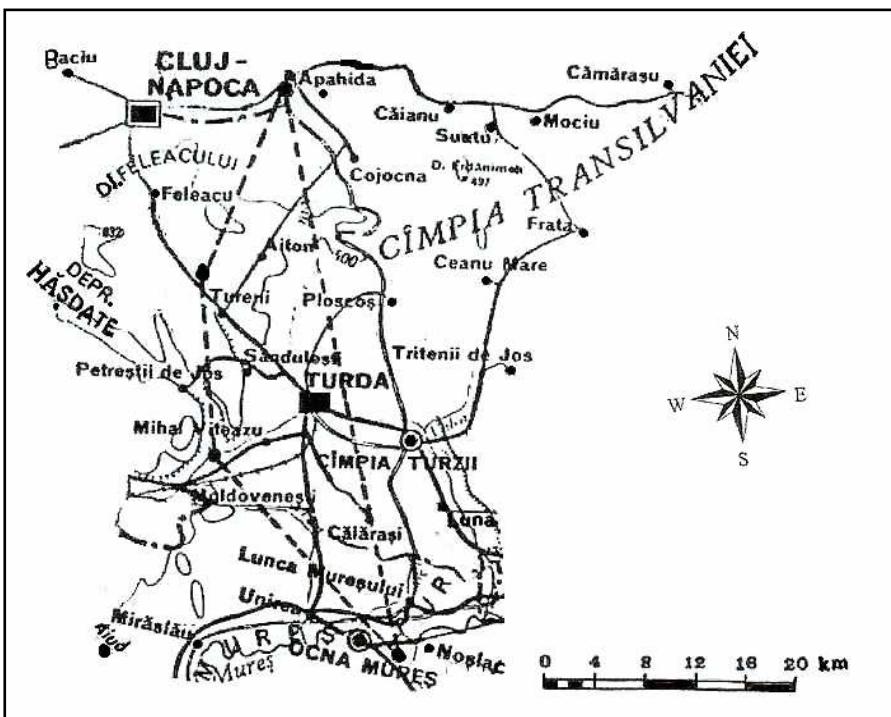


Fig. 1. Presumptive limits of the Golden Eagle pair territory.

Being top predators, the Golden Eagles have very large territories, the surface of which is between 50 and 200 km² and in some regions that are less suitable for the species like the Apennine Mountains this surface can exceed 300 km² [14].

This kind of territory in which Golden Eagles perform all their vital activities is called reproduction territory [3] and it is made up of two components: the vital range and the feeding range and there is contiguity between these two components. The establishment in time of the two components is done gradually and it is achieved in the moment when the eggs are laid and the nest becomes the biological centre of the territory.

Between November and January 2000–2002, the Golden Eagles were missing from the Gorge of Turda. Fidelity towards the territory is a characteristic feature of this species. They leave it only when food becomes insufficient. Starting from the middle of February the couple comes back into the territory. This is the moment when they have to consolidate their position as owners of the territory and to decide upon the nest that they would use.

The most usual way for marking the territory is based on optic signals, given the extraordinary eyesight of the eagles. These signals are represented by the presence of one or both partners flying above the territory.

In order to investigate the couple's return to the nest, observations were done for a period of 9 days in 2000 and 7 days in 2001. The increasing periods of time spent by the partners flying above the territory was done gradually (Fig. 2).

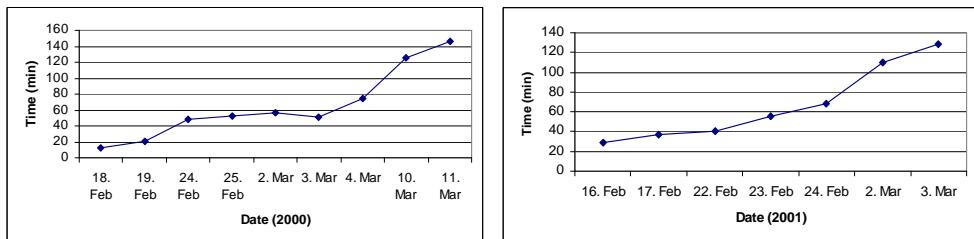


Fig. 2. Dynamics of the presence of Golden Eagle in the Gorge of Turda before the laying period.

Preparing the nest for laying the eggs. In the Gorge of Turda, the zone of minimal altitude in Romania (about 500m) where the Golden Eagle breeds, there are 7 nests, all of which belong to the studied pair. Five of them are situated on the “Peretele Vulturilor” and other two on the “Colțul cel Lat” (Fig. 3).

Supernumerary nests are a characteristic of the species; the average being of 5.3-nests/pair [28] and the reported maximum number is 12-nests/pair [27].

The nests are used in turn. The Golden Eagles do not hatch two consecutive years in the same nest. This is an efficient adaptation for diminishing the attack of ecoparasites (*Mallophaga*), the reproductive cycle of these being synchronised with emergence of the hatchlings [30].

After the adults return to the territory an extremely important first step is choosing the nest into which the eggs are to be laid, the decisive role belonging to the female. This aspect result from the longer time spent by the female at the nests (Fig. 4).

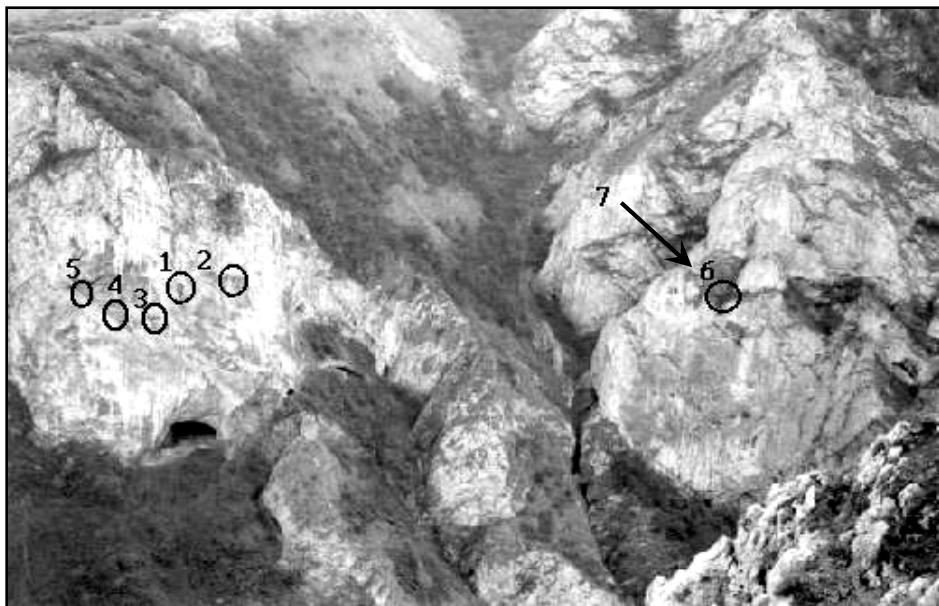


Fig. 3. Position of the nests of the studied Golden Eagle Pair.

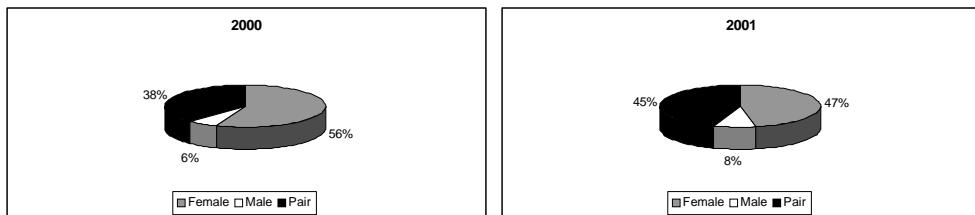


Fig. 4. Proportion of the presence of partners at the nests before the laying period.

In each year, before laying the eggs, the couple rebuilt two or three nests and finally one of them was chosen. Thus, by the multi-year use of the nest, some of them acquired impressive dimensions. The dimensions of nest no. 1, the oldest in the area, can be estimated to 4 m in height and 2.5 m in diameter, with a volume of about 7 m³. Nests with a diameter varying between 1.5 and 3 m and 2 to 4 m in height, having a volume of up to 6 m³ are reported in the literature [4]. Probably, these are used in successive generations (the nest no. 1 was photographed by Nyárády in 1939 [23]).

From the seven existing nests, in the year 2000 the female rebuilt nests no. 1, 2 and 3, finally choosing nest no. 2 and in 2001 fresh vegetal materials (pine and spruce branches) was seen in nests no. 3 and 6, the eggs being laid in nest no. 3.

Laying and brooding biology. In each of the studied four breeding seasons the female laid two eggs. In the year 2000, the first egg was laid on the 18th of March and in 2001 on the 10th of the same month. This 8-day delay probably was due to higher temperatures in 2001 as compared to 2000. In 2000, the second egg was laid on the 21st of March; between the layings of the two eggs there was a delay of about 76 hours. In 2001, the second egg was laid on the 13th of March at an interval of 73 hours after laying of the first egg. The laying of the Golden Eagle is most often composed of two eggs, rarely one egg and only in exceptional cases, of three eggs. In Scotland, from 82 studied nests, 18% contained one egg, 72% contained two eggs and only 10% of them containing three eggs [13]. Between the laying of the first egg and that of the second there is a delay of three or even of four days [2].

Brooding began with the laying of the first egg, the incubation period lasting for about 42 days both in 2000 and in 2001. Observations upon this aspect have been done for a period of 27 days in 2000 and for a period of 24 days in 2001.

Both parents took part in brooding, but in different proportions, the main role belonging to the female (Figs. 5 and 6).

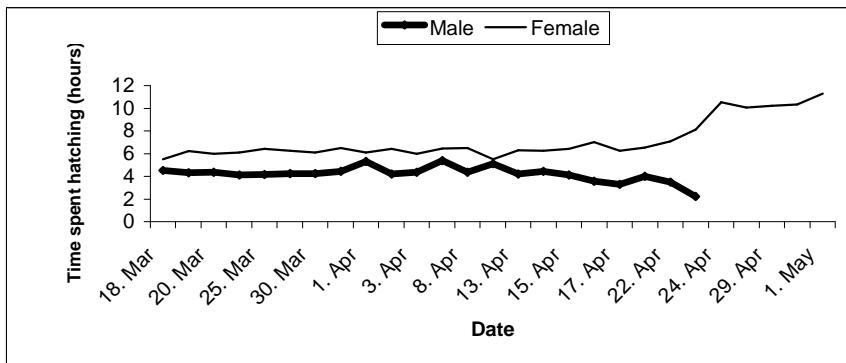


Fig. 5. Time spent brooding by the two partners in 2000.

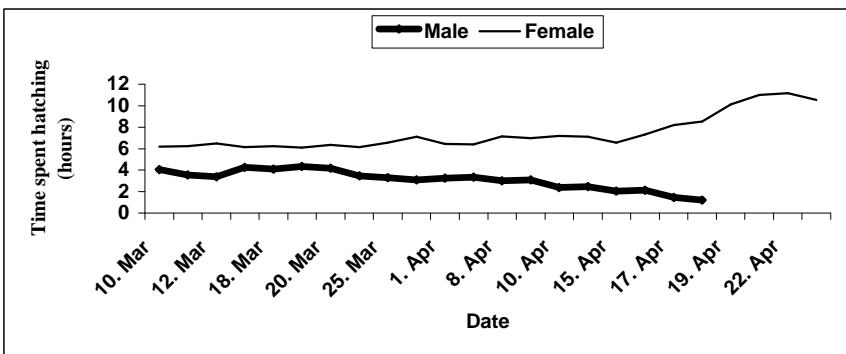


Fig. 6. Time spent brooding by the two partners in 2001.

Thus, during the daytime, the female insured brooding in the proportion of 67.53% in 2000 and 70.7% in 2001. The observations have shown that during the night only the female spent the night at the nest brooding. The time spent by male brooding decreased gradually toward the moment of hatching. In 2000, the male was not seen brooding for about 40 hours before the hatching of the first young and in 2001 for about 35 hours before hatching. Also, the male no longer brooded in the interval between the hatching of the first and of the second young. During the whole period of brooding both adults got their food independently. The participation of both partners to brooding is common to many species of eagles and hawks, including the representatives of the genus *Aquila* [22], but in that case the female's contribution to brooding is more significant than that of the male [2, 20].

The ethogram of the shifts of partners at the nest. The partners shift is done in a ritual manner and it is a consequence of the rhythmicity of behavioural acts. In order to know when the partner who is to be changed is prepared to do it, the other partner provokes an answer on the question: will it take over the brooding or not. In getting close to the nest and sitting onto the eggs, the Golden Eagle makes a series of characteristic movements (Table 1).

Table 1
Ethogram of the shifts of partners at the nest

Male	Female
1. Appearance closer to the nest.	2. Optically received signal, change of position from normal to attentive (stretched neck).
3. Flying in the front of the nest and emitting the sound signal.	4. Acts of shifting the surroundings objects. Usually, branches are taken out from one part of the nest and put in another part. Response to the sound signal.
5. Landing on the edge of the nest.	6. Cleaning the feathers by means of the beak, followed by the repeated shakings of the feathers and stretching of the wings.
7. Getting close to the eggs, bowing of the head and pulling it between the shoulders.	8. Taking off in flight.
9. Rotating the eggs.	10. Flying in large circles in front of the nest.
11. Binding of the legs and sitting on the eggs.	12. Flying in larger and larger circles and away to the feeding range.
13. The normal brooding position, followed or not by sleep.	

Note: in the case of changing the male by the female the procedure is reversed.

The ritual of the shifting of partners at the nest is a mixt one, a combination of clock, optical and sound ritual [25]. The most important stimuli that unleash the ceremony shifting are the optical ones. Being birds with small chances of obtaining their food in equal time limits (the smaller size of the male confers greater agility and thus greater chances to capture the prey) and in short time (usually these last

for hours, in which smaller or greater distances are covered), they would shift at the nest only when the partner in search for food, appears around the nest. The clock ritual, characteristic to grain feeding birds, for example, cannot be respected due to these reasons. Nevertheless, the time allocated by the individual for brooding (with few exceptions in unfavorable weather conditions) is relatively well delimited and respected as it is.

As a result of the observations in 2000 and 2001, an average of about 3 hours and 10 minutes elapsed until the replacement of the female by the male and an average of about 1 hour 20 minutes was necessary until the replacement of the male by the female. The attempts to replace the brooding partner under the mentioned time limit, that is, without having spent a certain number of hours, did not have positive results in replacement of the partner.

These data only account for the rigid mathematical value of the process. In reality, there have been situations in which, because of the unfavorable weather conditions, the period allocated to brooding by one individual was prolonged to 5 hours, in the case of both partners.

There is also the possibility of replacing the partners at the nest without any ceremony. This happens when one of the partners delays its arrival with much time over the average allocated to brooding by its partner, in favorable weather conditions. So, the time factor has deeper implications in the ethology of the Golden Eagle for the given sequence, probably with more complex origins in the phylogeny of the *Falconiformes* order itself. The acoustic signals seem to be of secondary importance.

The hatching. In 2000, the hatching of the first young took place on the 28th of April and that of the second on the 1st of May, at an interval of about 72 hours. The same interval was kept in 2001, when the first hatchling emerged on the 20th of April and the second on the 23rd of April.

Rotating the eggs stopped completely with about two days before the hatching. In this period of time the male was not observed brooding, although this made short visits to the nest, bringing fresh vegetal material, which the female arranged in the nest. In all the cases in the moment of hatching only the female was present to the nest. It does not have a typical behaviour during the hatching, although the phenomenon is indicated in its phenology by its almost imperceptible rising from the eggs and the shift stretching of the wings. The hatchlings emerged by themselves, without help of the female.

In many of the species, the shells of eggs are immediately removed after hatching [25], because the intense heat leads to the drying and intense whitening of these and, thus, the presence of the nest may be revealed. In the animal world, white as well as red has signaling significance. This behaviour was not observed in this case. The female has consumed most of the shells in order to restore the calcium level in blood.

Roles of the parents during the fledging period. In 2000, this period lasted for 65 days and for 64 days in 2001. Regular observations were done only in 2000 for a period of 34 days (53.30% of the total). Depending on the conditions the duration of fledging period can vary between 64 and 70 days [2, 7, 20].

In this period the roles of the two adults became distinct, being superposed only to a small extent. The male provided food for the entire family in the proportion of 85.24%. The female hunted for the first time 35 days after hatching, her contribution to providing food being of 14.76%. Warming and feeding the hatchlings were provided only by the female.

Most of the preys were captured and brought to the nest in the first half of the day and their number varied between one and three preys/day. The maximum number of preys was brought to the nest in the middle of the fledging period (Fig. 7).

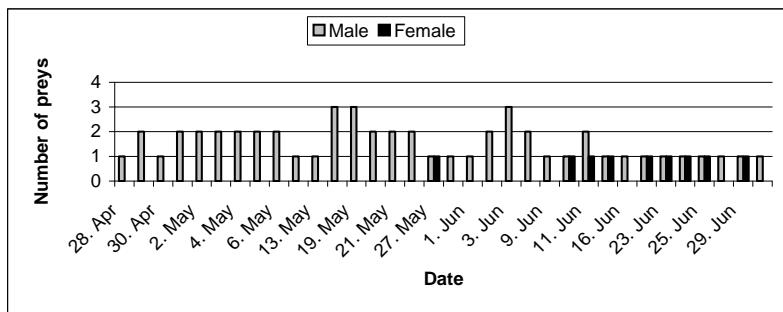


Fig. 7. Number of preys captured by the adults in 2000.

The female participated actively to providing food only after the hatchling has achieved the age of 35 days, when the food brought by the male was not satisfying any more the nutritional requirements of the hatchling. In this period the hatchling had its plumage already formed so it was not necessary any more to be warmed by the female. Until this moment, but to same, even later, the female fed from the prey captured and brought by the male.

The active participation of the female to providing food in the second stage of fledging period also explains the accentuated dimorphism in favour of the female. The smaller size of the male confers him greater agility, so that he is capable of capturing more abundant small-sized prey. The larger size of the female deprives her of the agility of the male and does not confer her the possibility of capturing this type of prey, but she is capable of hunting large-sized prey, which can provide food to her and hatchling for a longer period [29].

As for the male, the time not allocated to hunting was used in the proportion of about 70% for the grooming behaviour.

The active feeding for the hatchlings were exclusively the duty of female. They were fed for the first time about seven hours after hatching. There is a correlation of inverse proportionality between the number of daily feedings and the average duration of one feeding (Fig. 8).

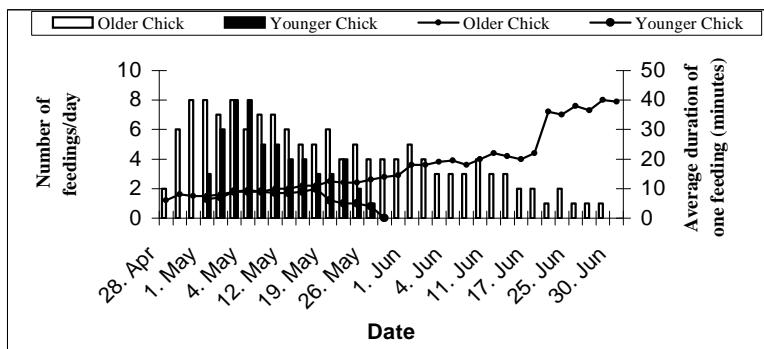


Fig. 8. Number of daily feedings and the average duration of one feeding.

Thus, in the first phase of fledging period, the hatchlings were feed many times (up to 8 feedings/day) and as they grew up, the number of feedings decreased (to 1 feeding/day). On the other hand, the average duration of one feeding was short (6 minutes) at the beginning of the fledging period and it was increasing gradually (until 40 minutes). Thus, one may conclude that as the hatchling advanced in age, the amount of consumed food increased. This aspect was valid for both of the hatchlings until a certain moment. The situation changed about three weeks after hatching of the younger chick. This died of hunger even in the period when the male brought the greatest number of preys to the nest, which means that the principal cause of its death was the strong competition exercised by its older brother. Although the older hatchling was capable of feeding itself about 30 days from hatching, the female continued to feed it until its leaving the nest.

The relationship between the brothers. Right after the hatching of the younger chick, an intense competition begins between the two brothers, materialised by aggressive acts of the older one upon its younger brother. These manifestations were observed for the first time in 2000, on the 2nd of May, when the older hatchling had the age of four days and the younger of one day. The aggressive acts continued with increased frequency close to the moment of the death of the younger hatchling (Fig. 9).

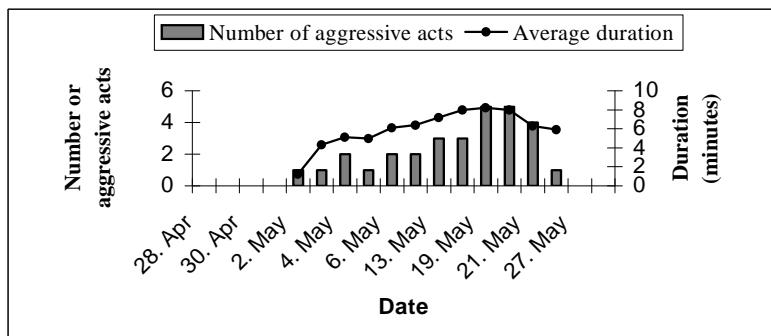


Fig. 9. The relationship between number of aggressive acts and their average duration in 2000.

It has been observed that with ageing the older hatchling manifestations of aggressivity upon the younger one became more and more frequent. The increase in the number of these manifestations was accompanied by the increase of their average duration.

The aggressive acts were observed before as well as during after feeding. The reaction of the younger hatchling always consisted in adopting a state of total immobility. After finishing the acts of aggression the older hatchling was always feed by the female. The younger chick was fed only during the sleeping period of the older one. In many cases its feeding was interrupted by a new series of aggressive manifestation of the older hatchling, which, in order to celebrate its "victory", consumed some more pieces of meat, after which it slept again. The female never intervened in the relationship between the youngs.

Starting with the 21st of May, about six days before the moment of death of the younger hatchling (the 27th of May), the number as well as the average duration of the aggressive manifestations decreased. On the 26th of May, the older hatchling was at the last capable of feeding itself. Although the difference of age between the hatchlings was only three days, the older hatchling was almost double in size.

Although between the years 2001-2003 no more regular observations were made, the aggressive manifestations of the older hatchling were present and they lead to the same final result: the death of the younger hatchling. The exception took place in 2003 when the younger one passed over the critical period and managed to survive.

The result of the aggressions upon the younger hatchling was its gradual physical exhaustion, which lead to its death. The death was caused by starvation, due to the competition exerted by the older brother and not due to lack of food. These results, evidentiate the reproductive strategy of the Golden Eagle. This produces two hatchlings every year only as a measure of safety for the case the older hatchling dies from different reasons (strong attack form various ectoparasites, its falling from the nest) [7, 12, 20].

Breeding success. In each of the four years of observations, the Golden Eagle pair produced 2 hatchlings. Due to the fact that in 2003 both hatchlings survived, the breeding success was of 62.5%. The survival of both hatchlings is extremely rare in the Golden Eagle, few such cases having been cited in literature [19, 30]. In Scotland, between 1964 and 1968, the breeding success was studied for 64 pairs of Golden Eagles and an average of 1.22 hatchlings/nest was fund, which means 61% breeding success.

About 100 days after leaving the nest the hatchlings become fully independent. After this period no juvenile or sub-adult has been seen at the Gorge of Turda.

Conclusions. 1. The breeding season for the Golden Eagle begins at the end of February by the gradual return of the couple to the territory, to which they show a great fidelity, the same couple being studied in each of the four seasons.

2. The vital space of the couple is located in the Gorge of Turda, where the seven nests have been identified. The total surface of the territory, including the two components, was estimated to be about 300 km².

3. After the couple's return to the territory, three nests were rebuilt in 2000 and two in 2001. The role of choosing and rebuilding the nests belonged to the female.

The nests were used in turn; the couple did not brood in two consecutive years in the same nest.

4. Two eggs were laid in each of the four years. In 2000, between the moment of laying the first egg and the second one there was a period of 76 hours, and in 2001 a period of 73 hours.

5. Brooding the eggs started with the laying of the first egg and was provided by both partners, but in different proportions. In 2000, during the day, the female did the hatching in the proportion of 67.53% and in 2001 in the proportion of 70.7%. At night, the female was always the one who hatched. The brooding period has lasted for 42 days in both 2000 and in 2001.

6. Replacing the partners at the nest is ritualized and it is based mainly on optical signals. On the average, the female was replaced at brooding after a period of about 3 hours and 10 minutes and the replacement of the male occurred after 1 hour and 20 minutes.

7. The hatching was phased, the young chick hatched 72 hours after the older one. The fledging period lasted for 65 days in 2000 and 64 days in 2001.

8. During the fledging period the roles of the adults became distinct. The male got involved in providing food for the entire family and the female warmed and fed the hatchlings. The female hunted for the first time only one month after the hatching. The male was never seen feeding or warming the chicks.

9. In the first stage of the fledging period the hatchlings were feed often (8 feedings/day) and for short periods (about 6 minutes/meal). As they progressed in age, they were fed more and more seldom (1 meal/day) for longer and longer periods (up to 40 minutes/meal).

10. Right after hatching of the younger chick, a relationship of competition developed between the two brothers, materialised in aggressive manifestations of the older chick upon the younger one. These manifestations were seen mainly during feeding. Their frequency and duration increased in time resulting finally in the death of the younger hatchling.

11. In three of the four investigated seasons only the older hatchling survived. In the year 2003 both hatchlings survived. The breeding success during the four seasons was 62.5%.

R E F E R E N C E S

1. Bates, J., Moretti, M., *Golden Eagle (Aquila chrysaetos) population ecology in eastern Utah*, "Great Basin Nat.", **54**, 1994, 248-255.
2. Beecham, J., Kochert, M., *Breeding biology of the Golden Eagle in southwestern Idaho*, "Wilson Bull.", **87**, 1975, 506-513.
3. Brooke, M., Kirkhead, T., *The Cambridge Encyclopedia of Ornithology*, Cambridge Univ. Press, Cambridge, 1991.
4. Brown, L. H., Amadon, D., *Eagles, Hawks and Falcons of the World*, McGraw-Hill, New York, 1968.
5. Brown, L. H., Watson, A., *The Golden Eagle in relation to its food supply*, "Ibis", **106**, 1964, 78-100.
6. Ciochia, V., *Păsările clocitoare din România*, Ed. Şti., Bucureşti, 1992.

7. Crampton, S., *Handbook of the Birds of Europe, the Middle East and North Africa: the Birds of the Western Palearctic*, Vol. IV, Oxford Univ. Press, Oxford, 1985.
8. Dementev, G. P., Gladkov, N. A., *Ptitsy SSSR. Tom 1, Khishchnye Ptitsy*, Izd. Sov. Nauka, Moskva, 1951.
9. Dennis, R. H., Ellis, P. M., Brood, R. A., Langslow, D. R., *The status of the Golden Eagle in Britain*, "Brit. Birds", **77**, 1984, 592-607.
10. Dobrovský, R. R., *Păsările României (Ornis Romaniae)*, Ed. Fundația pentru Literatură și Artă, București, 1946.
11. Fould, H., Craighead, J., *Food requirement of the Golden Eagle*, "Auk", **75**, 1958, 312-317.
12. Gutz, B., Bauer, K., Bezel, E., *Handbuch der Vögel Mitteleuropas*, Vol. 4, Akad. Verlagsges., Wiesbaden, 1975.
13. Gordon, S., *The Golden Eagle, King of Birds*, Collins, London, 1955.
14. Hagemeijer, J., Blair, M., *The EBCC Atlas of European Breeding Birds. Their Distribution and Abundance*, Poyser, London, 1997.
15. Heath, M., Borggreve, C., Petit, N., *European Birds Populations. Estimates and Trends*, Bird Life Conservation Series, Cambridge, 2000.
16. Kohl, S., *Date ornitologice din Cheile Turzii*, "Rev. Muz", **1**, 1969, 79-80.
17. Linția, D., *Păsările din R. P. R.*, Vol. II, Ed. Acad. R. P. R., București, 1954.
18. Lorenz, K., *Betrachtungen über das Erkennen der arteigenen Triebhandlungen der Vögel*, in Über tierisches und menschliches Verhalten, München, 1932.
19. Marquis, M., Ratcliffe, D., Roxburgh, R., *Breeding success and diet of Golden Eagle in southern Scotland in relation to change in land use*, "Biol. Conserv.", **34**, 1985, 121-140.
20. McGrady, M., *The Golden Eagle*, Birds of Western Palearctic Update, **1**, 1997, 99-114.
21. Munteanu, D., *Atlasul păsărilor clocitoare din România*, Publ. Soc. Ornitol. Rom., Cluj-Napoca, 2002.
22. Newton, I., *Population ecology of raptors*, Buteo Books, Vermillion, South Dakota, 1979.
23. Nyáryady, E., *Monografia Cheia Turzii*, Imprimeria Națională, București, 1937.
24. Răsinaru, R., *Contribuții la cunoașterea avifaunei Cheilor Turzii*, "An. Bucovinei", **2** (1), 1995, 159-165.
25. Stănescu, D., *Etograma speciei Recurvirostra avosetta L. 1758 pentru perioada de clocit*, "Stud. Comun. – Șt. Nat.", **23**, 1976, 367-381.
26. Şenchea, N., *Cheile Turzii*, "Rev. Șt. V. Adamachi", **24** (2), 1938, 109-112.
27. Tjernberg, M., *Spacing of Golden Eagle Aquila chrysaetos nests in relation to nest site and food availability*, "Ibis", **127**, 1985, 250-255.
28. Watson, A., Payne, S., Rae, S., *Golden eagles Aquila chrysaetos: land use and food in northeast Scotland*, "Ibis", **131**, 1989, 336-348.
29. Watson, J., *The Golden Eagle*, Poyser, London, 1997.
30. Watson, J., Rae, S., Stillman, R., *Nesting density and breeding success of Golden Eagles Aquila chrysaetos in relation to food supply in Scotland*, "J. Animal Ecol.", **61**, 1992, 543-550.
31. Yatman, L., *Atlas des oiseaux nicheurs de France de 1970 à 1975*, Ministère de la Qualité de la Vie Environnement, Paris, 1976.
32. *** 1979 – Directive 79/409/EEC on the conservation of wild birds. Addressed to the Member States, Luxembourg, 2 April 1979.
33. *** 1993 – Legea nr. 13 pentru aderarea României la Convenția privind conservarea vieții sălbaticice și a habitatelor naturale din Europa, adoptată la Berna la 19. Septembrie 1979, Monit. Oficial al României, **62**/25.03.1993.
34. *** 2001 - Legea 462 pentru aprobația Ordonanței de urgență a Guvernului nr. 236/2000 privind regimul ariilor naturale protejate, conservarea habitatelor naturale, a florei și faunei sălbaticice, Monit. Oficial al României, **433**/02.08. 2001.

PEROXIDASE ACTIVITY, STOMATAL CONDUCTANCE AND CARBOHYDRATE METABOLISM IN SOYBEAN PLANTS UNDER HEAVY METAL STRESS

SILVIA ONAC^{*1}, VICTORIA CRISTEA^{*3}, MARTIN KEUL^{**},
DANA BÁTHORY^{**} and MIHAI TRIFU^{*2}

SUMMARY. – The influence of heavy metals from mine spoils (Cavnic, Baia Mare mining area, Romania) on the guaiacol peroxidase (GPOD) activity, stomatal conductance and the content of carbohydrates in soybean leaves was studied. Soybean plants (*Glycine max* cvs. Agat and Diamant) were grown under field conditions in the following variants: control (unpolluted soil), 50% spoils+50% unpolluted soil, 100% spoils. The activity of GPOD increased in the leaves of both cvs. of soybean plants grown on 100% spoils and decreased in the plants of both cvs. cultivated on 50% spoils. Plants grown in spoil variants showed a decrease in stomatal conductance. The decrease was higher in plants from 100% spoil variant for both cultivars. For both cvs., the content of reducing carbohydrates and disaccharides increased in plants cultivated on 50% spoils and decreased in plants from 100% spoils, as compared to control, whereas the amount of starch diminished in plants from 50% spoil variant and increased in plants from 100% spoils, as compared to control. The results obtained suggest that heavy metals from mine spoils can induce oxidative stress in soybean plants, affecting the plant metabolism.

Reactive oxygen species (ROS) are generated in normal cell metabolism and their regulation is a common cellular event, but under stress conditions (high and low temperatures, high light intensities, UV irradiation, drought and salt stress, exposure to pollutant gases, herbicides, metal toxicity, mechanical and physical stress, pathogen attack, senescence) their production is increased, resulting in oxidative damage [8, 21]. Besides exacerbating cellular damage, ROS can act as signal molecules during the acclimation of plants to stress, mark the formation of tracheary element, lignification and other cross-linking processes in the cell wall [8, 34, 40]. At low concentrations, ROS induce defence genes and adaptative responses, but at high concentrations, cell death is initiated [40].

^{*1} Babeş-Bolyai University, Department of Experimental Biology, University Library “Lucian Blaga”, 400084 Cluj, Romania. E-mail: sonac@bioge.ubbcluj.ro

^{*2} Babeş-Bolyai University, Department of Experimental Biology, 400084 Cluj, Romania

^{*3} Babeş-Bolyai University, “Alexandru Borza” Botanical Garden, 400015 Cluj, Romania

^{**} Biological Research Institute, 400015 Cluj, Romania

One of the causes of heavy metals toxicity in plants is the production of ROS [2, 9, 15, 19, 35, 44], which can react with lipids, proteins, pigments and nucleic acids. ROS are involved in senescence of cells and organs such as leaves and the appearance of chlorosis and necrosis of leaves as visual symptoms of heavy metal toxicity is often related to elevated levels of harmful oxygen species, which inhibit the export of photosynthates and cause the accumulation of large amount of starch and sucrose in leaves [20].

Heavy metals can induce oxidative stress directly (Cu, Fe) by generating ROS via the Fenton-type reactions or indirectly (Cd, Zn) by producing the degradation of chlorophyll and carotenoids or the inhibition of their biosynthesis [28], leading to disturbances in the electron transport rates of photosystems and, consequently, to the generation of ROS [33], or by inducing lipoxygenase, which produces O_2^- by oxidising NADPH [31].

H_2O_2 is generated in various cellular compartments: chloroplasts, mitochondria, peroxisomes, cytoplasm, at the plasma membrane level or extracellularly in the apoplast [8, 12, 40]. Unlike other reactive oxygen species, H_2O_2 is relatively stable and able to diffuse across cell membranes, allowing the interaction between different scavenging systems, even those located in separate organelles [10]. The toxicity of H_2O_2 itself is relatively weak compared with that of other reactive oxygen species, but in the presence of superoxide, H_2O_2 can generate highly reactive hydroxyl radicals (OH^-) via the metal-catalysed Haber-Weiss reaction. OH^- can potentially react with all biological molecules, and because cells have no enzymatic mechanism to eliminate it, its excessive production leads ultimately to cell death [40, 43]. Thus, the scavenging of H_2O_2 in cells is critical to avoid oxidative damage [43].

Plants possess a range of antioxidative defense systems for detoxification of ROS, including several enzymes and metabolites: superoxide dismutases, which catalyse the dismutation of O_2^- to H_2O_2 , as well as catalases and peroxidases, which convert H_2O_2 to H_2O and O_2 , ascorbate, glutathione, tocopherol, carotenoids, flavonoids etc. [8, 9, 20, 34, 43].

Peroxidases are oxidoreductases that catalyse the oxidation of a diverse group of organic compounds using hydrogen peroxide as the ultimate electron acceptor [32]. They are widely distributed in the plant kingdom, highly specific for H_2O_2 as electron acceptor and less specific for the donors, which can be ascorbate, pyrogallol, guaiacol etc. [24]. Peroxidases are generally considered to be the most sensitive indicators of pollutants in the absence of visible injury. The most commonly observed response of peroxidases to pollution is an increase of their activity [14].

Heavy metals can affect stomatal behaviour, that controls both transpiration and photosynthesis in plants, indirectly by generating ROS, which are involved in the regulation of stomatal movement [17, 21, 33, 45], or directly by the action of toxic metals at the guard cell level [30].

The purpose of this study was to investigate the effects of the heavy metals from mine spoils on the peroxidase activity, stomatal conductance and the content of carbohydrates in the leaves of soybean plants.

Materials and methods. The plant used in the present study was soybean, *Glycine max* (L.) Merrill. The plants of two soybean cultivars, Agat and Diamant, were cultivated under field conditions. The experiment was described in a previous paper [29]. The activity of guaiacol peroxidase (GPOD) was determined according to Bouche et al. [3]. The extraction of peroxidase was carried out by grounding 500 mg of fresh leaves with quartz sand, in the presence of K-phosphate buffer (pH 6.1). The extract was centrifuged for 30 minutes at 5000 rpm. For the enzymatic determinations, the supernatant was used. The formation of tetraguaiacol was monitored spectrophotometrically (UNICAM SP1800), at 420 nm and 45 seconds after the initiation of reaction. The activity of GPOD was expressed as $\mu\text{g g}^{-1}$ fresh weight. The stomatal conductance was measured with a porometer (AP4 Delta T, Cambridge) and expressed as mmol water vapour $\text{m}^{-2}\text{s}^{-1}$. Carbohydrates were determined spectrophotometrically (UNICAM SP1800), at 660 nm [26, 36]. Reducing carbohydrates were determined from an aqueous extract and disaccharides were determined by the hydrolysis of this extract with HCl 2% for 5 minutes at 68-70°C. Starch content was obtained by the hydrolysis of dried vegetal material with HCl 1N for one hour at 100°C. The results were expressed as g 100 g^{-1} dry weight.

Results and discussion. The results presented in Fig. 1 show that the activity of guaiacol peroxidase (GPOD) increased in the leaves of soybean plants grown on 100% spoils by 18 and 68% of control in cv. Agat and Diamant, respectively, whereas in the plants of both cvs. cultivated on 50% spoils it decreased by 16 and 48% of control in cv. Agat and Diamant, respectively.

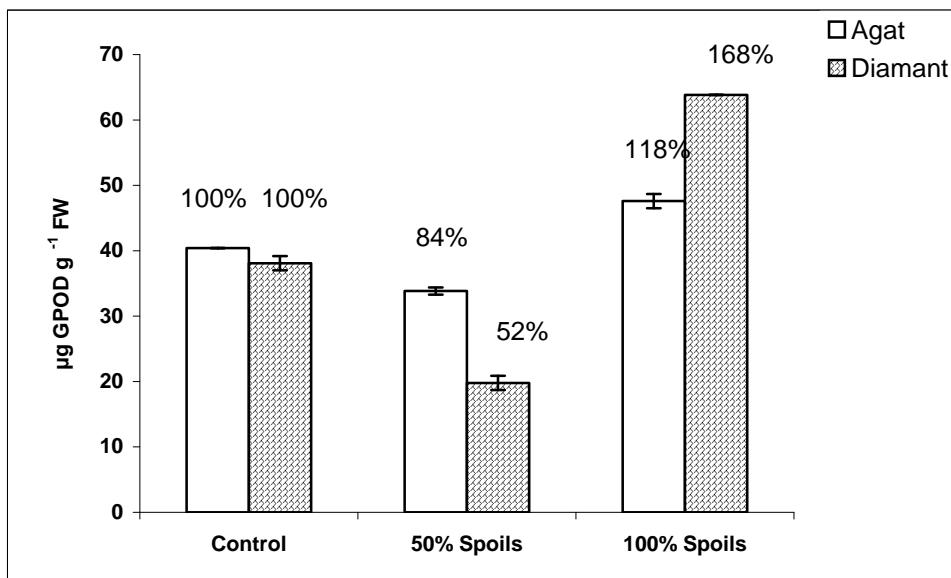


Fig. 1. Guaiacol peroxidase activity in leaves of two soybean cvs., Agat and Diamant, as influenced by the heavy metals from spoils (mean values \pm SD, n=3).

With reference to an increase or a decline in the activity of GPOD in plants treated with heavy metals, variable responses have been reported [6, 19, 33, 38], likely due to the experimental conditions, which strongly differ from one to another. W e c k x and C l i j s t e r s [41] found that the activity of peroxidase was enhanced as a result of Cu toxicity, with a parallel decrease in the H₂O₂ level in the leaves of *Phaseolus vulgaris*. Z a c c h i n i *et al.* [44] reported an increase of peroxidase activity in maize callus treated with lead. H₂O₂ can induce defence or stress-related genes [10] and may play a signalling role during the acclimation of plants to stress [8, 25, 40]. The stimulation of peroxidase capacity in soybean plants grown on 100% spoils indicates increased levels of H₂O₂ generated by heavy metal toxicity and an attempt of plants to protect themselves against oxidative stress. However, the cellular antioxidative system of these plants might not have been completely successful since their growth and production of pods and seeds were reduced [27].

The increased activity of GPOD in plants cultivated on 100% spoils could be an indication of enhanced senescence caused by heavy metals [18, 33]. The decreased activity of GPOD in plants grown on 50% spoils might suggest not a weak antioxidative system, but the production of small amount of H₂O₂ in these plants, because their growth was stimulated compared to the plants cultivated on 100% spoils and even to control [27]. However, this is only a speculation, since the amount of H₂O₂ was not determined in our present study.

The peroxidase activity is higher in the leaves of cv. Diamant on 100% spoils as compared to cv. Agat and this could represent an appropriate protection against overproduction of peroxides [22] caused by the higher accumulation of heavy metals [29]. As in the case of other studied parameters [27], cv. Diamant appeared to be more tolerant to heavy metal toxicity and/or nutrient deficiency from mine spoils.

Y a m a s a k i and G r a c e [42] suggested that cations like Zn, Al and Cd could promote oxidative damage due to stabilization of phenoxy radicals, which are formed by peroxidase-catalysed reaction between phenolics and H₂O₂ [37] and are toxic to living systems.

Heavy metals affect the stomatal physiology. Plants grown in spoil variants showed a decrease in stomatal conductance (Fig. 2). The decrease was higher in plants from 100% spoil variant for both cultivars. The leaf conductance in plants of Diamant cv. was lower as compared to that of cv. Agat in all the cultivation variants.

S a n d a l i o *et al.* [33] found that Cd stimulated stomata to close. Stomatal conductance may be decreased by a direct interaction of the toxic metals at the guard cell level [30]. P e r f u s - B a r b e o c h *et al.*[30] reported that Cd reduced stomatal conductance, leading to the decrease of CO₂ uptake, and did not significantly affect the photosynthetic apparatus, suggesting that Cd perturbed stomatal regulation and limited photosynthesis by CO₂ diffusion at low leaf conductance resulting from stomatal closure. The same authors concluded that the stomatal closure induced by Cd was independent of the abscisic acid (ABA)-related process and Cd did not block the K⁺ channels in guard cells, but the Ca²⁺ channels in the plasma membrane of the

guard cells are highly permeable to Cd, suggesting that Cd could enter guard cells through Ca^{2+} channels, resulting in a perturbation of the intracellular calcium level and calcium signalling, a key element in guard cell osmoregulation.

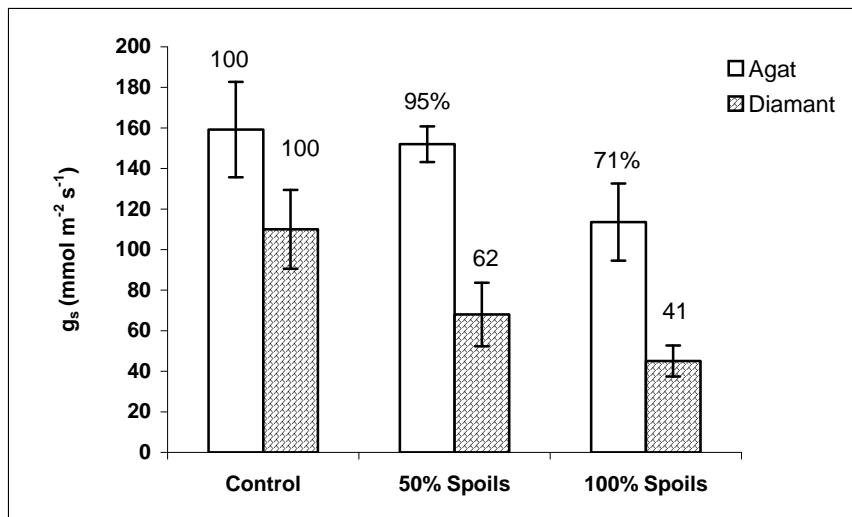


Fig. 2. Stomatal conductance in leaves of two soybean cvs., Agat and Diamant, as influenced by the heavy metals from spoils (mean values \pm SD, n=3).

Stomatal closure may be a consequence of toxic effects of heavy metals in various plant tissues, leading to decreased water availability in leaves and, finally, to decreased stomatal regulation [30]. Heavy metals inhibit the water transport to shoots, causing a water deficit in plants [7]. Under water stress conditions, ABA accumulates in leaf tissues, generating a net loss of guard cell turgour that leads to stomatal closure, thus reducing water loss through transpiration [13]. Z h a n g *et al.* [45] reported that H_2O_2 was involved in the stomatal closure induced by abscisic acid. The findings in our study, namely the increased activity of guaiacol-peroxidase, indicating a high amount of H_2O_2 , and the decrease of stomatal conductance in plants grown on 100% spoils, suggest that ABA, the production of which is increased in plant tissues during heavy metal exposure [11], could have induced the stomatal closure in the leaves of soybean plants.

An increase in ABA content alters the metabolism of indole-3-acetic acid (IAA) and greatly decreases the content and proportion of free IAA, leading to a decrease in cell wall extension [20]. Guaiacol peroxidase is inactivated by IAA [1]. Enhanced synthesis and higher levels of ABA in roots and shoots are a typical response to water deficiency and also to nitrogen and phosphorus deficiency [20]. In our study, the activity of GPOD increased in plants grown on 100% spoils and also P [29] and N (data not shown) levels in 100% spoils were very low, suggesting that ABA might be involved in decreasing the stomatal conductance and also in the reduction of leaf expansion.

M c A i n s h *et al.* [21] and Z h a n g *et al.* [45] reported that H₂O₂ inhibited stomatal opening and promoted stomatal closure, in a concentration-dependent manner. At low concentrations, H₂O₂ induced changes in stomatal aperture that were reversible and due to increases in cytosolic free Ca²⁺ from guard cells induced by the oxidative stress. At high concentrations, the changes in stomatal aperture were irreversible, indicating a reduction in membrane integrity and guard cell viability. An increase in cytosolic free Ca²⁺ concentrations depolarises the plasma membrane, decreasing membrane potential and enhancing K⁺ efflux, thus leading to loss in turgour of the guard cells and stomatal closure [20].

Stomatal closure is associated with K⁺ efflux across the plasma membrane, which is preceded by an increase in cytosolic free Ca²⁺ and cytosolic alkalinisation in the guard cells, which occur in response to ABA [16]. Heavy metals and peroxide induce K⁺ leakage which may occur through K⁺ channels or through nonspecific lesions in the cell membrane caused by lipid peroxidation [23, 41]. However, K ö h l e r *et al.* [17] reported that H₂O₂ and ABA activated the same Ca²⁺ channel and in a similar manner, increasing the cytosolic free calcium, but H₂O₂ inhibited the outward-rectifying K⁺ channels, while ABA activated them, suggesting that ABA and H₂O₂ pathways diverge further downstream in their actions on the K⁺ channels and, thus, on stomatal control.

Stomatal closure inhibits the ability of plants to carry out photosynthesis; therefore a tolerance of plants to water stress can be an important factor affecting growth. The reduction in stomatal conductance accompanied by decreased levels of chlorophyll content in soybean leaves grown on 100% spoils [28] could cause a reduction of leaf photosynthesis and, consequently, a decrease in plant growth [27].

Heavy metals could also affect carbohydrate metabolism in plants. Carbohydrates are the primary products of biosynthesis processes in plants, out of which all the other substances are synthesised. They represent the source of energy and carbon needed for adaptative and/or defensive responses of plants to stresses. Sucrose is the main form in which the carbohydrates are transported in most species of vascular plants. It has an important role in protecting cells from water stress, participating in osmoregulation or protecting the macromolecules or membranes [4, 5, 39].

The content of carbohydrates in the leaves of soybean plants (Fig. 3) presented a similar evolution in both cvs., having close values for all the variants, except for the starch in plants of 50% spoil variant, the level of which decreased in a higher percentage in cv. Agat as compared to cv. Diamant. For both cvs. the content of reducing carbohydrates and disaccharides increased in plants cultivated on 50% spoils and decreased in plants from 100% spoils, as compared to control, whereas the amount of starch diminished in plants from 50% spoils and increased in plants from 100% spoils, as compared to control.

The increase of the amount of disaccharides (sucrose) in leaves of soybean plants grown on 50% spoils could be ascribed to the adaptation tendency of carbohydrate metabolism to the stress caused by heavy metals present in the cultivation substratum. Plants on 50% spoils tried to cope with this stress by compensatory,

energy-consuming mechanisms, the increase of the required energy inducing the synthesis of sucrose. Plants grown on 100% spoils had a decreased content in soluble sugars as compared to the control, being unable to adapt to the unfavourable growth medium. The results regarding the carbohydrate metabolism are in accordance with the morphological observations. The growth of soybean plants from 100% spoils was strongly inhibited, whereas the growth of plants from 50% spoils was stimulated [27]. As for other studied parameters [27], the higher disaccharide content in the leaves of Diamant cv. plants grown on 50% and 100% spoils, as compared to cv. Agat, indicates a better resistance of this cultivar to the stress represented by heavy metals.

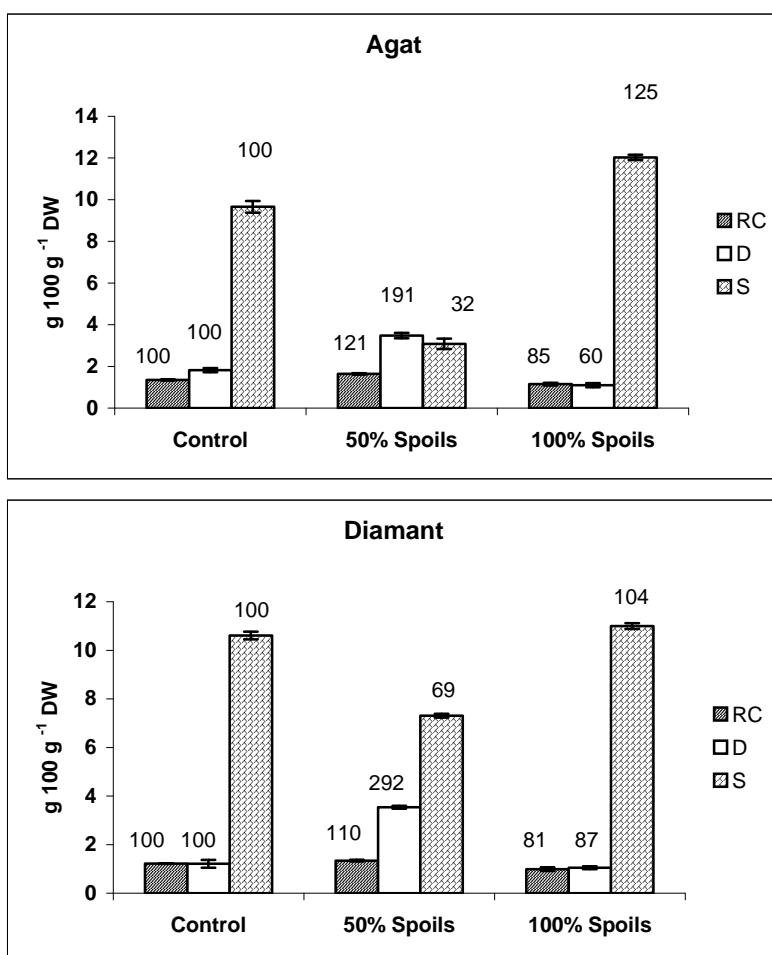


Fig. 3. The content of reducing carbohydrates (RC), disaccharides (D) and starch (S) in leaves of soybean plants, Agat and Diamant cvs., grown on mine spoils (mean values \pm SD, n=3).

High levels of toxic oxygen species cause premature leaf senescence by inhibiting the export of photosynthates, leading to the accumulation of large amounts of starch and sucrose in the source leaves [20].

Reduction of root growth strongly reduces shoot growth by limiting the leaf expansion and lateral stem formation. Simultaneously, carbohydrates (starch in particular) accumulate in the leaves and the carboxylation efficiency of photosynthesis decreases [20]. This could be one of the explanations for the high levels of starch in the leaves of soybean plants grown on 100% spoils, where the elongation of roots was reduced by half when compared to the control [27] and the leaf expansion was decreased (data not shown). The content of phosphorus in 100% spoils is also very low [29] and one of the typical features of phosphorus deficiency is the accumulation of large amounts of starch in the chloroplasts [20].

Conclusions. 1. The activity of guaiacol peroxidase increased in the leaves of soybean plants, Agat and Diamant cvs., grown on 100% spoils and decreased in the plants of both cvs. cultivated on 50% spoils.

2. Soybean plants grown in spoil variants showed a decrease in stomatal conductance. The decrease was higher in plants from 100% spoil variant for both cultivars.

3. Heavy metals from mine spoils affect carbohydrate metabolism in soybean plants, causing an increase in the content of reducing carbohydrates and disaccharides in plants cultivated on 50% spoils and a decrease in plants from 100% spoils, as compared to control, whereas the amount of starch diminished in plants from 50% spoil variant and increased in plants from 100% spoils, as compared to control.

4. By mixing the mine spoils with soil in equal percentages, the noxious action of heavy metals on soybean plants was diminished.

5. The Diamant cv. plants appeared to be more resistant to the stress represented by heavy metals from spoils, when compared to the Agat cv. plants.

R E F E R E N C E S

1. Asada, K., *Ascorbate peroxidase - a hydrogen peroxide-scavenging enzyme in plants*, "Physiol. Plant.", **85**, 1992, 235-241.
2. Bonnert, M., Camara, O., Vessiere, P., *Effects of zinc and influence of Acremonium lolii on growth parameters, chlorophyll a fluorescence and antioxidant enzyme activities of ryegrass (Lolium perenne L. cv. Apollo)*, "J. Exp. Bot.", **51**, 2000, 945-953.
3. Bouchet, M., Joassin, L., Schwindt, F., Gaspar, T., *Effets à court et à long terme d'une variation de température et d'un traitement chimique sur les peroxydases de la betterave sucrière*, "Med. Fac. Landbouww. Rijksuniv. Gent", **42**, 1977, 1671-1679.
4. Bray, E. A., *Plant responses to water deficit*, "Trends Plant Sci.", **2**, 1997, 48-54.
5. Burzó, I., Tomá, S., Dobrescu, A., Ungurean, L., Ţefan, V., *Fiziologia plantelor de cultură. Vol. 2: Fiziologia culturilor de câmp*, pp. 239-283, Știința, Chișinău, 1999.
6. Chaooui, A., Mazzoudi, S., Horbal, M. H., El Ferjani, E., *Cadmium and zinc induction of lipid peroxidation and effects on antioxidant enzyme activities in bean (Phaseolus vulgaris L.)*, "Plant Sci.", **127**, 1997, 139-147.
7. Costa, G., Morel, J.-L., *Water relations, gas exchange and amino acid content in Cd-treated lettuce*, "Plant Physiol. Biochem.", **32**, 1994, 561-570.

8. Dat, J., Vandenebeele, S., Vranová, E., Van Montagu, M., Inzé, D., Van Breusegem, F., *Dual action of the active oxygen species during plant stress responses*, "Cell. Mol. Life Sci.", **57**, 2000, 779-795.
9. Dixit, V., Pandey, V., Shyam, R., *Differential antioxidative responses to cadmium in roots and leaves of pea (Pisum sativum L. cv. Azad)*, "J. Exp. Bot.", **52**, 2001, 1101-1109.
10. Durner, J., Klessig, D. F., *Inhibition of ascorbate peroxidase by salicylic acid and 2,6-dichloroisonicotinic acid, two inducers of plant defense responses*, "Proc. Natl. Acad. Sci. USA", **92**, 1995, 11312-11316.
11. Fediuc, E. C., *Aspecte fiziológice și biochimice ale toxicității cadmiului și mecanismelor protective induse la modele experimentale cu importanță în fitoremediere*, Cluj-Napoca, 2002 (Ph.D. Thesis).
12. Foyer, C. H., Lopez-Delgado, H., Dat, J. F., Scott, I. M., *Hydrogen peroxide- and glutathione-associated mechanisms of acclimatory stress tolerance and signalling*, "Physiol. Plant.", **100**, 1997, 241-254.
13. García-Mata, C., Lamattina, L., *Nitric oxide and abscisic acid cross talk in guard cells*, "Plant Physiol.", **128**, 2002, 790-792.
14. Gaspar, T., Penel, C., Thorpe, T., Greppin, H., *Peroxidases 1970-1980 : a survey of their biochemical and physiological roles in higher plants*, Université de Genève, Genève, 1982.
15. Hartley-Whitaker, J., Ainsworth, G., Meharg, A. A., *Copper- and arsenate-induced oxidative stress in Holcus lanatus L. clones with differential sensitivity*, "Plant Cell Environ.", **24**, 2001, 713-722.
16. Irving, H. R., Gehring, C. A., Parish, R. W., *Changes in cytosolic pH and calcium of guard cells precede stomatal movements*, "Proc. Natl. Acad. Sci. USA", **89**, 1992, 1790-1794.
17. Köhler, B., Hills, A., Blatt, M. R., *Control of guard cell ion channels by hydrogen peroxide and abscisic acid indicates their action through alternate signalling pathways*, "Plant Physiol.", **131**, 2003, 385-388.
18. Lee, K. C., Cunningham, B. A., Paulsen, G. M., Liang, G. H., Moore, R. B., *Effects of cadmium on respiration rate and activities of several enzymes in soybean seedlings*, "Physiol. Plant.", **36**, 1976, 4-6.
19. MacFarlane, G. R., Burchardt, M. D., *Photosynthetic pigments and peroxidase activity as indicators of heavy metal stress in the grey mangrove, Avicennia marina (Forsk.) Vierh.*, "Marine Pollut. Bull.", **42** (3), 2001, 233-240.
20. Marschner, H., *Mineral Nutrition of Higher Plants*, 2nd ed., pp. 131-183, Acad. Press, London, 1995.
21. McCain, M. R., Clayton, H., Mansfield, T. A., Hetherington, A. M., *Changes in stomatal behavior and guard cell cytosolic free calcium in response to oxidative stress*, "Plant Physiol.", **111**, 1996, 1031-1042.
22. Meloni, D. A., Oliveira, M. A., Martinez, C. A., Cambräia, J., *Photosynthesis and activity of superoxide dismutase, peroxidase and glutathione reductase in cotton under salt stress*, "Environ. Exp. Bot.", **49**, 2003, 69-76.
23. Murphy, A. S., Eisinger, W. R., Shaff, J. E., Kochian, L. V., Tai, L., *Early copper-induced leakage of K⁺ from Arabidopsis seedlings is mediated by ion channels and coupled to citrate efflux*, "Plant Physiol.", **121**, 1999, 1375-1382.
24. Neamtu, G., Cîmpleanu, G., Socaciu, C., *Biochimie vegetală (partea structurală)*, pp 227-295, Edit. Did. Pedag., București, 1993.
- 25.Neill, S. J., Desikan, R., Clarke, A., Hurst, R. D., Hancock, J. T., *Hydrogen peroxide and nitric oxide as signalling molecules in plants*, "J. Exp. Bot.", **53**, 2002, 1237-1247.
26. Nelson, N. J., *A photometric adaptation of Somogyi method for determination of glucose*, "J. Biol. Chem.", **153**, 1944, 375-376.

27. Onac, S., *The growth and seed production of soybean plants cultivated on mine spoils*, "Stud. Univ. Babeş-Bolyai, Biol.", **47** (2), 2002, 43-53.
28. Onac, S., Bercea, V., Trifu, M., *Accumulation of photosynthetic pigments and proteins in the leaves of soybean plants cultivated on mine spoils*, "Stud. Univ. Babeş-Bolyai, Biol.", **48** (1), 2003, 39-47.
29. Onac, S., Trifu, M., *Accumulation of some heavy metals from mine spoils by soybean plants*, "Stud. Univ. Babeş-Bolyai, Biol.", **47** (1), 2002, 61-70.
30. Perfus-Barbeoch, L., Leonhardt, N., Vavasseur, A., Forestier, C., *Heavy metal toxicity: cadmium permeates through calcium channels and disturbs the plant water status*, "Plant J.", **32**, 2002, 539-548.
31. Quartacci, M. F., Cosi, E., Navarini-Izzo, F., *Lipids and NADPH-dependent superoxide production in plasma membrane vesicles from roots of wheat grown under copper deficiency or excess*, "J. Exp. Bot.", **52**, 2001, 77-84.
32. Quiroga, M., Guerrero, C., Botella, M. A., Barceló, A., Amaya, I., Medina, M. I., Alonso, F. J., Milrad de Forchetti, Tigier, H., Valpuesta, V., *A tomato peroxidase involved in the synthesis of lignin and suberin*, "Plant Physiol.", **122**, 2000, 1119-1127.
33. Sandalio, L. M., Dalurzo, H. C., Gómez, M., Romero-Puertas, M. C., del Río, L. A., *Cadmium-induced changes in the growth and oxidative metabolism of pea plants*, "J. Exp. Bot.", **52**, 2001, 2115-2126.
34. Schützendübel, A., Pöllle, A., *Plant responses to abiotic stresses: heavy metal-induced oxidative stress and protection by mycorrhization*, "J. Exp. Bot.", **53**, 2002, 1351-1365.
35. Schützendübel, A., Schwanz, P., Teichmann, T., Gross, K., Langenfeld-Heyer, R., Godbold, D. L., Pöllle, A., *Cadmium-induced changes in antioxidative systems, hydrogen peroxide content, and differentiation in scots pine roots*, "Plant Physiol.", **127**, 2001, 887-898.
36. Somogyi, M., *Notes on sugar determination*, "J. Biol. Chem.", **195**, 1952, 19-23.
37. Takahama, U., *Redox state of ascorbic acid in the apoplast of stems of Kalanchoë daigremontiana*, "Physiol. Plant.", **89**, 1993, 791-798.
38. Teisseire, H., Guy, V., *Copper-induced changes in antioxidant enzymes activities in fronds of duckweed (Lemma minor)*, "Plant Sci.", **153**, 2000, 65-72.
39. Todaka, D., Matsushima, H., Morohashi, Y., *Water stress enhances β -amylase activity in cucumber cotyledons*, "J. Exp. Bot.", **51**, 2000, 739-745.
40. Vranová, E., Inzé, D., Van Breusegem, F., *Signal transduction during oxidative stress*, "J. Exp. Bot.", **53**, 2002, 1227-1236.
41. Wecx, J. E. J., Clijsters, H., *Oxidative damage and defense mechanisms in primary leaves of Phaseolus vulgaris as a result of root assimilation of toxic amounts of copper*, "Physiol. Plant.", **96**, 1996, 506-512.
42. Yamashaki, H., Gräfe, S. C., *EPR detection of phytophenoxyl radicals stabilized by zinc ions: evidence for the redox coupling of plant phenolics with ascorbate in the H_2O_2 -peroxidase system*, "FEBS Lett.", **422**, 1998, 377-380.
43. Yamashaki, H., Sakihama, Y., Ikeharra, N., *Flavonoid-peroxidase reaction as a detoxification mechanism of plant cells against H_2O_2* , "Plant Physiol.", **115**, 1997, 1405-1412.
44. Zucchini, M., Rea, E., Tullio, M., de Agazio, M., *Increased antioxidative capacity in maize calli during and after oxidative stress induced by a long lead treatment*, "Plant Physiol. Biochem.", **41**, 2003, 49-54.
45. Zhang, X., Zhang, L., Dong, F., Gao, J., Galbraith, D. W., Song, C.-P., *Hydrogen peroxide is involved in abscisic acid-induced stomatal closure in Vicia faba*, "Plant Physiol.", **126**, 2001, 1438-1448.

UPTAKE OF HEAVY METALS BY MAIZE (*ZEA MAYS*) PLANTS CULTIVATED ON MINE SPOILS

DORINA PODAR*, CRISTINA DOBROTĂ* and MIHAI TRIFU*

SUMMARY. – The uptake of heavy metals (Cd, Cu, Pb and Zn) by maize (*Zea mays*) plants cultivated, under field conditions, on mine spoils from Cavnic (Baia Mare) was studied. The experiment, consisting of 4 treatments: control (agricultural soil) and 3 mixtures of agricultural soil with contaminated spoils from Cavnic, in the following proportions: 10, 20 and 30% spoils, was repeated for three years. The concentration of Cd in the maize organs was under the detection limit of the Atomic Absorption Spectrometry (AAS). The accumulation of Zn and Cu in stems and leaves of maize was in the range of normal values: 20-150 mg Zn kg⁻¹ d. wt. and 5-20 mg Cu kg⁻¹ d. wt., whereas the concentrations of Pb in the leaves of plants treated with 20 and 30% spoils were between 1.2 to 2.2 fold higher than the upper normal limit: 6-8 mg Pb kg⁻¹ d. wt. Lead accumulated mainly in roots, whereas Zn and Cu accumulated both in roots and leaves. In all treatments, the concentrations of Cd, Cu, Pb and Zn in grains of maize were in the range of normal values. The above-ground “green” biomass of maize plants was not severely reduced by the heavy metals from spoils, but the grain yield was significantly diminished, especially in treatments with 20 and 30% spoils. The climatic factors are also able to modify the influence of heavy metals from spoils on the growth and yield of maize plants.

Heavy metals present in the edible seeds may potentially enter the human food chain. These elements retained in the stems, leaves and other parts of the crops are either used as fodder or recycled when the inedible crop residues are returned to the soil. The mobilisation of heavy metals from storage sites in the vegetative organs and their transfer to deposition sites in the reproductive organs (seeds) during plant maturation have been intensively studied [9, 15]. Since the heavy metals rise a serious health risk to living organisms, much research has been centred on how to clean up the heavy metals from the soil and how to eliminate the threat.

The aim of our present study was to investigate whether maize, cultivated on soils contaminated with heavy metals, is capable to accumulate metallic elements under the conditions of maintaining the productivity and reducing the toxicity index to a level that is not affecting the health of human and/or animal consumers.

For this study, mine spoils from Cavnic were selected. At Cavnic, there is a large centre of extraction and processing of the metallic ore, which is a very important source of pollution of the environment, including, of course, the soil, too, due to

* Babeş-Bolyai University, Department of Experimental Biology, 400084 Cluj, Romania.
E-mail: dorina.podar@pcnet.ro

the storing process of the mine spoils in the area [1]. Leaching of heavy metals from the spoils and from the polluted soils into the ground water by rainfall makes the contaminants available for being incorporated into the human food chain. In the Cavnic area, such studies are also important due to the fact that people use the spoils for planting vegetables on rocky substrates. So it is important to establish the amount of heavy metals that the plant species accumulates and the way in which the phytoremediation of the polluted soils may be achieved. Using the plants in the phytoextraction-like activities represents a very convenient way to phytoremediate the soils, especially when the plants have also an economic importance, and the heavy metals do not accumulate in toxic quantities in the edible organs.

Materials and methods. The experiment was carried out using maize plants (*Zea mays* L. c.v. Turda 200), because of its ability to accumulate relatively high levels of toxic metals [11, 12, 14, 19] and also because it has a large biomass and it is one of the most important crops used for human and animal consumption in Romania. The intention was to establish the conditions under which maize can be used as a productive plant and as a plant also involved in the phytoremediation process of the soil.

The experiment was carried out in Turda and consisted of 4 treatments: control (agricultural soil) and 3 mixtures of agricultural soil with spoils in the following proportions: 10, 20 and 30%. These percentages of spoils in the substrates were chosen based on a preliminary laboratory experiment which had shown that they do not significantly affect the above-ground biomass of maize and, therefore, the total uptake of heavy metals from soil would not be reduced because of a low biomass. A treatment with 100% spoils was not performed for this study, because the preliminary experiment showed a 60% inhibition of plant growth and very high heavy metal content in maize organs, which would reduce the phytoremediating potential and/or the possibility of using the above-ground organs for human and/or animal consumption.

The spoils from Cavnic were collected from 0-50 cm depth. The mixed treatments were prepared by removing the agricultural soil to 30 cm depth and mixing it with the appropriate quantity of spoils. After thorough mixing in a cement blender, the mixture obtained was placed back into the field. Each treatment plot was established on 2 m² land. The same experiment was repeated for three years: 2000, 2001 and 2002. The pH of the soil was measured in water (1:2.5), each year, at the beginning of the experiment (Table 1). The total heavy metal concentrations in the agricultural soil and in the mixed treatments are specified in Table 2. Length of shoots of plants was measured at 20, 30, 45, 60 and 75 days after germination. Two plants per treatment were harvested each month for heavy metal content determination. The plants were separated into roots, shoots, leaves and grains. A composite sample was taken after thorough mixing of each organ from the two plants collected for each treatment. Before drying, roots were carefully washed with deionised water to remove any soil adhering to their surface.

UPTAKE OF HEAVY METALS BY MAIZE PLANTS

Table 1
**pH of the substrates, measured in each year at the beginning
of the experiment** (mean \pm SE, n = 3)

	Treatment	pH	
		Control (agricultural soil)	10% Spoils
	20% Spoils	7.64 \pm 0.05	7.68 \pm 0.05
	30% Spoils	7.78 \pm 0.05	

Table 2
**Concentrations of heavy metals (Zn, Pb, Cu and Cd) (mg kg⁻¹ d. wt.)
in the substrates compared to the reference values [18]**

M e t a l	Treatment	Measured values (mg kg ⁻¹ d. wt.)			Normal values	Alert value		Intervention value	
		2000 (ICP-AES)		2001 \pm SD (AAS)		2002 \pm SD (AAS)	Standard land-use	Standard land-use	
		n = 1	n = 4	n = 4		A ^a	B ^b	A ^a	B ^b
Zn	Control	92.32	86.65 \pm 3.90	89.43 \pm 5.47	100	300	700	600	1500
	10% Spoils	208.43	228.19 \pm 8.93	239.56 \pm 14.42					
	20% Spoils	314.90	322.89 \pm 7.12	335.28 \pm 24.56					
	30% Spoils	400.10	415.70 \pm 6.53	418.20 \pm 7.93					
Pb	Control	20.73	22.72 \pm 2.35	23.83 \pm 1.97	20	50	250	100	1000
	10% Spoils	88.34	83.45 \pm 9.31	81.46 \pm 6.67					
	20% Spoils	181.70	172.62 \pm 4.94	174.52 \pm 5.21					
	30% Spoils	289.60	265.50 \pm 6.76	279.40 \pm 6.56					
Cu	Control	22.24	23.82 \pm 2.40	26.82 \pm 4.24	20	100	250	200	500
	10% Spoils	36.85	37.17 \pm 3.33	34.67 \pm 3.68					
	20% Spoils	58.13	53.18 \pm 5.18	51.44 \pm 5.97					
	30% Spoils	71.59	69.98 \pm 4.97	68.21 \pm 3.67					
Cd	Control	0.50	0.49 \pm 0.13	0.47 \pm 0.17	1	3	5	5	10
	10% Spoils	1.27	1.31 \pm 0.05	1.37 \pm 0.03					
	20% Spoils	1.72	1.80 \pm 0.07	1.83 \pm 0.04					
	30% Spoils	2.44	2.39 \pm 0.12	2.36 \pm 0.11					

A^a) – Residential with plant uptake and/or allotment land-use.

B^b) – Commercial/industrial land-use.

On each plot, fifty seeds of maize were sown in two rows. After germination, the seedlings were thinned to 24 on each plot. No fertilizers, pesticides or additional water were added. In each year, the plants were grown in the field, from May to October for an entire vegetative cycle.

Samples of soil and plant material were oven-dried at 75°C to constant weight and then were ground in a stainless steel mill. The concentrations of heavy metals were measured by inductively coupled plasma atomic emission spectrometry (ICP-AES) in 2000 and by flame atomic absorption spectrometry (AAS – AAnalyst Perkin Elmer 100) in 2001 and 2002. For ICP-AES, the soil and plant samples were digested with concentrated nitric acid and hydrogen peroxide and for AAS, their digestion was carried out with concentrated nitric and perchloric acids [23]. Duplicate analysis, reagent blanks and certified reference materials (U.S. National Institute of Standards and Technology- NIST SRM (Standard Reference Material) 2709, 2710 and 2711 – for soil, and spinach leaves – NIST SRM 1570a for plant material) were used for analytical quality. Analytical bias and precision for Cu, Pb and Zn were estimated as being <10%.

Tukey tests were used to compare the differences between mean values.

Results and discussion. The concentrations of heavy metals (Zn, Pb and Cu) in roots, shoots and leaves of maize plants are presented in Table 3. Concentration of Cd was also measured but it was under the detection limit of the AAS - 0.015 mg Cd kg⁻¹ d. wt. The concentrations of heavy metals in maize organs increased with the increase of the level of elements in the substrates (Table 3). The concentrations of metals in the organs of plants cultivated on substrate containing 30% spoils were higher than in those of the control plants. The concentrations were between 1.2-2.7 fold higher at 20 days and between 1.2-3.8 fold higher at the end of the experiment for Zn, between 1.4-13 fold higher at 20 days and between 3-9 fold higher at the end of the vegetative cycle for Pb and between 1.2-2.5 fold higher at 20 days and between 1.3-3.4 fold higher at the end of the vegetative cycle for Cu (Table 3).

Normal concentrations of metals in plants show a large variation between plant species, cultivars, plant tissues and age of plants. However, levels of 20-150 mg kg⁻¹ d. wt. for Zn [6, 13, 20], 5-20 mg kg⁻¹ d. wt. for Cu [1, 17, 19], 0.05-2 mg kg⁻¹ d. wt. for Cd [8] and 0.2-8 mg kg⁻¹ d. wt. for Pb [8, 9, 13] are considered normal. According to these values, the concentrations of metals measured in our study, in the vegetative organs of maize, were all in the range of normal values for Zn and for Cu, except for the roots of plants cultivated on substrate with 30% spoils (Table 3). With regard to Pb, although this element is considered less bioavailable than other heavy metals (Cd, Zn), its concentrations in the roots of plants cultivated on substrates containing spoils were up to 4.5 fold higher than the upper limit considered normal. Moreover, the concentrations of Pb in the leaves of plants from treatments with 20 and 30% spoils were up to 2.2 fold larger than the normal values during 2001 and 2002 (Table 3). These results disagree with those of C a r l s o n *et al.* [4], who found that the amount of Pb in leaves of sunflower and corn treated with up to 500 mg L⁻¹ Pb, as lead chloride, averaged 4 mg L⁻¹ Pb and was not statistically different from the Pb concentration in the control (untreated) plants. The relative high concentrations of Pb found in the leaves of maize plants corroborate with findings of H u a n g *et al.* [11, 12], who concluded that maize could be used for phytoremediation of Pb-contaminated soils.

UPTAKE OF HEAVY METALS BY MAIZE PLANTS

Table 3

Heavy metal concentrations (mg kg⁻¹ d. wt) in the vegetative organs of maize plants cultivated on substrates containing different percentages of mine spoils

Metal	Age of plants (days)	Year	Treatment											
			Control			10% Spoils			20% Spoils			30% Spoils		
			R	S	L	R	S	L	R	S	L	R	S	L
Zn	20	2000	30	54	30	44	100	39	59	92	46	78	107	47
		2001	36	62	36	48	61	39	57	55	40	73	53	43
		2002	48	42	35	86	51	37	114	58	43	130	89	51
	40	2000	24	32	28	29	36	29	35	38	29	41	43	29
		2001	19	63	31	30	69	36	32	71	37	40	78	38
		2002	20	24	31	29	28	33	33	32	38	43	39	38
	70	2000	23	16	30	55	17	32	57	26	36	87	33	39
		2001	17	11	26	24	9	40	27	19	42	32	50	47
		2002	20	11	23	23	15	23	32	27	23	52	32	25
	100	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	18	8.0	11	28	14	18	37	36	29	48	40	33
		2002	20	8.0	23	31	10	25	32	12	31	45	165	38
Pb	130	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	16	10	28	22	11	29	30	11	31	31	15	31
		2002	17	8.0	28	25	11	37	42	13	40	46	24	56
	20	2000	0.9	0.0	0.0	6.5	1.4	0.6	8.8	1.9	1.2	11	2.9	3.5
		2001	4.3	3.3	6.3	17	4.4	7.0	18	5.5	8.3	22	7.5	9.0
		2002	4.6	1.6	2.1	22	2.6	2.7	25	2.8	5.7	35	4.3	6.0
	40	2000	1.3	0.0	0.0	5.0	1.1	1.1	8.2	1.3	1.7	14	1.4	2.1
		2001	5.0	3.0	4.4	12	2.7	5.2	15	3.9	5.6	23	5.0	7.1
		2002	4.3	1.6	5.0	9.6	2.5	2.3	12	3.8	4.8	21	5.0	5.0
	70	2000	2.6	0	0	11	1.8	2.4	14	2.0	2.9	16	3.0	3.1
		2001	4.7	1.1	2.8	14	1.5	4.0	15	1.5	4.7	19	1.7	4.4
		2002	4.6	1.9	2.5	6.2	2.0	2.8	14	3.0	3.0	23	3.6	3.1
Cu	100	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	4.4	0.4	4.0	16	3.8	6.1	23	4.1	13	37	5.8	14
		2002	5.7	1.6	6.8	18	2.4	7.2	17	2.8	8.1	29	3.1	12
	130	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	4.6	1.0	2.9	15	4.2	5.2	21	4.8	8.8	28	5.1	9.6
		2002	3.5	0	2.0	8.2	0.8	6.6	12	0.9	14	26	1.0	18
	20	2000	5.4	4.7	4.9	7.5	5.1	5.4	8.2	5.9	5.4	9.7	6.0	7.0
		2001	9.7	6.8	9.5	15	7.5	11	19	9.2	11	20	11	15
		2002	8.8	5.0	7.0	18	7.0	9.3	19	8.0	9.3	23	8.5	11
	40	2000	6.3	3.6	5.6	6.6	3.7	5.6	8.9	4.4	8.2	10	5.6	10
		2001	9.0	6.6	11	15	7.0	11	19	9.4	13.1	20	11	16
		2002	8.7	3.8	11	10	3.9	10.5	12	4.6	14.8	16	5.2	17
	70	2000	4.5	2.3	9.7	10	2.4	13	14	2.8	13	15	4.2	14
		2001	7.4	3.7	14	14	5.9	14	14	7.9	17	18	10	18
		2002	7.6	2.3	6.2	8.5	2.4	6.2	9.3	5.7	6.3	12	7.0	6.7
	100	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	12	4.2	12	16	7.3	13	20	9.6	16	25	11	19
		2002	6.4	2.2	6.0	9.8	2.4	7.4	12	2.5	11	16	3.5	11
	130	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	8.2	5.7	10	12	6.6	12	16	6.2	15	16	7.4	17
		2002	13	1.8	7.7	17	1.9	15	20	2.8	17	21	3.8	19

R – Roots. S – Shoots. L – Leaves.

Depending on the metal, year-to-year variations in the accumulation of heavy metals in the maize parts were registered. These variations from one year to another were probably caused by the climatic factors, as the same soil, treatments, plant species and location of the experiment were used during each year. Year-to-year variations, due to climate variations, were also found by Brown *et al.* [3] for lettuce. In our experiment, the concentration of Zn in roots and shoots of maize tended to decrease to the end of the vegetative cycle; the most evident decrease was, in general, registered between day 20 and 40 of the experiment (Table 3). This trend was observed previously for maize [25] and was probably due to the diluting effect of the enlargement of plant biomass. In our study, it could also be due to the expansion of the roots down, in the uncontaminated agricultural soil, situated beneath the 30-cm layer of substrate containing spoils.

The decrease in the Zn concentration with the increase of plant age was on average by a factor of 2 and 3 for roots, in 2001 and 2002 and by a factor of 3.0, 5.0 and 4.5 for shoots, in 2000, 2001 and 2002, respectively (Table 3). However, the concentration of Zn in maize leaves, although it followed some ups and downs during the vegetative cycle, was similar at the end of the vegetative cycle comparing to that of 20 days old seedlings and, moreover, it was similar in the three years of the experiment. For Cu and Pb, although their concentrations in the maize parts also tended to decrease between day 20 and 40, but to a lower extent than for Zn, they, in general, increased again after day 70 of the experiment (Table 3).

The results showed that Pb, a non-essential element, was accumulated mainly in roots, whereas Zn and Cu, essential microelements, were accumulated in both roots and leaves (Tables 3 and 4). Previous experiments showed that most plants appear to have natural barriers to translocation of lead from root into stem, leaves and edible fruiting parts [4, 9, 24]. Lead retention within the roots of plants was also observed in our experiment, in which more than 50% of the Pb taken up by maize was found at root level. With regard to Zn, the percentage of this element in roots increased, whereas in leaves it decreased at higher amounts of spoils in the substrates. The increase of the percentage of Zn in roots (up to 21, 9 and 12% in year 2000, 2001 and 2002, respectively) was equivalent with the decrease of its percentage in leaves (up to 20, 9 and 9% in 2000, 2001 and 2002, respectively), showing that in highly contaminated soils, the plants tend to protect themselves by retaining the metals in the roots system [15], the xylem translocation occurring at a small degree (Table 4). The results of our study showed that Cu was relatively equally distributed in both roots and leaves of maize plants, which agrees with the results of Chlopecka [5] and Mantovani *et al.* [17], who also found that this element was intensively translocated from the roots to the above-ground tissues. Depending probably on the climatic factors, the percentage of Cu found in leaves was either not significantly different or it was slightly higher than the percentage of Cu found in the roots (Table 4).

UPTAKE OF HEAVY METALS BY MAIZE PLANTS

Table 4

**Heavy metal concentrations (% of the total amount taken up by plants)
in the vegetative organs of maize plants cultivated on substrates
containing different percentages of mine spoils**

M e t a l	Age of plants (days)	Year	Treatment											
			Control				10% Spoils				20% Spoils			
			R	S	L	R	S	L	R	S	L	R	S	L
Zn	20	2000	26	48	26	25	53	22	30	47	23	34	46	20
		2001	27	46	27	33	41	26	38	36	26	44	31	25
		2002	39	33	28	50	29	21	53	27	20	48	33	19
	40	2000	28	39	33	31	38	31	34	37	29	36	38	26
		2001	16	56	28	22	51	27	23	50	27	26	50	24
		2002	27	32	41	32	31	37	32	31	37	36	32	32
	70	2000	33	23	44	53	16	31	48	21	31	54	21	25
		2001	31	20	49	33	12	55	31	22	47	25	38	37
		2002	37	20	43	38	24	38	39	33	28	48	29	23
	100	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	48	22	30	46	24	30	36	35	29	40	33	27
		2002	39	16	45	47	15	38	43	16	41	46	16	38
	130	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	30	18	52	35	18	47	41	16	43	39	19	42
		2002	32	14	54	34	15	51	44	14	42	36	19	45
Pb	20	2000	100	0	0	76	16	7	74	16	10	64	16	20
		2001	46	19	35	60	15	25	57	17	26	57	20	23
		2002	55	20	25	80	10	10	75	8	17	84	7	9
	40	2000	100	0	0	68	16	16	73	12	15	80	8	12
		2001	41	24	35	60	14	24	61	16	23	66	14	20
		2002	52	20	28	60	15	25	59	18	23	68	16	16
	70	2000	100	0	0	73	12	15	74	11	15	72	14	14
		2001	56	13	32	72	8	20	71	7	22	76	7	17
		2002	52	21	27	57	18	25	70	15	15	77	12	11
	100	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	51	4	45	63	14	23	58	10	32	63	13	24
		2002	41	11	48	65	9	26	61	10	29	67	7	26
	130	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	55	11	34	61	17	22	61	14	25	66	22	22
		2002	64	0	36	53	5	42	45	3	52	58	2	40
Cu	20	2000	36	31	33	42	28	30	39	29	32	43	27	31
		2001	37	26	37	46	23	32	48	23	29	44	24	32
		2002	42	24	34	52	21	27	52	22	26	54	20	26
	40	2000	41	23	36	37	20	43	42	20	38	39	22	39
		2001	34	25	41	43	21	36	46	23	31	43	23	34
		2002	38	16	46	38	14	48	39	14	47	42	14	44
	70	2000	27	14	59	40	10	50	46	10	44	46	13	42
		2001	30	15	55	40	17	43	36	21	43	38	22	40
		2002	47	14	39	50	14	36	44	26	30	47	27	26
	100	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	44	15	41	44	20	36	45	21	34	45	20	35
		2002	44	15	41	50	12	38	48	10	42	54	12	34
	130	2000	-	-	-	-	-	-	-	-	-	-	-	-
		2001	34	24	42	39	22	39	43	17	40	40	18	42
		2002	56	9	35	49	6	45	50	7	43	48	9	43

R – Roots. S – Shoots. L – Leaves.

Highly significant correlation coefficients were obtained between the total concentration of metals (Zn, Pb and Cu) in soil and their concentration in roots (Table 5). For shoots, no significant correlations between the metal concentrations in plants and soil were found, for any of Zn, Pb or Cu, except for Cu in 2002. Similar significant positive correlation between total Cu in soil and its concentration in maize shoots was found by Chlopecka [5]. In disagreement with the data of Chlopecka [5], we found significant positive correlation between the total concentration of Pb in substrate and in the leaves of maize plants (Table 5). However, this positive correlation for Pb might have been falsely significant because the concentrations of Pb in leaves were under the detection limit for control, 10 and 20% spoils (Table 6). The concentrations of metals in grains were positively correlated with their total concentration in soil for Zn and Cu, but not for Pb, which agrees with previous observations on maize [5] and *Silene vulgaris* [7]. Other authors found that effects of soils on the concentrations of Cd, Cu, Pb and Zn were more evident in the leaves than in the grains [15], whereas in a study performed with rice, Herrera et al. [10] observed no correlation between the content of Cd, Cu and Zn in plant tissues and the amount of these metals found in soil.

Table 5
Correlations between the metal concentrations (mg kg⁻¹ d. wt.) in maize organs and in soil at the end of the vegetative cycle

Plant parts	Zn		Pb		Cu				
	2000	2001	2002	2000	2001	2002	2000	2001	2002
Roots	0.98*	0.96*		0.98*	0.97*		0.95*	0.95*	
Shoots	0.89	0.90		0.82	0.82		0.87	0.99*	
Leaves	0.91	0.93		0.96*	0.99*		0.99**	0.87	
Grains	0.97*	0.97*	0.98*	no Pb	0.81	0.97*	0.99**	0.99*	0.98*

* and ** indicate that the correlation coefficients are significant at P < 0.05 and 0.01, respectively.

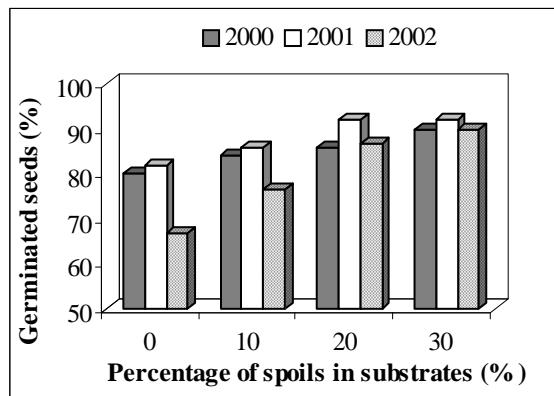
Table 6
Concentrations of Cu, Pb and Zn in grains of maize plants (mg kg⁻¹ d. wt.) at the end of the vegetative cycle

Treatment	Zn		Pb		Cu				
	2000	2001	2002	2000	2001	2002	2000	2001	2002
Control	23.5	26.1	22.4	0.0	0.0	0.0	1.6	1.5	1.6
10% Spoils	28.7	31.7	27.8	0.0	0.0	0.0	2.1	1.9	2.0
20% Spoils	29.4	33.5	28.7	0.0	0.0	0.1	2.6	2.9	2.3
30% Spoils	33.6	35.6	32.9	0.0	0.4	0.2	3.0	3.3	2.8

Distribution of Cd, Cu, Pb and Zn in the seeds and other above-ground parts of crops is of nutritional importance to humans. The grains of crops represent the part most likely to be eaten. The apparent ease with which Zn moves into the grains and the apparently restricted movement of Cd and Pb into seeds could both be important to human health [15].

Grain and seed crops serve as natural barriers to the movement of potentially toxic heavy metals, such as Cd and Pb, into the animal/human food chain, minimising their transfer from soils, while conserving Zn and Cu levels in the edible parts of these crops (Kubota *et al.* [15], working on maize, wheat and soybean). The copper content of maize grains was below the lowest limit for copper contents in plants ($5\text{--}20 \text{ mg kg}^{-1}$ d. wt., which is considered the critical value for plant growth [2, 17, 19]).

The presence of spoils in the substrates had a positive influence on the germination of maize seeds (Fig. 1), slightly stimulating this process. Similar results were also obtained for maize [13] and soybean [21].



F i g. 1. Germination capacity of maize seeds sown on substrates containing different percentages of mine spoils.

The results of our study are in agreement with previous findings which showed a reduced biomass of plants cultivated on polymetal-contaminated substrate [5, 16, 22]. However, the reduction in the length of the shoots of maize plants grown on contaminated substrates compared to control plants, depended largely on the climatic conditions. Thus, during 2000, the driest and the warmest among the three years (data not shown), the length of shoots was significantly smaller in all treatments containing spoils compared to control plants, whereas during 2001, the coolest and the wettest year, there was no difference in the growth of maize plants. Moreover, length of shoots in 2001 was significantly higher than in 2000 and 2002 (Fig. 2). This was due to lower-normal temperatures and to a “normal” regime of rainfalls during 2001, compared to 2000 and 2002. In 2000, stunted growth became obvious in the case of 30 days old plants and it heightened with increasing plant age, whereas in 2002, the differences in the length of shoots were significant only at the end of the vegetative growth period (Fig. 2). In 2000, the length of shoots was reduced by a factor of 1.12 for plants grown on 10 and 20% spoils and by a factor of 1.25 for plants on 30% spoils as compared to control plants, whilst in 2002, the length of shoots of plants from treatments with 20 and 30% spoils was by 1.15 fold smaller than that of the control

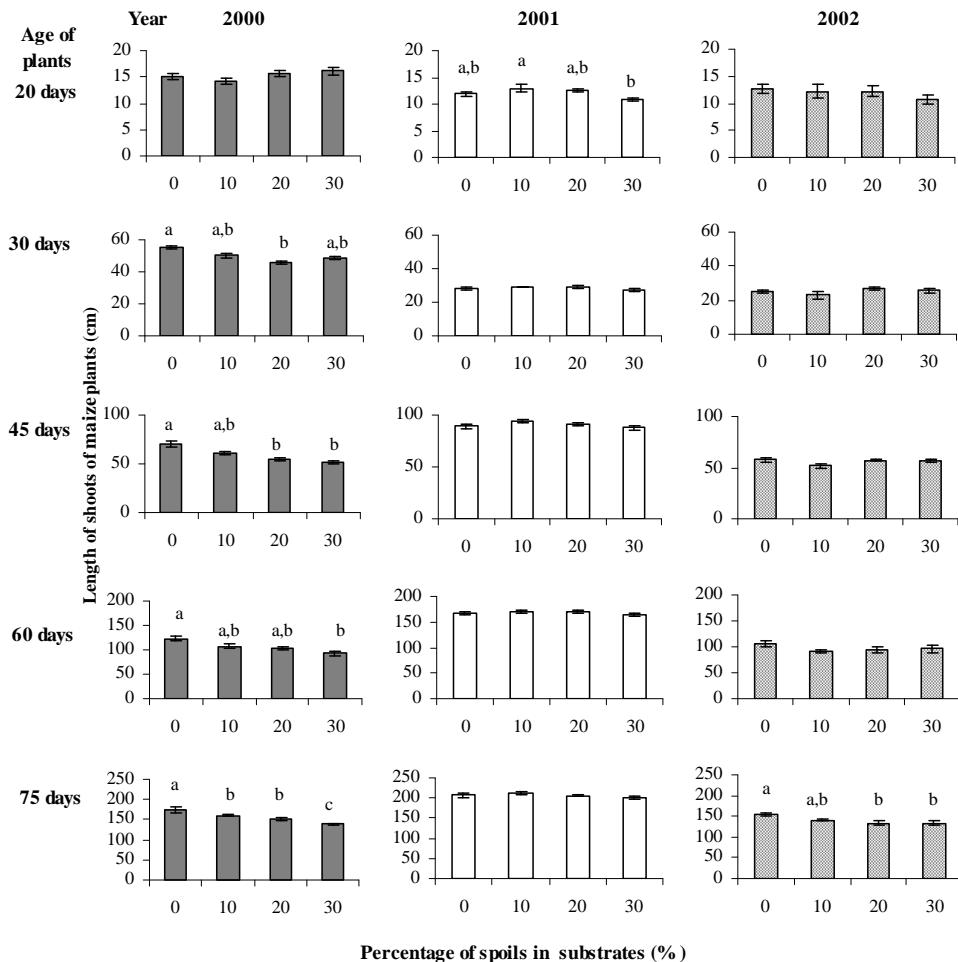


Fig. 2. Length of shoots of maize plants(cm) (mean \pm SE, n = 15). Bars with different letters are significantly different (confirmed by Tukey test). Tukey test was performed separately for each graph; therefore, letters from different graphs are not comparable. When no letters are attached, the means are not significantly different (confirmed by Tukey test).

plants. The statistically significant reduction of growth of maize plants registered in our study was, however, much lower than in other experiments, probably because of the capacity of maize to tolerate relatively high concentrations of heavy metals [11, 12, 14]. Shoot length of alfalfa (*Medicago sativa* L. c.v. Mesa) grown on soil contaminated with 50 mg of each Cd, Cu, Ni and Zn kg⁻¹ d. wt. was on average 2 fold lower than that of the control plants [22]. The authors concluded that the low growth of plants was due to the combined stress caused by the mixture of heavy metals, because the alfalfa

plants have demonstrated the ability to grow well in soils individually contaminated with more than 50 mg kg⁻¹ d. wt. of heavy metals [22]. Similar results were obtained by Chlopecka [5], who showed that Cd, Cu, Pb and Zn added jointly as carbonates, at a rate of 50, 500, 500 and 1500 mg kg⁻¹ d. wt., respectively, reduced the biomass of shoots and leaves of maize by a factor of 2 and the grain yield by a factor of 2.5. Lugo and Rimmer [16] also obtained an average of 10% decrease in the shoot biomass of barley when grown in substrate containing all four metals (Cd, Cu, Pb and Zn) than in treatments with only one, two or three of these metals.

Despite the fact that the growth of shoots was not severely inhibited and that the concentrations of heavy metals in the maize parts were in general in the range of normal values, the grain yield of maize was significantly affected in each year of the experiment. Similarly to the length of the shoots, the grain yield was also influenced by the climatic factors, being more severely reduced under drier and hotter weather (Fig. 3). The reduction in the grain yield in 2000, 2001 and 2002, respectively, expressed as a percentage of the control, was 9.3, 4.9 and 3.8% for the treatment with 10% spoils, 39.3, 10.4 and 16.7% for the treatment with 20% spoils and 65.5, 32.1 and 50.1% for the treatment with 30% spoils. Outstandingly, the concentrations of heavy metals in grains of maize were in the range or even below the normal values (Table 6). Due to the complexity of the soil composition and of the climatic conditions, it was difficult to establish the factors that influenced the grain yield the most. Probably, the high pH of soil and the complex interactions between elements in soil and in plants affected the uptake of macro- and micronutrients, which might have been the cause for the low yield obtained for the treatments containing spoils.

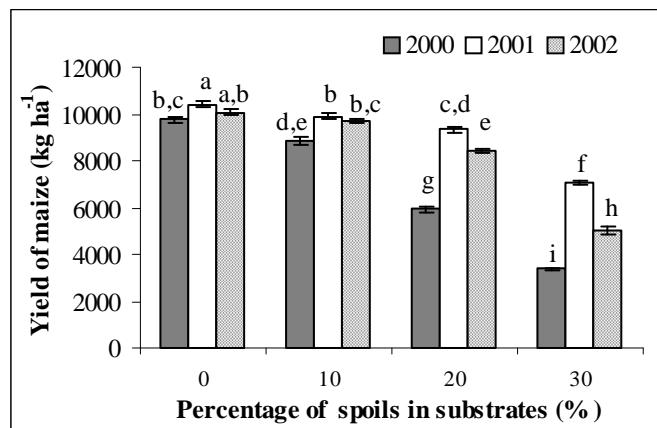


Fig. 3. Grain yield of maize plants (kg ha⁻¹) (mean ± SE, n = 7). Bars with different letters are significantly different (confirmed by Tukey test).

Conclusions. 1. The accumulations of Zn and Cu in stems and leaves of maize grown, under field conditions, on mixtures of agricultural soil with 10, 20 and 30% of spoils from Cavnic, were in the range of normal values: 20-150 mg Zn kg⁻¹ d. wt. and 5-20 mg Cu kg⁻¹ d. wt. Despite this, the above-ground organs of maize plants could not be used to feed the animals because the concentration of Pb in the leaves of treated plants was between 1.2 to 2.2 fold higher than the upper normal limit: 6-8 mg Pb kg⁻¹ d. wt.

2. The concentrations of Cd, Cu, Pb and Zn in the grains of maize were in the range of normal values in all treatments used in our study. Therefore, the grains of maize could be used for animal and/or human consumption.

3. Lead, a non-essential element, accumulated mainly in roots, whereas Zn and Cu, essential microelements, accumulated in both roots and leaves.

4. Mixtures of agricultural soil with 10, 20 and 30% spoils from Cavnic did not severely reduce the above-ground “green” biomass of maize plants, but they significantly diminished the grain yield, especially in the treatments with 20 and 30% spoils. The influence of heavy metals found in the spoils on the above-ground biomass and grain yield of maize was modified by climatic factors, too (*i.e.* by high temperatures and low humidity in soil).

5. Maize can be used for phytoremediation of polymetal-contaminated soils, but its phytoremediating efficiency is relatively low, because of the reduced uptake and translocation of heavy metals from soil to the above-ground organs. Caution should be taken if the above-ground maize material is to be used for animal consumption.

REF E R E N C E S

1. Agenția pentru Protecția Mediului, Baia Mare (APMBM), *Raport privind starea mediului în județul Maramureș*, Ministerul Apelor, Pădurilor și Protecției Mediului (MAPPM), 1999, 2000, 2001, 2002, [www.apmbm.ro/Starea mediului/rapoarte/](http://www.apmbm.ro/Starea%20mediului/rapoarte/).
2. Alloway, B. J., *Heavy Metals in Soils*, Blackie, Glasgow, 1995.
3. Brown, S. L., Chaneey, R. L., Angle, J. S., Ryan, J. A., *The phytoavailability of cadmium to lettuce in long-term biosolids-amended soils*, “J. Environ. Qual.”, **27**, 1998, 1071-1078.
4. Carlson, R. W., Bazzaz, F. A., Rolfe, G. L., *The effect of heavy metals on plants. II. Net photosynthesis and transpiration of whole corn and sunflower plants treated with Pb, Cd, Ni and Tl*, “Environ. Res.”, **10**, 1975, 113-120.
5. Chlopecka, A., *Forms of Cd, Cu, Pb and Zn in soil and their uptake by cereal crops when applied jointly as carbonates*, “Water, Air, Soil Pollut.”, **87**, 1996, 297-309.
6. Choudhary, M., Baille, L. D., Grant, C. A., Leslie, D., *Effect of Zn on the concentration of Cd and Zn in plant tissue of two durum wheat lines*, “Can. J. Plant Sci.”, **75**, 1995, 445-448.

UPTAKE OF HEAVY METALS BY MAIZE PLANTS

7. Ernst, W. H. O., Nelissen, H. J. M., Ten Boom, W. M., *Combination toxicology of metal-enriched soils: physiological responses of a Zn- and Cd-resistant ecotype of Silene vulgaris on polymetallic soils*, "Environ. Exp. Bot.", **43**, 2000, 55-71.
8. Ferguson, J. E., *The Heavy Elements: Chemistry, Environmental Impact and Health Effects*, pp. 377-405, Pergamon, Oxford, 1990.
9. Finster, M. E., Gray, K. A., Binnns, H. J., *Lead levels of edibles grown in contaminated residential soils: a field survey*, "Sci. Total Environ.", **320**, 2004, 245-257.
10. Herawati, S., Herawati, N., Susuki, S., Hayashi, K., Riva, I. F., Koyama, H., *Cadmium, copper and zinc levels in rice and soil of Japan, Indonesia, and China by soil type*, "Bull. Environ. Contam. Toxicol.", **64**, 2000, 33-39.
11. Huang, J. W., Chen, J., Bert, W. R., Cunningham, S. D., *Phytoremediation of lead-contaminated soils: role of synthetic chelates in lead phytoextraction*, "Environ. Sci. Technol.", **31**, 1997, 800-805.
12. Huang, J. W., Cunningham, S. D., *Lead extraction: species variation in lead uptake and translocation*, "New Phytol.", **134**, 1996, 75-84.
13. Keul, M., Vintilă, R., Lazăr - Keul, G., Nicoraru, A., Bercea, V., *Acumularea și fitotoxicitatea unor metale grele la plantule de porumb (Zea mays) crescute pe soluri poluate*, "Contrib. Bot." (Cluj-Napoca), 1991-1992, 189-195.
14. Khan, A. G., Kuek, C., Chaudhry, T. M., Khoo, C. S., Hayes, W. J., *Role of plants, mycorrhizae and phytochelators in heavy metal contaminated land remediation*, "Chemosphere", **41**, 2000, 197-207.
15. Kubota, J., Welch, R. M., Van Campen, D. R., *Partitioning of cadmium, copper, lead and zinc amongst above-ground parts of seed and grain crops grown in selected locations in the USA*, "Environ. Chem. Health", **14**(3), 1992, 91-100.
16. Luo, Y., Rimmer, D. L., *Zinc-copper interaction affecting plant growth on a metal contaminated soil*, "Environ. Pollut.", **88**, 1995, 79-83.
17. Mantovani, P., Bonazzi, G., Mestrini, E., Marimoli, N., *Accumulation of copper and zinc from liquid manure in agricultural soils and crop plants*, "Plant Soil", **250**, 2003, 249-257.
18. Ministerul Apelor, Pădurilor și Protecției Mediului (MAPPM), Ord. MAPPM, nr. 756, Monit. Oficial al României, **303**/06.11.1997.
19. Mocquot, B., Vangronsveld, J., Clijsters, H., Mench, M., *Copper toxicity in young maize (Zea mays L.) plants: effects on growth, mineral and chlorophyll contents and enzyme activities*, "Plant Soil", **182**, 1996, 287-300.
20. Mulligan, C. N., Young, R. N., Gibbs, B. F., *Remediation technologies for metal contaminated soils and groundwater: an evaluation*, "Eng. Geol.", **60**, 2001, 193-207.
21. Onac, S., *The growth and seed production of soybean plants cultivated on mine spoils*, "Stud. Univ. Babeș-Bolyai, Biol.", **47**(2), 2002, 43-53.
22. Peralta-Videa, J. R., Gardeda-Torresdey, J. L., Gomez, E., Tiemann, K. J., Parsons, J. G., Carrillo, G., *Effect of mixed cadmium, copper, nickel and zinc at different pHs upon alfalfa growth and heavy metal uptake*, "Environ. Pollut.", **119**, 2002, 291-301.
23. Thompson, M., Walsh, J. N., *A Handbook of Inductively Coupled Plasma Spectroscopy*, Blackie, Glasgow, 1983.

24. T u n g, G., T e m p l e, P. J. *Histochemical detection of lead in plant tissues*, “Environ. Toxicol. Chem.”, **15**, 1996, 906-914.
25. W e n g e r, K., G u p t a, S. K., F u r r e, G., S c h u l i n, R., *Zinc extraction potential of two common crop plants*, Nicotiana tabacum and Zea mays, “Plant Soil”, **242**, 2002, 217-225.

ASPECTE HISTOLOGICE ȘI ULTRASTRUCTURALE ALE TIMUSULUI ÎN INTOXICAȚIA ACUTĂ CU VENIN DE ALBINE

ADRIAN FLOREA*, CONSTANTIN CRĂCIUN **
și CONSTANTIN PUICĂ ***

SUMMARY. – **Histological and Ultrastructural Aspects of the Thymus in the Acute Intoxication with Honey Bee Venom.** In this study the histological and ultrastructural modifications in the rat thymus were investigated after the experimental administration of very high doses of the bee venom water-soluble fraction. We have taken into account this organ since the honey bee (*Apis mellifera L.*) venom is well known to be a powerful immuno-stimulating agent. First of all, a massive lymphocyte migration in the blood flow was recorded after two hours of the acute treatment. The presence of a high amount of exogenous, toxic substances, as well as the severe alterations occurred in the structure of the most organs and tissues are mainly responsible for this effect. On the other hand, we have also recorded ultrastructural alterations in all cell populations, many cells being totally destroyed; the number of apoptotic cells is increased and the number of mitotic cells is decreased. At the subcellular level, in most of the cells the nuclear envelope was deeply affected and vacuolisations of endoplasmic reticulum were observed. Mitochondria also respond strongly to the bee venom: they appeared to be swollen and with a rarefied matrix. These results indicate both a specific response of the thymus against the action of bee venom molecules, and a direct, unspecific effect of the venom upon all cell membranes leading to the death of many cells.

Timusul este un organ limfoid, de origine epitelială infiltrat cu limfocite și alte celule mezenchimale. El îndeplinește un rol esențial în dezvoltarea aparatului imun, dar și în dobândirea și întreținerea imunității organismului prin formarea, diferențierea și maturarea limfocitelor T la acest nivel. Deși producerea limfocitelor T este specifică și zonelor periferice ale nodulilor splenici sau din alte organe limfopoietice, timusul reprezintă principala sursă pentru acest tip de celule, fiind deci principalul factor responsabil pentru apărarea organismului. Scopul studiului nostru a fost de a urmări reacțiile provocate la nivel structural și ultrastructural în acest țesut la administrarea experimentală a veninului de albine în doze foarte mari, fiind bine cunoscută capacitatea acestui produs biologic de a dezvolta diverse reacții imunologice [2, 8, 17].

*Universitatea de Medicină și Farmacie "Iuliu Hațieganu", Facultatea de Medicină, Catedra de Biologie Celulară și Moleculară, 400349 Cluj-Napoca, România. E-mail: aflorea@personal.ro

**Universitatea "Babeș-Bolyai", Facultatea de Biologie și Geologie, Centrul de Microscopie Electronică, 400006 Cluj-Napoca, România

***Institutul de Cercetări Biologice, 400015 Cluj-Napoca, România

Materiale și metode. *Animale.* Experimentele s-au realizat în două etape pe 4 loturi de *șobolani albi Wistar* (câte 3 animale în fiecare lot), crescuți în biobaza U.M.F. “Iuliu Hațieganu”, fără restricții în ce privește hrana, apa sau mișcarea.

Veninul de albine. Am utilizat venin de albine pur cristalizat, recoltat de noi prin metoda stimulării abinelor cu impulsuri electrice [15, 33]. Cu 24 ore înainte de administrarea experimentală, veninul a fost resuspendat într-o soluție izotonă (150 mM NaCl, 5,5 mM glucoză, 5 mM HEPES, la pH 7,4) și păstrat la frigider, ferit de acțiunea luminii.

Tratament. Animalelor din două loturi – *TL1 și TL2* – li s-a injectat subcutan, dorso-lateral posterior o doză unică, foarte mare de venin de albine (DL₅₀ – letală pentru 50% din subiecți), echivalentă cu cantitatea eliberată la un număr de 100 înțepături de albine (62 mg/kg corp). Probele de țesut recoltate au fost prelucrate pentru realizarea de preparate histologice (*Lotul TL1*), respectiv pentru microscopia electronică de transmisie (*Lotul TL2*). Doza semiletală a fost stabilită în jurul valorii de 60 mg venin/kg corp pe baza datelor din literatură [16, 19]. Pentru evidențierea eventualelor efecte, în paralel s-au recoltat și prelucrat fragmente de timus de la animale din 2 loturi martor – *ML1 și ML2* – coresponzător fiecărui lot de animale tratate.

Histologie. Pentru realizarea preparatelor histologice, piesele biologice recoltate au fost fixate în fixator Bouin, deshidratate, incluse în parafină și apoi secționate la microtom. După etalarea pe lame și deparafinare, colorarea s-a făcut prin metoda Hurduc (pe bază de orange G, albastru de metil, xilidină și acid acetic glacial) [27]. Examinarea s-a realizat la un microscop Olympus BX51.

Microscopia electronică de transmisie. Probele de timus recoltate în vederea studierii la microscopul electronic de transmisie au fost prefixate în soluție de glutar-aldehidă 2,7%, postfixate în acid osmic 2%, spălate și deshidratate în acetonă, apoi incluse în Epon 812 și secționate la un ultramicrotom LKB-III. Secțiunile preluate pe grile electrolitice au fost contrastate cu acetat de uranil și apoi cu citrat de plumb [32]. Examinarea s-a realizat la un microscop electronic de transmisie Jeol JEM 1010.

Rezultate. Examinarea la microscopul optic (Fig. 1), respectiv la microscopul electronic (Fig. 3) a probelor de timus de la loturile martor a relevat structura normală a lobulilor timici, aspectul caracteristic al timocitelor (limfocite T) și al altor tipuri celulare prezente: polimorfonucleare, macrofage, rare monocite, în corticală, respectiv plasmocite, macrofage, puține eozinofile, numeroase celule reticulare și corpusculii Hassal în zona medulară [1, 5, 14, 25, 26, 28, 34-37].

Capilarele sanguine sunt concentrate în special la nivelul medularei, fiind rar observate în corticală. Perivasculare se poate întâlni un spațiu în care ocazional se află fibre de colagen; vasele de sânge nu sunt în contact cu timocitele, între aceste elemente structurale fiind interpusă o barieră formată din celulele reticulare [7, 25, 35].

La examinarea probelor prelevate de la *lotul TL1*, s-a remarcat histologic o reactivitate puternică a glandei, manifestată prin dispariția totală a delimitării între zonele corticală și medulară. Aceste două componente ale timusului se întrepătrund

TIMUSUL ÎN INTOXICAȚIA ACUTĂ CU VENIN DE ALBINE

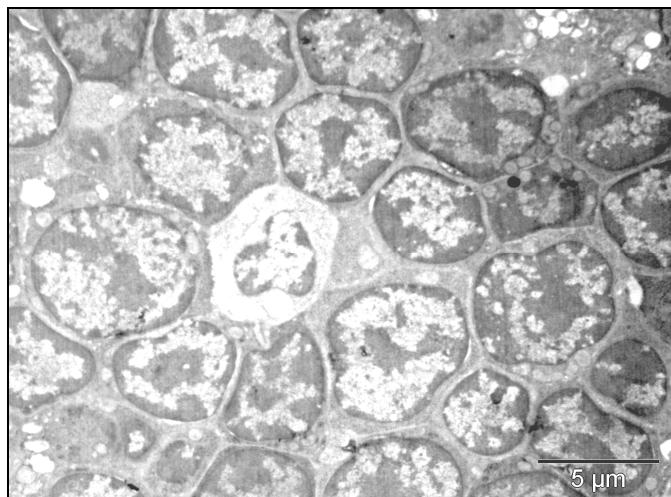
și confluăză. Aspectul morfologic cel mai pregnant îl reprezintă hipoplazia marcată a limfocitelor timice din parenchimul timic, atât din regiunea corticală, cât și din cea medulară. Se remarcă prezența a numeroase spații clare, acelulare, de diferite dimensiuni, rezultate ca urmare a migrării timocitelor din aceste zone ale timusului. Toate aceste modificări contribuie la apariția aspectului de „cer înstelat” al timusului (Fig. 2) care este caracteristic în stările acut sau în unele infecții ale organismului. În medulara lobulilor timici se observă o ușoară hiperplazie a corpusculilor Hassal de formă rotundă sau ovoidală.



F i g. 1. Structura normală a lobulului timic – lotul MLI (ob. 10 ×).

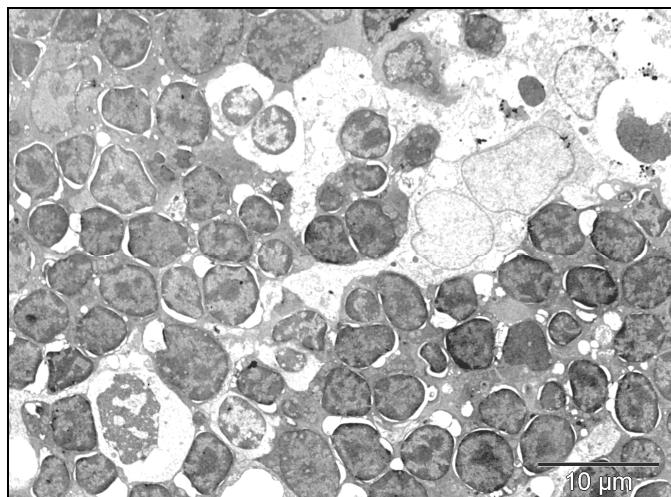


F i g. 2. Aspectul de “cer înstelat” al lobulului timic – lotul TLI (ob. 10 ×).



F i g. 3. Timocite dens împachetate, grupate în „cuiburi” – lotul ML2.

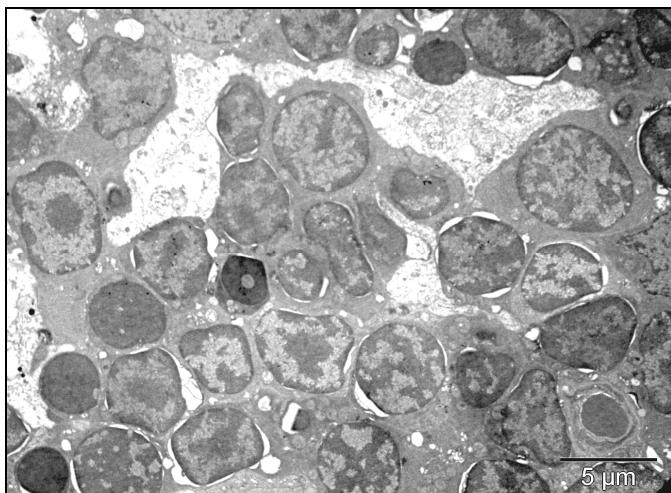
Electronmicroscopic, studiul secțiunilor ultrafine, prin timusul animalelor din *lotul TL2*, evidențiază și în acest caz o densitate scăzută a timocitelor și vacuolizări citoplasmatice de ampoloare ce privesc în primul rând spațiul perinuclear, dar și profilele reticulului endoplasmic (Fig. 4-7). Pe de altă parte, există în numeroase zone celule în care afectarea citoplasmei este totală – aici celulele apar fără conținut citoplasmatic perinuclear (Fig. 4, 6, 7). Acest aspect este observat și în cazul unor celule surprinse în timpul mitozei (Fig. 4), sau chiar în cazul unor celule apoptotice (Fig. 4 și 7).



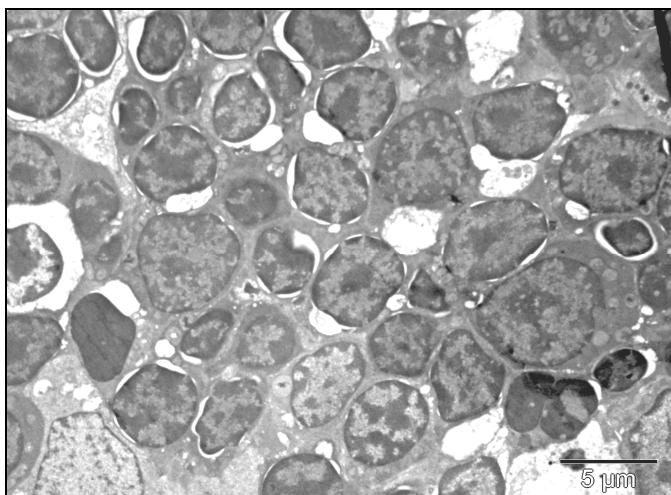
F i g. 4. Migrarea timocitelor evidențiază reticulocitele rarefiate; rarefieri și liză în timocitele aflate în diverse etape ale ciclului celular – lotul TL2.

TIMUSUL ÎN INTOXICAȚIA ACUTĂ CU VENIN DE ALBINE

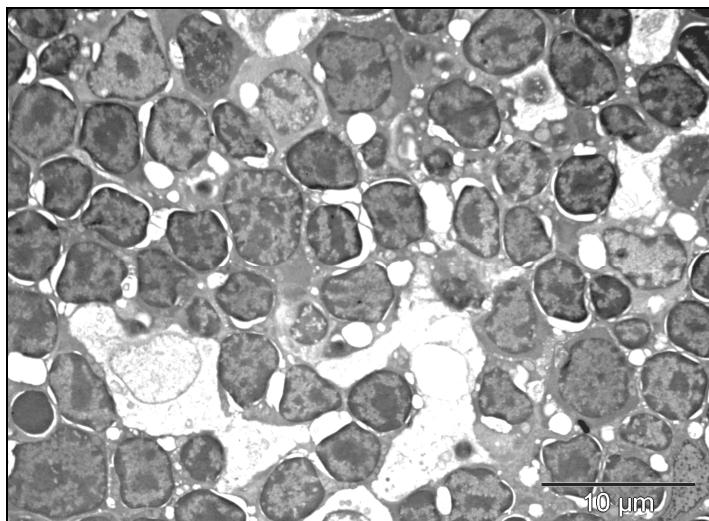
În timocitele ce prezintă un grad de afectare mai redus, totuși mitocondriile au matricea rarefiată (Fig. 4-6). În condițiile prezenței acute a veninului în cantitate mare, apare un număr sporit de timocite surprinse în diferite etape ale proceselor de apoptoză comparativ cu loturile martor (Fig. 4-7). În același timp, numărul mitozelor înregistrate este mai scăzut.



F i g. 5. Densitate celulară redusă; număr mare de timocite apoptotice – lotul TL2.



F i g. 6. Vacuolizări în toate celulele; spațiul perinuclear dilatat; mitocondrii cu matricea rarefiată – Lotul TL2.



F i g. 7. Spațiu perinuclear dilatat; reticulocite electronotransparente – Lotul TL2.

Discuții. Efectele produse asupra timusului în urma administrației veninului de albine în doze mari indică faptul că în această variantă experimentală răspunsul acestui organ este, în principal, unul imunologic. Astfel, timusul răspunde acut la prezența și acțiunea substanțelor străine care atacă structurile altor țesuturi, în primul rând prin punerea masivă a limfocitelor în circulație. Complementar cu acest mecanism, migrarea timocitelor din întreaga structură a timusului este modulată și indirect.

Pe de o parte, Lundin [citat în 29] arată că există o strânsă interrelație între activitatea timusului și cea a adenohipofizei, iar pe de altă parte, s-a demonstrat că eliberarea timocitelor este stimulată și ca urmare a prezenței unei cantități crescute de glucocorticoizi circulańi (răspunsul neuro-endocrin clasic la stres), organele limfopoietice fiind deosebit de sensibile la steroizi [21]. S-a raportat că acțiunea acestor hormoni (de exemplu a cortizonului) poate fi în multe situații răspunzătoare de depopularea și involuția rapidă a timusului [25].

În același context, Oancea și Cojocnea [29] au descris experimente efectuate de mai mulți cercetători care au arătat că extirparea experimentală a suprarenalelor a fost urmată de creșterea în greutate a timusului, cu accentuarea imediată a activității limfocitoproliferative timice. Din contră, administrarea de prednisolon a induș o creștere imediată a numărului de limfocite eliberate de timus, pentru că apoi, în scurt timp, numărul acestora să fie din ce în ce mai scăzut, asociat cu scăderea în greutate a timusului [29]. Această situație este similară cu efectul obținut în experimentul nostru, când, la administrarea acută de venin de albine se înregistrează, la un interval de două ore de la administrare, o scădere a limfocitelor circulante cu 30% față de martor (date preliminare încă nepublicate).

Rezultatele noastre sunt asemănătoare cu cele descrise de Hermenean și colaboratorii [20], care au înregistrat o reducere a populațiilor de timocite din zona corticală în urma administrării pe o perioadă scurtă a unor doze mari dintr-o altă substanță cu efect citotoxic pronunțat (ciclofosfamidă). De altfel, acestea se corelează și cu modificările observate la nivelul splinei după administrarea aceleiași substanțe într-o singură doză terapeutică, când s-a observat o inducție a fenomenelor apoptotice la nivelul limfocitelor [31].

Al doilea efect important al veninului de albine evidențiat prin experimentul nostru este cel timotoxic. Acesta este, de asemenea, foarte pronunțat fiind exercitat direct asupra tuturor categoriilor celulare din structura organului. Veninul a ajuns în cantități mari în proximitatea celulelor ca urmare a dilatărilor de amploare ale capilarelor, proces însotit de diminuarea fluxului sanguin, respectiv de stagnarea săngelui în aceste zone. A urmat rapid traversarea sau, în unele situații, distrugerea pereților capilarelor, ceea ce a permis moleculelor din compoziția veninului de albine să se infiltreze pe calea spațiilor intercelulare spre toate tipurile de celule, până în profunzimea organului, respectiv să-și exerce efectul citolitic la nivelul structurilor timusului. Modificările structurale și ultrastructurale observate de noi au fost determinate, prin acțiunea directă a componentelor veninului, mai ales a melitinei și fosfolipazei A₂ asupra tuturor categoriilor de membrane celulare [4, 6, 18, 19, 22-24, 30, 38]. Vacuolizările citoplasmatiche afectează întreaga populație celulară; observarea lor și la nivelul celulelor endoteliale reticulare sugerează ideea că aceste alterări au contribuit decisiv la scăderea coeziunii timocitelor și au favorizat migrarea lor.

Modificările survenite la nivel ultrastructural sunt deci nespecifice și extinse, de aceeași natură cu cele înregistrate în alte țesuturi investigate de noi în condiții similare [9-11]. Leziunile multiple produse în timus nu sunt însă la fel de ample și generalizate ca cele observate în cazul administrării experimentale a veninului de albine pentru o perioadă mai îndelungată de timp [12].

Rezultatele prezentate de noi sunt în concordanță cu cele raportate de alți cercetători care au abordat studiul efectelor veninului de albine, prin alte metode sau tehnici de investigare [3, 6, 16]. De asemenea, rezultatele noastre experimentale concordă cu cele raportate în cazul unor accidente soldate cu multiple înțepături de albine în urma cărora s-au efectuat variate investigații clinice, inclusiv examinarea unor preparate prin microscopie electronică de transmisie [13, 30].

Concluzii. În urma analizei rezultatelor obținute, putem afirma că dozele foarte mari de venin de albine influențează semnificativ structura și ultrastructura timusului. Pe de o parte, efectele produse la nivelul întregului organism au contribuit la depopularea bruscă a timusului, care încearcă astfel să facă față invaziei acute a moleculelor bioactive străine. Alterările profunde survenite la nivelul diferitelor țesuturi, inclusiv al săngelui sunt deci și ele reflectate, în morfologia și fiziologia timusului, fiind, în ansablu lor responsabile pentru colapsul organismului. Pe de altă parte, componentele veninului au provocat, prin acțiunea lor directă asupra timusului,

lezioni majore la nivelul tuturor categoriilor de celule, dar în primul rând la nivelul limfocitelor T. Ca și în cazul altor țesuturi studiate de noi, consecutive administrației experimentale de venin de albine, efectele veninului la nivelul timusului sunt direct dependente de doză.

B I B L I O G R A F I E

1. Anderson, G., Jenkinson, E. J., Moore, N. C., Owen, J. J. T., *MHC class II positive epithelium and mesenchyme cells are both required for T-cell development in the thymus*, "Nature", **362**, 1993, 70-73.
2. Annila, I. T., Karjalainen, E. S., Annila, P. A., Kuusisto, P. A., *Bee and wasp sting reactions in current beekeepers*, "Ann. Allerg., Asthma, Immunol.", **77**, 1996, 423-427.
3. Azevedo-Marques, M. M., Cupo, P., Hering, S. E., Costa, R., Ferreira, D. B., *Clinical-laboratory and therapeutic aspects of envenomation caused by africanized bees*, "J. Ven. Anim. Tox.", **5** (2), 2001, 154-171.
4. Berg, O. G., Gelb, M. H., Tsai, M. D., Jain, M. K., *Interfacial enzymology: The secreted phospholipase A₂-paradigm*, "Chem. Rev.", **101**, 2001, 2613-2653.
5. Crăciun, C., Miclăuș, V., Crăciun, V., *Studiu ultrastructural al timusului la vârste diferite*, "Rev. Med. Orăd.", **3**, 1996, 95-104.
6. Diaz-Achirica, P., *Permeabilization of the mitochondrial inner membrane by short cecropin-A-melittin hybrid peptides*, "Eur. J. Biochem.", **224**, 1994, 257-263.
7. Eliséiev, V. G., Afanásiev, Y. A., Kotovski, E. F., *Atlas de la estructura microscópica y ultramicroscópica de las células, tejidos y órganos*, Ed. Mir, Moscú, 1987.
8. Florea, A., Crăciun, C., *Reacții ale organismului la veninul de albine și toxicitatea acestuia*, "Stud. Univ. Babeș-Bolyai, Biol.", **48** (2), 2003, 103-114.
9. Florea, A., Crăciun, C., Puică, C., Gherghel, P., Oprea, M. C., *Hepatotoxicitatea veninului de albine. Studiu histologic și ultrastructural*, "Stud. Univ. Babeș-Bolyai, Biol.", **48** (1), 2003, 49-66.
10. Florea, A., Oprea, M. C., Puică, C., Gherghel, P., Crăciun, C., *Efecte structurale și ultrastructurale determinate de veninul de albine în miocard*, "Stud. Univ. Babeș-Bolyai, Biol.", **48** (2), 2003, 89-102.
11. Florea, A., Puică, C., Crăciun, C., Gherghel, P., Oprea, M. C., *Modificări histologice și ultrastructurale induse experimental la nivelul rinichiului prin administrarea veninului de albine în diferite doze*, "An. Soc. Naț. Biol. Cel.", **7** (1), 2002, 327-341.
12. Florea, A., Puică, C., Gherghel, P., Oprea, M. C., Crăciun, C., *Efecte morfofiziologice produse la nivelul timusului prin administrarea subcronică a veninului de albine*, "An. Soc. Naț. Biol. Cel.", **8**, 2003, 169-178.
13. França, F. O. S., Benvenuti, L. A., Fan, H. W., Dos Santos, D. R., Hain, S. H., Picchi-Martins, F. R., Cardoso, J. L. C., Kamiguti, A. S., Theakston, R. D. G., Warrell, D. A., *Severe and fatal mass attacks by "killer" bees (Africanized honey bees – Apis mellifera scutellata) in Brazil: clinicopathological studies with measurement of serum venom concentration*, "Quart. J. Med.", **87**, 1994, 269-282.
14. Fukumoto, T., *Precise analyses of the structure of the thymus for establishing details of the mechanisms underlying thymocyte proliferation and maturation*, "Arch. Histol. Cytol.", **60** (1), 1997, 1-8.

TIMUSUL ÎN INTOXICAȚIA ACUTĂ CU VENIN DE ALBINE

15. Gelderblom, R., *Collecting bee venom – The other way*, "Hivelights", **13** (4), 1997, 12-13.
16. Ghelasvili, D. B., Orlov, B. N., *Proprietățile neurotrope ale veninului de albine și ale fracțiilor sale*, II Int. Simp. Apither., p. 107-109, Ed. Apimondia, București 1976.
17. Graft, D. F., *Maintenance venom immunotherapy*, "Curr. Opin. Allerg. Clin. Immunol.", **2** (4), 2002, 359-362.
18. Habermann, E., *Chimie, pharmacologie et toxicologie du venin*, in Chauvin, R. (Ed.), *Traité de biologie de l'abeille*, p.363-387, Masson, Paris, 1968.
19. Habermann, E., *Bee and wasp venoms. The biochemistry and pharmacology of their peptides and enzymes are reviewed*, "Science", **177**, 1972, 314-322.
20. Hermenean, A., Ardelean, A., Crăciun, C., Borșa, M., Miclăuș, V., *Testarea comparativă a efectului imunomodulator a două extracte de Echinacea în terapia cu ciclofosfamidă la șobolan*, "An. Soc. Naț. Biol. Cel.", **7** (1), 2002, 291-299.
21. Jelea, S., Giurgea, R., *Efectele acetatului de plumb asupra timusului la șobolanii Wistar*, "An. Soc. Naț. Biol. Cel.", **7** (1), 2002, 380-383.
22. Ladokhin, A. S., Selsted, M. E., White, S. H., *Sizing membrane pores in lipid vesicles by leakage of co-encapsulated markers: pore formation by melittin*, "Biophys. J.", **72**, 1997, 1762-1766.
23. Lee, S. Y., Park, H. S., Lee, S. J., Choi, M.-U., *Melittin exerts multiple effects on the release of free fatty acids from L1210 cells: lack of selective activation of phospholipase A₂ by melittin*, "Arch. Biochem. Biophys.", **389** (1), 2001, 57-67.
24. Lin, Y., Nielsen, R., Murray, D., Hubbell, W. L., Mailer, C., Robinson, B. H., Gelb, M. H., *Docking phospholipase A₂ on membranes using electrostatic potential-modulated spin relaxation magnetic resonance*, "Science", **279**, 1998, 1925-1929.
25. Lundin, P. M., Schelin, U., *Ultrastructure of the rat thymus*, "Acta Pathol. Microbiol. Scand.", **65** (3), 1965, 379-394.
26. Miclăuș, V., Crăciun, C., Lisovschi-Cheleșan, C., Rotaru, O., Pușcașiu, M., Crăciun, V., *Presence and evolution of mastocytes in involutional thymus*, in Crăciun, C., Ardelean, A. (Eds.), *Current Problems and Techniques in Cellular and Molecular Biology*, pp.440-443, Mirton, Timișoara, 1996.
27. Mureșan, E., Gaboreanu, M., Bogdan, A. T., Baba, A. I., *Tehnici de histologie normală și patologică*, Ed. Ceres, București, 1974.
28. Nabarra, B., Andrianarison, I., *Ultrastructural studies of thymic reticulum: I. Epithelial component*, "Thymus", **9**, 1987, 95-121.
29. Oancea, T., Cojocea, V., *Timusul. Fiziopatologie, patologie, chirurgie*, pp.14-36, Ed. Militară, București, 1986.
30. Ownby, C. L., Powel, J. R., Jiang, M. S., Fletcher, J. E., *Melittin and phospholipase A₂ from bee (Apis mellifera) venom cause necrosis of murine skeletal muscle in vivo*, "Toxicon", **35**, 1997, 67-80.
31. Pașca, C., Miclăuș, V., Kis, E., Sandu, V.-D., Papuc, I., *Does a single therapeutic dose of cyclophosphamide induce the apoptosis of the lymphocytes?*, "Stud. Univ. Babeș-Bolyai, Biol.", **46** (1), 2001, 119-124.
32. Ploaie, P. G., Petre, Z., *Introducere în microscopia electronică cu aplicații în biologia celulară și moleculară*, Ed. Acad. Rom., București, 1979.
33. Popescu, N., *Unele măsuri necesare pentru menținerea puritatei și calității veninului de albine*, "Apic. România", **1**, 1990, 13-14.

34. Ross, M. H., Romrell, L. J., Kaye, G. I., *Histology – A Text and Atlas*, Williams & Wilkins, Baltimore, 1995.
35. Ushiki, T., Takeda, M., *Three-dimensional ultrastructure of the perivascular space in the rat thymus*, "Arch. Histol. Cytol.", **60** (1), 1997, 89-99.
36. Vicente, A., Varas, A., Sacedón, R., Zapata, A. G., *Histogenesis of the epithelial component of rat thymus: An ultrastructural and immunohistological analysis*, "Anat. Rec.", **244**, 1996, 506-519.
37. von Gaudecker, B., *Functional histology of the human thymus*, "Anat. Embryol.", **183**, 1991, 1-15.
38. Yang, Y., Harroun, T. A., Weiss, T. M., Ding, L., Huang, H. W., *Barrel-stave model or toroidal model? A case study on melittin pores*, "Biophys. J.", **81**, 2001, 1475-1485.

ASPECTE HISTOPATOLOGICE ALE TOXICITĂȚII PANCREATICE
A UNUI AGENT ALCHILANT ANTITUMORAL (CISPLATIN)
LA ȘOBOLANII ALBI WISTAR DE VÂRSTE DIFERITE

CRISTINA PAŞCA* și VICTORIA-DOINA SANDU*

SUMMARY. - Histopathological Aspects of the Toxicity of an Antitumour Alkylating Agent (Cisplatin) on the Pancreas of White Wistar Rats of Different Ages. Cisplatin is an anticancer drug, more exactly an alkylating agent belonging to the family of platinum-containing products, widely used in the chemotherapy of many types of malignant diseases in humans. Unfortunately, according to the previous studies concerning its side effects, this drug has an antitumour activity correlated with a significant toxicity on different vital organs, consisting of the appearance of many and serious structural and functional alterations, which, sometimes, can endanger the life of the patient. Cisplatin has a significant toxicity on the whole digestive system (digestive tube and glands), which, so far, has been just tangentially investigated. Therefore, our researches intended to evaluate the histological and functional alterations induced by a single therapeutic dose of Cisplatin on the pancreas in puberal and mature white Wistar rats.

Our results demonstrated that this antitumour drug, in monochemotherapy, has a moderate toxic effect on the rat pancreas, depending on the age of the rats, the most sensitive being the puberal individuals, in which the destructive processes appeared and got worse more quickly, and affected more seriously the structure of the pancreas. The toxic effect consisted of the appearance of certain histological modifications which affected both the cellular and vascular components of the organ. But, the histopathological alterations, both at the level of the exocrine and endocrine secretory units, seem to be due (induced and aggravated owing) to the grave circulatory disturbances (blood stasis, congestion, oedemas, intravascular coagulation phenomena) and thrombocytopenia, which determined a deficient supply with oxygen and nutritive substances, and an accumulation of some toxic metabolic products.

The exocrine units (or acini) seem to be more sensitive than the islets of Langerhans. They appeared affected by an obvious diffuse oedema correlated with a serious perturbation of the intraacinar and intraductal transit of the pancreatic fluid. Besides, many and abundant lymphocyte and granulocyte infiltrations could be observed, especially perivascularly, around the pancreatic acini.

The endocrine units of the pancreas (islets of Langerhans) were a little more resistant, the toxic action of Cisplatin determining only a transitory and discrete oedema and intravascular coagulation phenomena correlated with a moderate granulo-vacuolar dystrophy, and zonal necrosis processes. In addition, the destructive processes at the level of the islets of Langerhans were not associated with serious perturbation of the secretions of insulin and/or glucagon and, thus, the level of glucose in the blood was not significantly affected.

Fortunately, both in the puberal and mature rats, the histological alterations induced by this dose of Cisplatin, administered in monochemotherapy, although perturbed both the exocrine and endocrine units of the pancreas, had no irreversible character, as at the end of the experimental period, without any protective or regenerative treatment, an obvious natural recovery process could be observed.

* Universitatea "Babeş-Bolyai", Catedra de Biologie Experimentală, 400006 Cluj-Napoca, România.
E-mail: cpasca@biolog.ubbcluj.ro

Extinderea chimioterapiei antineoplazice din ultimii ani, precum și folosirea citostaticelor ca imunosupresoare în pregătirea organismelor pentru realizarea transplantelor de grefe și organe impun cu necesitate studierea și cunoașterea cât mai completă a acțiunii acestor medicamente nu numai asupra structurilor tumorale (cărora le sunt în principal destinate), dar și asupra țesuturilor și organelor sănătoase, la nivelul cărora ele exercită anumite efecte secundare datorită acțiunii selective extrem de reduse pe care o au.

În plus, literatura de specialitate denotă clar faptul că mecanismul de acțiune al citostaticelor în general și, implicit, și al Cisplatinului se bazează pe capacitatea acestor medicamente de a forma legături ireversibile cu ADN-ul și de a interfera cu sinteza ARN-ului și a proteinelor [6, 8, 10]. În consecință, administrarea de citostatice blochează atât diviziunea celulară, cât și sinteza proteică, ambele esențiale pentru conservarea structurii tisulare, pentru buna desfășurare a activității țesuturilor, organelor și, în ultimă instanță, a întregului organism. Datorită acestui mecanism de acțiune al citostaticelor, structurile cel mai drastic afectate în condiții de chimioterapie vor fi acelea, la nivelul cărora indicele mitotic este ridicat și/sau acelea care sunt implicate în desfășurarea unor procese de sinteză proteică mai intense.

Toate aceste considerente justifică pe deplin cercetările noastre, care și-au propus depistarea unor aspecte inedite ale toxicității pancreatică exercitatate la nivel structural, la șobolanii albi Wistar de vârstă diferită, de către Cisplatin (citostatic frecvent utilizat în terapia antineoplazică și într-o oarecare măsură și în transplantul de grefe și organe), cunoscut fiind că acesta este un organ vital, cu o activitate de sinteză proteică constantă și extrem de intensă. În urma parcurgerii unui material bibliografic destul de bogat, am putut constata că aceste aspecte sunt doar tangențial și, implicit, insuficient abordate la ora actuală pe plan mondial.

Material și metodă. Cercetările s-au efectuat pe șobolani albi Wistar, masculi, juvenili (cu o greutate medie de 100 ± 10 g) și adulți (cu o greutate medie de 190 ± 10 grame). Animalele de experiență au fost întreținute în condiții optime de laborator, asigurându-li-se îngrijirea și hrana corespunzătoare și apă *ad libitum*. S-a lucrat pe două serii experimentale (una constituită din juvenili și alta din adulți), fiecare serie incluzând câte 8 loturi a către 8 indivizi fiecare:

-loturile martor M₁-M₄ – constituite din șobolani sănătoși, neutratați cu medicament;

-loturile C₁-C₄ - constituuite din șobolani sănătoși, tratați cu o doză terapeutică unică de 13 mg Cisplatin/kg corp, administrată intravenos în vena codală și sacrificiați la 24 ore, 4, 11 și, respectiv, 18 zile de la tratament în vederea realizării unor dozări biochimice sanguine și a unor investigații structurale.

Sacrificarea animalelor s-a făcut dimineața, la 16 ore de inaniție, după o anestezie profundă cu eter etilic, prin decapitare și exsangvinizare. Imediat după sacrificare s-au recoltat probe de sânge și fragmente de ficat, care au fost imediat prelucrate conform tehniciilor uzuale, astfel încât să fie posibilă dozarea glicemiei și realizarea în condiții optime a investigațiilor de microscopie optică [11, 13].

Rezultate și discuții. Cu toate că în cadrul experimentului nostru am administrat doar o singură doză terapeutică de Cisplatin, în condiții de monochimioterapie, rezultatele investigațiilor structurale și biochimice reflectă cu certitudine faptul că acest citostatic exercită un anumit efect toxic asupra pancreasului de şobolan, atât asupra componentei exocrine, cât și endocrine. Gravitatea, amploarea și extinderea proceselor tisulare distructive par să strâns corelate cu vîrsta organismului supus tratamentului, precum și cu intervalul de timp scurs de la administrarea citostaticului.

Cele mai pregnante și extinse modificări histopatologice au fost înregistrate la *juvenili*, la care efectul toxic se instalează rapid după administrarea medicamentului, se accentuează progresiv și semnificativ în prima jumătate a perioadei experimentale, după care începe să scadă atât ca intensitate, cât și ca extindere. La finele celor 18 zile ale experimentului se ajunge la o refacere structurală avansată a ambelor componente pancreatică, deși, zonal, se mai păstrează încă amprenta acțiunii toxice exercitate de acest xenobiotic.

Astfel, comparativ cu lotul de control, cu o structură histologică normală a pancreasului (Fig.1), am constatat că primele aspecte histopatologice sunt deja decelabile microscopo-optic la doar 24 ore de la tratament și se materializează în prezența unor acini pancreatici tumefiați, cu celule acinoase supraîncărcate cu produși de secreție concentrați apical, care concură la deplasarea nucleului mult înspre polul basal al celulelor. În plus, apare și un edem difuz intraacinar și diminuarea până la dispariție a lumenului adenomerului (Fig. 2).

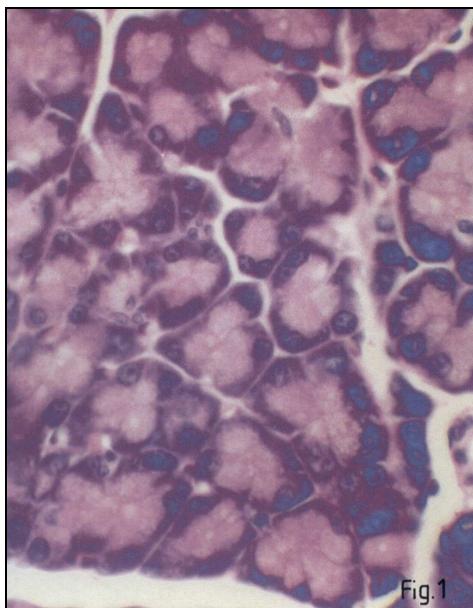


Fig.1

Fig. 1. Aspectul histologic al pancreasului la lotul martor ($\times 1280$).

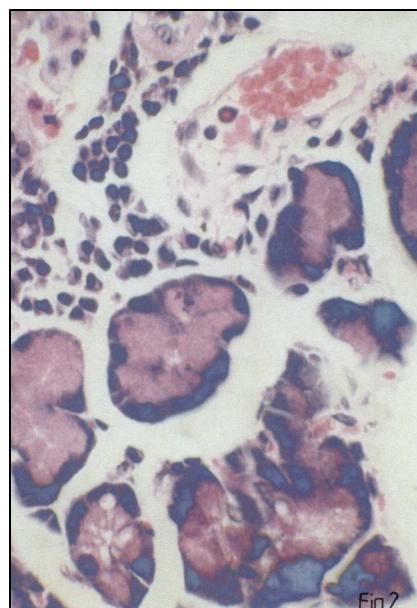


Fig.2

Fig. 2. Acini pancreatici ușor tumefiați, cu celule supraîncărcate cu produși de secreție ($\times 1280$).

Componenta endocrină este moderat afectată, la nivelul ei înregistrându-se o stază ușoară a rețelei vasculare corelată cu procese evidente de coagulare intravasculară diseminată și un edem în insulele Langerhans (Fig. 3).

După 4 zile de la tratament, staza și edemul acinar înregistrează o ușoră intensificare, aspectul tumefiat al acinilor accentuându-se. Totodată, în canalele pancreaticice intra- și interlobulare se constată prezența unei secreții abundente de material oxifil, granular. Deci, caracteristic pentru componenta exocrină a acestui lot este perturbarea activității secretorii acinare, cu stagnarea produsului de secreție atât intracelular, cât și intraductal. În plus, la nivel intersticial se constată prezența unui bogat infiltrat limfogranulocitar (Fig. 4).

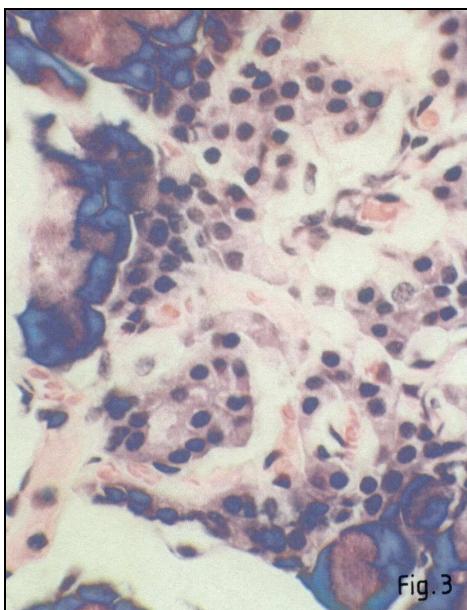


Fig. 3

F i g. 3. Stază sanguină și fenomene discrete de coagulare intravasculară diseminată, edem accentuat în insulele Langerhans (x 1280).

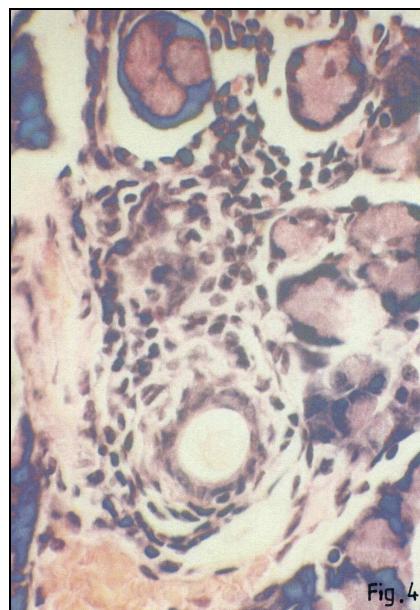


Fig. 4

F i g. 4. Infiltrat limfogranulocitar intersticial abundant (x 1280).

Insulele Langerhans continuă să prezinte tulburări vasculare de tipul perturbărilor de permeabilitate a peretelui vascular, care justifică edemul moderat existent la acest nivel. De asemenea, persistă și procesele de coagulare intravasculară diseminată, ele fiind ceva mai accentuate decât la 24 ore. Nici dimensiunea și nici densitatea celulară a majorității insulelor Langerhans nu par a fi modificate comparativ cu lotul de control (Fig. 5). Însă, unele insule apar mai puternic alterate, în sensul că ele se remarcă printr-o stază sanguină mai pronunțată, prin distrofia granulo-vacuolară și chiar necrozarea unor celule endocrine (Fig. 6).

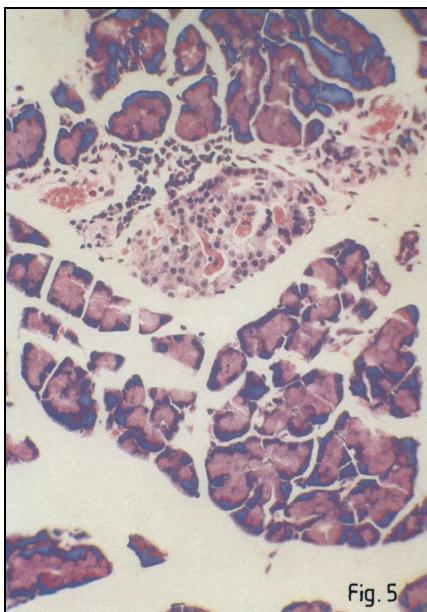


Fig. 5

F i g. 5. Congestie vasculară, stază sanguină și edem moderat în insulele Langerhans (x 512).

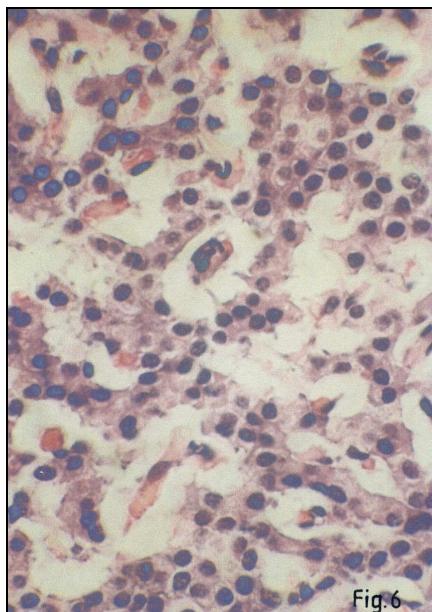


Fig. 6

F i g. 6. Distrofie granulo-vacuolară și chiar necroza unor celule ale insulelor Langerhans (x 1280).

La 11 zile, acinii pancreatici manifestă o tendință de normalizare sub aspectul dimensiunii și structurii lor. Staza sanguină și edemul, deși persistente și destul de extinse, sunt mai reduse, la fel și staza intracelulară și intracanaliculară (Fig. 7). Infiltratul limfogranulocitar intersticial (cu preponderență eozinofilelor) poate fi încă sesizat, ceea ce denotă că citostaticul continuu să exerce un efect iritativ pancreatic semnificativ (Fig. 8).

Componenta endocrină pancreatică apare și ea mai puțin afectată, staza vasculară fiind mai ușoară, iar edemul mai discret, în timp ce fenomenele de coagulare intravasculară diseminată, deși prezente, au doar un caracter zonal (Fig. 9).

La 18 zile de la administrarea citostaticului, aspectele histopatologice înregistrate la 11 zile sunt mai discrete și se mai păstrează doar într-un număr redus de zone pancreatici, în timp ce fenomenele de coagulare intravasculară diseminată nu mai sunt deloc semnalate. Pe ansamblu, atât componenta exocrină, cât și cea endocrină pancreatică au o structură oarecum comparabilă cu cea înregistrată la lotul martor (Fig. 10).

Comparativ cu juvenili, **adulții** se remarcă printr-o mai mare rezistență la atacul pancreaticotoxic al Cisplatinului, în sensul că modificările histopatologice, care sunt similare cu cele înregistrate la juvenili, devin decelabile microscopo-optic abia la 4 zile de la tratament, ating un maxim la 11 zile, după care aproape că dispar către finele perioadei experimentale. Însă, trebuie subliniat că toate perturbările

structurale (staza sanguină, congestia vasculară, coagularea intravasculară diseminată, edemele, staza pancreatică intraacinară și intraductală) au fost mult mai discrete și nu au avut un caracter generalizat ca la juvenili, ci doar unul zonal și moderat. În plus, nu au fost deloc semnalate fenomene necrotice la nivelul insulelor Langerhans, în timp ce distrofia granulo-vacuolară a fost extrem de discretă și de foarte scurtă durată.

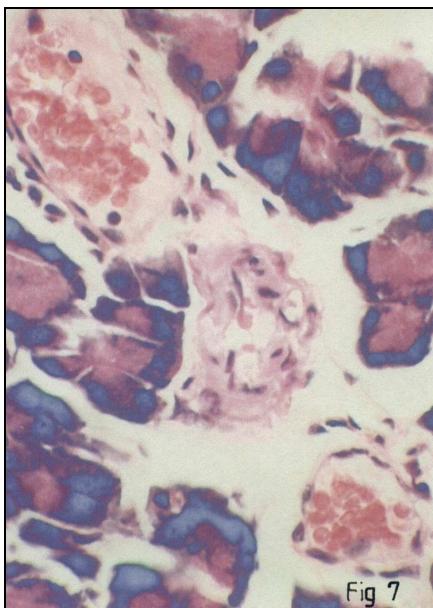


Fig. 7

F i g. 7. Stază sanguină pancreatică generalizată și edeme interstițiale (x 1280).

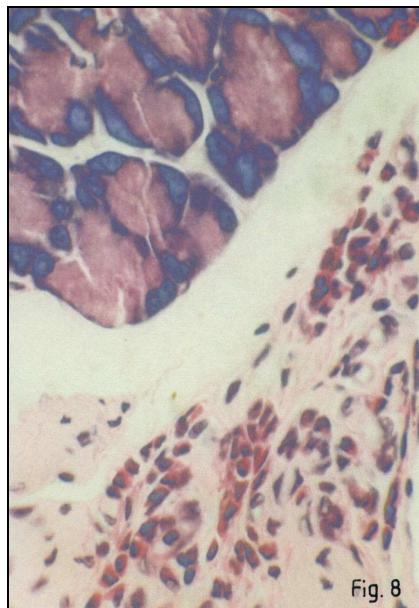
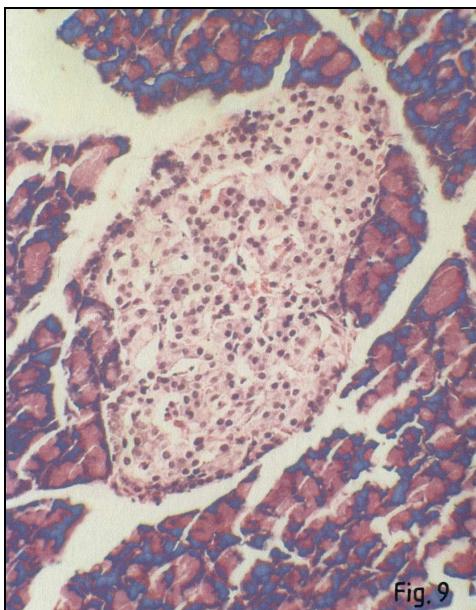


Fig. 8

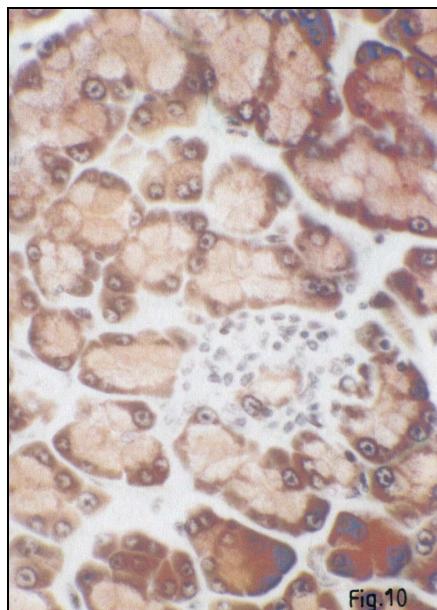
F i g. 8. Infiltrat limfogranulocitar interstițial masiv cu preponderență eozinofilelor (x 1280).

Subliniem că, deși atât la adulți, cât mai ales la juvenili au fost înregistrate modificări structurale la nivelul insulelor Langerhans, ele nu au fost corelate cu perturbări funcționale ale acestora, deci cu alterarea secreției de insulină și glucagon, ceea ce justifică menținerea glicemiei în limite normale pe toată perioada desfășurării experimentului.

Investigațiile histologice și biochimice realizate în cadrul experimentului nostru demonstrează că Cisplatinul administrat chiar și într-o singură doză terapeutică și în monochimioterapie poate conduce la instalarea unor modificări structurale semnificative la nivelul pancreasului şobolanului alb Wistar, care prin gravitatea, ampoloarea și evoluția lor în timp pot altera, într-o oarecare măsură, funcțiile acestui organ vital și, implicit, pot compromite calitatea vieții acestor organisme. Cercetările noastre au demonstrat că acțiunea pancreaticotoxică a acestui citostatic cu platină pare a fi strâns corelată atât cu vîrsta organismului supus tratamentului, cât și cu intervalul de timp scurs din momentul administrării medicamentului.



F i g. 9. Stază vasculară ușoară și edem la nivelul insulelor Langerhans ($x 512$).



F i g. 10. Pancreas cu un aspect histologic foarte apropiat de cel al lotului martor ($x 1280$).

Astfel, cele mai pregnante și extinse modificări histopatologice au fost înregistrate la *juvenili*, la care efectul pancreaticotoxic devine decelabil la doar 24 de ore de la administrarea medicamentului, se accentuează progresiv și semnificativ în prima jumătate a perioadei experimentale, după care începe să scadă atât ca intensitate, cât și ca extindere. La finele celor 18 zile ale experimentului se ajunge la o refacere structurală avansată a ambelor componente pancreaticice, deși, zonal, se mai păstrează încă amprenta acțiunii toxice exercitate de acest xenobiotic.

Comparativ cu juvenilii, *adulții* se remarcă printr-o mai mare rezistență la atacul pancreaticotoxic al Cisplatinului. Așa se explică de ce modificările histopatologice, care sunt similare cu cele înregistrate la juvenili, au fost semnalate microscopoptic mai târziu (abia după 4 zile de la administrare), ating un maxim la 11 zile, după care scad rapid sub raportul amploarei și extinderii lor, pentru că la finele experimentului să dispară aproape complet, structura histologică a pancreasului fiind comparabilă cu cea a lotului de control.

La ambele serii experimentale, dar mai pregnant la juvenili, efectul toxic pancreatic s-a concretizat în apariția unor modificări structurale moderate atât la nivelul componentei vasculare (congestie, stază sanguină, fenomene de coagulare intravasculară diseminată), cât mai ales la nivelul componentei celulare implicate în îndeplinirea funcției exocrine și endocrină a acestei glande.

Mai afectat pare a fi *pancreasul exocrin*, unde s-a înregistrat instalarea stazei sanguine și congestiei vasculare, care concură la apariția unui edem difuz la nivelul

acinilor. În plus, efectul iritativ exercitat de acest medicament justifică prezența a numeroase infiltrate limfohistiocitare și limfogranulocitare interstițiale, cu dispoziție preponderent perivasculară.

Tranzitul secreției pancreaticice este și el perturbat, fapt materializat prin stagnarea produsului secretor atât intracelular cât și intraductal. Așa se justifică hipertrofia acinară generalizată și dilatarea canaliculelor secretorii. Posibil ca această alterare a fluxului de suc pancreatic să fie corelată cu modificările semnificative înregistrate la nivel intestinal, mai exact cu perturbarea sintezei și/sau activității enzimelor marginii în perie implicate în activarea enzimelor sucului pancreatic (potrivit unor aspecte semnalate de noi, dar încă nepublicate). Deci, acest citostatic cu platină pare să afecteze atât procesul secretor, cât și tranzitul produsului de secreție pancreatică.

Componența endocrină (insulele Langerhans) se dovedește a fi mai rezistentă la acțiunea toxică a medicamentului, care, și în acest caz, interesează atât componenta vasculară (congestie, stază sanguină și procese tranzitorii de coagulare intravasculară diseminată), cât și pe cea celulară. Tulburările de hemodinamică de tipul congestiei vasculare și stazei, în condițiile alterărării permeabilității peretelui vascular, par a fi creat condiții propice apariției unui edem difuz insular. Alterarea aportului adecvat de substanțe nutritive și O₂, corelată cu stagnarea locală a produșilor de catabolism cu un anumit potențial toxic, justifică apariția în unele insule a unei distrofii granulo-vacuolare discrete și tranzitorii și a cătorva celule necrozate. Aceste aspecte histopatologice semnalate în cadrul experimentului nostru sunt în concordanță cu datele furnizate de diversi histopatologi, care susțin că diferenți factori toxic, infecțioși și radiațiile pot cauza tulburări circulatorii de intensități diferite, care pot fi generatoare de fenomene distrofice și chiar necrotice insulare, care uneori interesează un număr semnificativ de celule de tip B, ajungându-se la instalarea unui diabet aloxanic [1].

Așa cum reiese din cele prezentate mai sus, fenomenele de coagulare intravasculară diseminată interesează atât vasele sanguine tributare componentei exocrine, cât și endocrine. Conform datelor din literatură, coagularea intravasculară diseminată este un proces patologic destul de complex, care implică participarea sistemului de coagulare sanguin și formarea de trombi în multe vase mici, cu instalarea unor alterări ischemice consecutive în diferite organe. Prezența cordoanelor de fibrină în vasele sanguine mici cauzează deformarea și lezarea hematiilor la trecerea lor prin rețea de fibrină intravasculară și, astfel, la organismele cu fenomene de coagulare intravasculară diseminată, apare, implicit, și o anemie hemolitică corelată cu prezența unor eritrocite anormale sub aspect morfologic – mai exact apare ceea ce se numește anemie hemolitică micro-angiopatică [14].

De altfel, în prezent este cunoscut că Cisplatinul provoacă destul de frecvent o anemie hemolitică cu caracter progresiv, care deseori impune realizarea unor transfuzii cu masă eritrocitară [9]. Una dintre cauzele majore ale coagulării intravasculare disseminate o reprezintă perturbările hematologice și, în special, trombocitopenia [4, 14]. Aceasta explică amploarea deosebită a coagulării intravasculare disseminate înregistrată

la şobolanii trataţi cu Cisplatin în cadrul investigaţiilor noastre, acest citostatic fiind cunoscut ca un mielodepresiv foarte puternic, factorul limitant al dozei fiind trombocitopenia, maximă la 3 săptămâni de la tratament, revenirea la normal având loc abia după 4 săptămâni [2, 3, 5, 7, 12, 15]. Putem concluziona că este posibil ca trombocitopenia indusă de acest agent alchilan cu platină să declanșeze fenomene de coagulare intravasculară diseminată, care, grație cheagurilor de fibrină constituite, creează condiții propice instalării unor tulburări ischemice în diferite organe și a unei anemii hemolitice.

Subliniem că, pe baza literaturii de specialitate destul de bogată consultată de către noi, instalarea unor astfel de fenomene de coagulare intravasculară diseminată la nivel pancreatic, după chimioterapie cu Cisplatin, este un aspect histopatologic nou, care nu a mai fost semnalat.

Referitor la intensitatea și dinamica modificărilor histopatologice induse de Cisplatin la cele două serii experimentale trebuie subliniat că la adulți staza sanguină, congestia vasculară, coagularea intravasculară diseminată, edemele, staza pancreatică intraacinară și intraductală au fost mult mai discrete și nu au avut un caracter generalizat ca la juvenili, ci doar unul zonal și moderat. În plus, nu au fost deloc semnalate fenomene necrotice la nivelul insulelor Langerhans, în timp ce distrofia granulo-vacuolară a fost discretă și de foarte scurtă durată.

Subliniem și faptul că, la nici una dintre cele două serii experimentale, alterările structurale înregistrate la nivelul insulelor Langerhans nu au condus la instalarea unui diabet aloxanic (indus prin distrofii și necroze ale celulelor B pancreatică), ceea ce explică menținerea glicemiei în limite normale pe parcursul întregii perioade experimentale.

Cu toate că unele aspecte histopatologice pancreaticice înregistrate atât la nivelul componentei exocrine, cât și endocrine au fost destul de puternice, extinse și de durată (în principal la juvenili), ele au avut un caracter reversibil, tranzitoriu, făcând posibilă refacerea naturală în timp a tuturor structurilor afectate.

Concluzii. 1. Cisplatinul, administrat într-o doză terapeutică unică de 13 mg/kg corp, în condiții de monochimioterapie, conduce la instalarea unor modificări structurale semnificative atât la nivelul pancreasului, cât mai ales la nivelul ficatului şobolanului alb Wistar, care prin gravitatea, amploarea și evoluția lor în timp pot altera într-o oarecare măsură funcțiile acestor organe vitale și, implicit, pot compromite calitatea vieții acestor organisme.

2. Cercetările noastre au demonstrat că acțiunea pancreaticotoxică a Cisplatinului pare a fi strâns corelată atât cu vîrsta organismului supus tratamentului, cât și cu intervalul de timp scurs din momentul administrării medicamentului.

3. Cele mai pregnante și extinse modificări histopatologice au fost înregistrate la *juvenili*, la care efectul pancreaticotoxic devine decelabil la doar 24 de ore de la administrarea medicamentului, se accentuează progresiv și semnificativ în prima jumătate a perioadei experimentale, după care începe să scadă atât ca intensitate,

cât și ca extindere. La finele celor 18 zile ale experimentului se ajunge la o refacere structurală avansată a ambelor componente pancreaticice, deși, zonal, se mai păstrează încă amprenta acțiunii toxice exercitate de acest xenobiotic.

4. Comparativ cu juvenili, *adulții* se remarcă printr-o mai mare rezistență la atacul pancreaticotoxic al Cisplatinului. Așa se explică de ce modificările histopatologice, care sunt similare cu cele înregistrate la juvenili, au fost semnalate microscopo-optic mai târziu (abia după 4 zile de la administrare), ating un maxim la 11 zile, după care scad rapid sub raportul ampoarei și extinderii lor, pentru ca la finele experimentului să dispară aproape complet, structura histologică a pancreasului fiind comparabilă cu cea a lotului de control.

5. La ambele serii experimentale, dar mai pregnant la juvenili, efectul toxic pancreatic s-a concretizat în apariția unor modificări structurale moderate atât la nivelul componentei vasculare (congestie, stază sanguină, fenomene de coagulare intravasculară diseminată), cât mai ales la nivelul componentei celulare implicate în îndeplinirea funcției exocrine și endocrine a acestei glande.

6. Tulburările hemodinamice pancreaticice induse de Cisplatin, de tipul stazei, congestiei și edemului, prin inducerea unui deficit de substanțe nutritive și O₂ și prin blocarea îndepărțării produșilor de catabolism cu potențial toxic, este posibil să se afle la baza fenomenelor distructive ce interesează celulele acinilor și insulelor Langerhans.

7. Mai sensibil la acțiunea Cisplatinului pare a fi *pancreasul exocrin* unde s-a înregistrat instalarea stazei sanguine și congestiei vasculare, care concură la apariția unui edem difuz la nivelul acinilor și a numeroase infiltrate limfohistiocitare și limfogranulocitare interstitionale, cu dispoziție preponderent perivasculară.

8. Tranzitul secreției pancreaticice este și el perturbat (tranzitoriu), fapt materializat prin instalarea stazei produsului secretor atât intracelular, cât și intraductal.

9. Componenta endocrină (insulele Langerhans) pare a fi mai rezistență la acțiunea toxică a medicamentului, constatăndu-se doar prezența unui edem difuz și a unor procese tranzitorii de coagulare intravasculară diseminată pe fondul unei distrofii granulo-vacuolare discrete și tranzitorii.

10. La nici una dintre cele două serii experimentale, alterările structurale înregistrate la nivelul insulelor Langerhans nu au fost corelate cu modificări semnificative ale glicemiei pe parcursul perioadei experimentale.

11. La adulți, stază sanguină, congestie vasculară, coagularea intravasculară diseminată, edemele, stază pancreatică intraacinară și intraductală au fost mult mai discrete și nu au avut un caracter generalizat ca la juvenili, ci doar unul zonal și moderat. În plus, nu au fost deloc semnalate fenomene necrotice la nivelul insulelor Langerhans, în timp ce distrofia granulo-vacuolară a fost discretă și de foarte scurtă durată.

12. Cu toate că unele modificări histopatologice pancreaticice înregistrate la nivelul componentei vasculare și celulare (exocrine și endocrine) au fost destul de puternice, extinse și de durată (în principal la juvenili), toate au avut un caracter tranzitoriu, fiind posibilă refacerea naturală în timp a tuturor structurilor afectate.

B I B L I O G R A F I E

1. Avram, N., Begnescu, R., Bucur, E., Ciocnitu, V., Lungeanu, A., Manolescu, N., Păltineanu, D., Păunescu, G., Șirbu, C., Voiculescu, I., *Citologie normală și patologică*, Ed. Ceres, București, 1980.
2. Bloommaert, F. A., Dijk-Knijnenburg, H. C. M., Dijt, F. J., Engelse, L., Baan, R. A., Berends, F., Fischtinger-Schepman, A. M. J., *Formation of DNA adducts by the anticancer drug carboplatin: different nucleotide sequence preferences in vitro and in cell*, "Biochemistry", **34**, 1995, 8474-8480.
3. Calvert, A. H., *Dose optimisation of alkylant agents in adults*, "Anticancer Res.", **14**, 1994, 2273-2278.
4. Chanarin, I., *The Megaloblastic Anaemias*, p. 880-884, Blackwell Sci., Oxford, 1979.
5. Crown, J., Hakes, T., Reichman, B., Lebwohl, D., Glewski, T., Surbone, A., Currie, V., Yao, T. J., Huddis, C., Seidman, A., *Phase II trial of cisplatin and etoposide in metastatic breast cancer*, "Cancer", **71**, 1993, 1254-1257.
6. Denham, J. W., Abbott, R. L., *The effect of lipid peroxidation on the activity of various membrane-bound ATPases in rat kidney*, "J. Clin. Oncol.", **9**, 1991, 458-464.
7. Hruban, R. H., Sternberg, S. S., Meyers, P., Fleisher, M., Menendez-Botet, C., Boitnott, J. K., *Fatal thrombocytopenia and liver failure associated with cisplatin therapy*, "Cancer Inv.", **9**, 1991, 263-268.
8. Lindauer, E., Holler, E., *Cellular distribution and cellular reactivity of platinum (II) complexes*, "Biochem. Pharmacol.", **52**, 1996, 7-14.
9. Marani, T. M., Trich, M. B., Armstrong, K. S., Ness, P. M., Smith, J., Minniti, C., Sandier, S. G., *Cisplatin-induced immune hemolytic anemia*, "Transfusion", **36**, 1996, 1016-1018.
10. Milicevic, Z., Slepcevic, V., Nicolic, D., *Role of DNA repair in the mechanisms of cell resistance to alkylating agents and cisplatin*, "Cancer Chemother. Pharmacol.", **32**, 1994, 85-89.
11. Mureșan, E., Gaboreanu, M., Bogdan, A. T., Babai, A. I., *Tehnici de histologie normală și patologică*, Ed. Ceres, București, 1974.
12. Slavutsky, I., Campossi, E., Gonzalez-Cid, M., Larripa, I., *Micronuclei induction by carboplatin in human lymphocyte subpopulation*, "Anticancer Drugs", **6**, 1995, 758-762.
13. Teuscher, A., Richterich, P., *Biochemische Technik, „Schweiz. Med. Wochenschr.“*, **101**, 1971, 345-390.
14. Woodiff, H. J., Herrmann, R. P., *Concise Haematology*, p. 208-212, Arnold, London, 1979.
15. Zoubek, A., Kajtar, P., Flucher-Wolfram, B., Holzinger, B., Mostbeck, G., Thun-Hohenstein, L., Fink, F. M., Urban, C., Mutz, I., Schuler, D., *Response of untreated stage IV Wilms' tumour to single dose cisplatin assessed by "up front" window therapy*, "Med. Pediatric Oncol.", **25**, 1995, 8-11.

FUNCTIONAL CORRELATIONS BETWEEN THE PERMEABILITY TRANSITION, MEMBRANE POTENTIAL COLLAPSE AND CALCIUM RELEASE BY LIVER MITOCHONDRIA OF ETHANOL-FED RATS

CORNELIU TARBA* and FELICIA SUĂRAŞAN*

SUMMARY. – Male white Wistar rats were maintained in our animal facility for 14-18 weeks and the evolution of their body weight was assessed periodically. One group of the rats, kept on a normal diet, served as control (C), while in another group (A) each rat was supplemented individually with 1.5 ml of 48% ethanol/100 g body weight daily. At the end of the period, the rats were sacrificed and their liver used for the preparation of mitochondria. For membrane potential ($\Delta\Psi$), calcium (Ca^{2+}) and swelling measurements, 1 mg of mitochondrial protein/ml was usually incubated directly in the spectrophotometer cuvettes, at room temperature (around 24°C), in different media, usually containing 5 mM Hepes buffer (pH 7.37), 1 mM KP_i and various concentrations of mannitol, sucrose, KCl and MgCl₂. 8 µM rotenone was always added to the mitochondrial suspension, as well as the appropriate probe for either $\Delta\Psi$ (2.5 µM diS-C₂-5(5)) or Ca^{2+} (30 µM arsenazo III). The respiration and all the associated phenomena monitored by us ($\Delta\Psi$, Ca^{2+} fluxes and swelling) were usually triggered by the addition of succinate (2.5 mM). Different amounts of CaCl₂ or other modulating factors were either added gradually or in one pulse, up to the desired final concentration. Recordings of either $\Delta\Psi$ or calcium fluxes in parallel with the matrix swelling were performed by a diode-array spectrophotometer. In addition, for quantitative measurements of the extension of swelling, special runs were also performed in which mitochondrial protein was used as the triggering factor. From the comparison of the spectrophotometric recordings, one can observe a strong correlation between the concentration of calcium added to the mitochondrial suspension, the moment of the $\Delta\Psi$ collapse, of the swelling and of the calcium release (massive calcium efflux). As expected, these phenomena occur faster (at shorter times and/or lower calcium concentrations) in ionic media, especially in those that lack magnesium. However, important differences were also noticed between the mitochondria of the control (C) rats and those of the animals treated with ethyl alcohol (A) regarding the degree (amplitude) and the kinetics of the absorbance changes associated with the generation/collapse of $\Delta\Psi$, Ca^{2+} movements and swelling. In general, mitochondria from ethanol-treated rats are more sensitive to calcium. Even though the amplitude of membrane potential does not seem to be affected, it collapses at much lower concentrations of added calcium. Even though, in general, the extension of swelling does not differ statistically significant, this sensitivity is also valid for the matrix swelling and the release of calcium. We have quantified this relationship by considering the number of calcium pulses (of 12.5 µM each) added to the cuvette before massive calcium efflux (release) and $\Delta\Psi$ collapse occur. As expected, this number is lower in ionic media (with the lowest in the swelling medium), but for the same medium it is also lower for the ethanol-fed rats and the differences are in general statistically very significant ($p<0.01$) in all media tested.

*Babeş-Bolyai University, Department of Experimental Biology, 400006 Cluj-Napoca, Romania.
E-mail: cntarba@biochem.ubbcluj.ro

Chronic ethanol consumption (in humans) or administration (to laboratory mammals) has been generally associated with specific hepatic structural and functional alterations known under the name of alcoholic liver disease (ALD). Despite many studies, the exact mechanism by which ethanol induces the disease is not entirely known. Whether ALD is determined by the nutritional defects induced by alcohol or by its direct hepatotoxic effect is still a matter of debate [17, 25, 40, 41]. For decades, alcoholic liver disease has been attributed to necrotic events associated, among others, with the production of proinflammatory cytokines, such as interleukin-1 (IL-1) and tumour necrosis factor- α (TNF- α) [11, 44, 68]. However, relatively recent data indicate that under certain circumstances both the acute and chronic ethanol administration is associated with a process of liver apoptosis [1, 12, 26, 36, 46], which, paradoxically, can facilitate latter necrotic events leading to ALD [1, 12, 13, 29, 45, 68]. On the other hand, progresses made during the last decade in the study of apoptosis (physiologic or programmed cell death) have placed mitochondria in a central position in controlling this very complicated process [3, 5, 27, 32-34, 42, 49, 52-54, 56, 57].

Excessive alcohol consumption has been known to be associated with a series of intracellular stresses, particularly detrimental to mitochondria [7, 9, 15, 35, 62, 68]. One of the possible ways of alcohol attack may be through distorting the oxidant/antioxidant balance of the cell. Oxidative stress, associated with chronic ethanol consumption, leads to the depletion of mitochondrial glutathione, decreased synthesis of the respiratory chain components and acetaldehyde adduct formation [2, 3, 11, 18, 25, 45, 71]. The result is an increase in the concentration of reactive oxygen species (ROS), such as superoxide anion, hydroxyl and peroxy radicals [15, 19, 20, 29, 35, 36, 39], which are considered to be among the main causes of triggering a drastic change in mitochondrial membrane permeability, known as the mitochondrial permeability transition (MPT), associated with the opening of a permeability transition pore (PTP). This permeability change seems to be a central event in different types of cell death, either apoptosis or necrosis [9, 24, 27, 32, 38, 42, 45, 50, 62, 68]. What exactly determines the type of cell death (apoptosis or necrosis) is not perfectly understood, although it seems to be related finally to the amount of ATP produced by mitochondria, as described elsewhere [37, 47, 62].

Without getting into the details of the structure and function of the pore, we mention that PTP is probably formed at the so-called membrane contact sites by the association of several proteins belonging to the two mitochondrial membranes and the intermembrane space, thus penetrating the whole insulating system of the mitochondrial matrix. The opening of this pore is accompanied by membrane potential ($\Delta\Psi$) collapse, uptake of electrolytes and water, matrix swelling and ruptures of the mitochondrial outer membrane. As a consequence, several factors present in the intermembrane space, among which cytochrome c (cyt c), are liberated into the cytosol. There is, however, an alternative point of view, which holds that cyt c can be released without a permeability transition and outer membrane breaking, due to certain pores created into the outer membrane by several proapoptotic factors (proteins), such as Bax and Bid [14, 31, 43, 69]. Indeed, the implication of Bax and

Bid in apoptosis can not be negated, but many investigators consider that these proteins only interact with the PTP components, activating or regulating the opening of the PT pores, a fact which could explain certain contradictory observations regarding the dependence/independence of cyt c release on $\Delta\Psi$ collapse [22, 70]. The groups of Martinou [14] and Orrenius [21] demonstrated in fact that there are two distinct mechanisms by which cyt c can be liberated: a Ca^{2+} -dependent one, which takes place apparently by matrix swelling and outer membrane rupture, and a Ca^{2+} -independent mechanism, in which the oligomeric form of Bax mediates or regulates the release of cyt c, without permeability transition. On the other hand, the implication of calcium itself in the cell death, in general, is also well documented [8, 21, 23, 24, 48, 51, 55, 58, 67].

In trying to clarify some aspects of such controversial problems of mitochondrial involvement in apoptosis (specifically in liver apoptosis), especially the relationship between cyt c release and membrane permeability, we selected chronic ethanol feeding as a natural model of producing hepatic apoptosis in combination with the use of suspending media of different ionic content and addition of different concentrations of calcium, while monitoring spectrophotometrically, in parallel, membrane potential, calcium release and matrix swelling. Liver slices and aliquots of both supernantant and mitochondrial sediment, taken at specific moments of incubation, were prepared or saved for later analysis, including electron microscopy. The present article describes and discusses only the spectrophotometric results of our study. A preliminary report has already been presented [65].

Material and methods. *Animals, treatment protocols and preparation of mitochondria.* Male white Wistar rats were kept under normal conditions in our animal facility for 14-18 weeks, starting from an average weight of 120 g/individual, while the evolution of their weight was assessed periodically. The rats were kept on a normal diet (a premix containing all the ingredients of the Larsen diet), with free access to water. One group served as control (C), while in another group (A) each rat was supplemented daily with 1.5 ml of 48% ethanol/100 g body weight, administered in the morning, on a little piece of bread, before getting access to the food. At the end of the period, the rats were fasted for 24 hrs and sacrificed by decapitation after a slight anaesthesia. Small pieces of liver were taken in some cases and prepared for electron microscopy, while the rest of the liver was used for preparation of mitochondria, essentially according to Johnson and Lardy [30], in a medium containing 200 mM mannitol, 70 mM sucrose, 5 mM Hepes-KOH (pH 7.37) and 0.5 mM Na-EDTA. The washing and preserving medium lacked the chelating agent (EDTA).

Spectrophotometric measurements. A diode-array spectrophotometer (Specord S 100B, Analytik Jena, Germany), which permits concomitant measurements at different wavelengths, was used for measurements of membrane potential and/or calcium in parallel with the matrix swelling, while the extension of swelling was quantitatively measured (at 540 nm) on separate recordings initiated by the addition of mitochondria to a medium containing succinate as a respiratory substrate. In parallel

recordings, membrane potential was estimated as a difference between 670 and 700 nm, by the use of 2.5 μM diS-C₂-(5) as a potential sensitive probe [59, 60], while calcium was measured by the use of 30 μM asenazo-III (650–700 nm) as a calcium specific dye. The change in absorbance at 700 nm (which represents a quasi-isobestic point for both dyes) was used for assessing the extent (amplitude) of swelling. For measurements, 1 mg of mitochondrial protein/ml was usually incubated directly in the spectrophotometer cuvettes, at room temperature (around 24°C), in different media, all containing 5 mM Hepes buffer (pH 7.37), 1 mM KP_i and various concentrations of mannitol, sucrose, KCl and MgCl₂, as specified below. The respiration and the associated phenomena monitored by us ($\Delta\Psi$, Ca²⁺ fluxes and swelling) were triggered by the addition of succinate (2.5 mM) in the presence of 8 μM rotenone. Different amounts of CaCl₂ were either added gradually, in several pulses, or in one pulse, up to the desired final concentration. The suspending media used are designated as follows: MS (210 mM mannitol, 70 mM sucrose, 1.5 mM MgCl₂); MSK (110 mM mannitol, 40 mM sucrose, 65 mM KCl, 1.5 mM MgCl₂); MSK-Mg (the same as MSK, but without Mg). In addition, a special swelling medium (KSW) was also used. It contained 100 mM KCl, 50 mM sucrose, 10 mM Hepes and 5 mM KP_i. Although we made some attempts at obtaining precise values of membrane potential amplitude and calcium concentrations (as described elsewhere [65]), the most important quantitative result was obtained by counting the number of calcium pulses (of 12.5 μM each) necessary to be added to the mitochondrial suspension for producing the calcium release, which is always more-or-less associated with the permeability transition (as monitored by matrix swelling) and $\Delta\Psi$ collapse. The statistical significance of the differences between different groups or subgroups was computed by the Student *t* test.

Results and discussion. *Mitochondria of the control rats.* The behaviour of the control mitochondria (C) in the 4 media used for suspension (incubation), regarding the kinetics of calcium fluxes, membrane potential and swelling, under stress conditions induced by calcium, are presented in Figs. 1–8. One of the curves in these figures (usually, the upper trace or the one noted A), represents the variation of absorbance difference between 650 nm (in the case of calcium fluxes) or 670 nm (in the case of $\Delta\Psi$) and 700 nm, whereas the second trace represents the absorbance change at 700 nm (the quasi-isobestic point of the two dyes), associated with the matrix swelling. Whereas the first is a continuous curve, lacking any electronic accidents (due to the differential way in which it was recorded), the “spikes” on the second trace reflect the electronic noise associated with the additions made into cuvettes. Although this noise is in general very visible (hence, the moments of addition are well marked), it must be mentioned that in some of the additions (see, for example, Figs. 1 and 6) the noise is barely perceivable. The 8 figures in this group should be considered in pairs (1–2, 3–4, 5–6 and 7–8), each pair reflecting the behaviour of mitochondria in one of the 4 media: MS, MKS, MKS-Mg and KSW, respectively.

From the analysis of the illustrations presented in these figures, one can notice the "saw tooth" behaviour in the upper traces of the odd number figures (1, 3, 5, 7), recorded by sequential additions of 12.5- μ M calcium pulses. The number of teeth varies from 7-9 in the MS medium to 2-3 in the KSW medium, after which the absorbance increases gradually in a hyperbolic fashion, indicating the massive calcium efflux (release) associated with the permeability transition, as indicated by the strong absorbance decrease in the lower curve (or the one noted B) of the same figure. The behaviour of membrane potential is illustrated in the upper traces of the even number figures (2, 4, 6, 8), also in parallel with the absorbance decrease (lower trace) associated with the extensive swelling due to the permeability transition. As expected, the permeability transition, $\Delta\Psi$ collapse and calcium release occur at a lower and lower calcium loads as we pass from a medium almost devoid of ions (MS) to an almost completely ionic medium (KSW), which demonstrates the involvement in these events of ion transport phenomena at the level of the inner mitochondrial membrane. In addition, the lack of Mg enhances even more this behaviour, a fact which is in accord with what is known from the literature with regard to the regulating role of Mg^{2+} in the calcium fluxes [4, 64, 72].

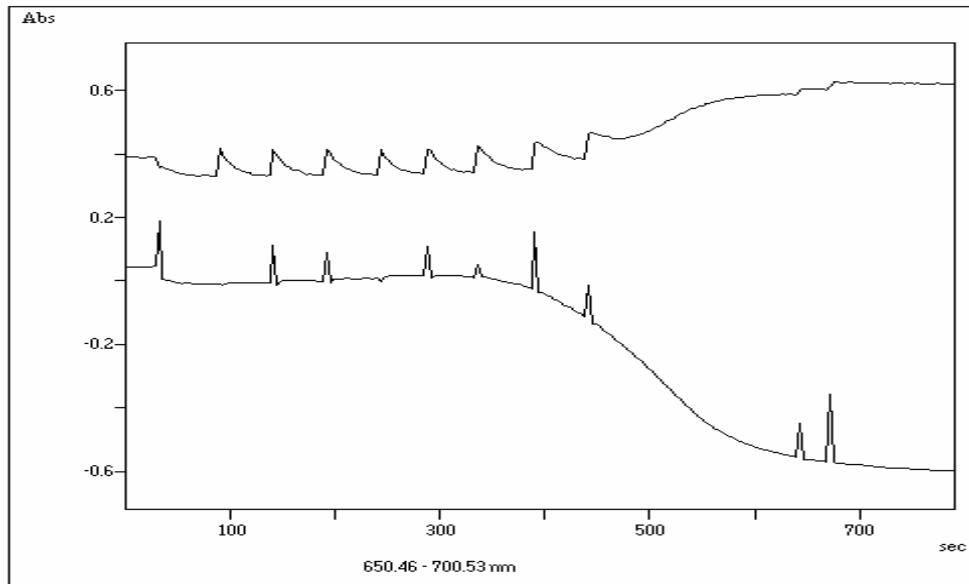


Fig. 1. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MS medium.

In all figures, the first spike is associated with the injection of succinate (2.5 mM), the rest being associated with $CaCl_2$ pulses of 12.5 μ M each, except those added after Ca^{2+} release (or $\Delta\Psi$ collapse).

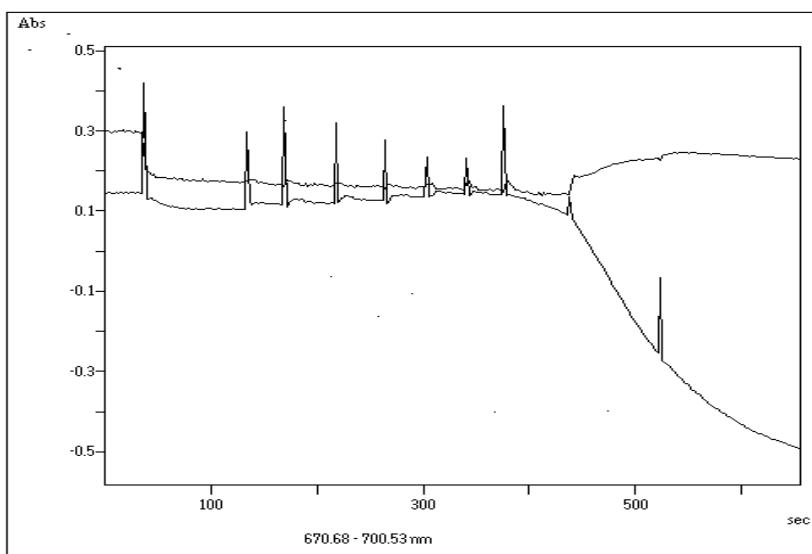


Fig. 2. Membrane potential (upper curve) and mitochondrial swelling (lower curve) in the MS medium. Note the relatively close association between the initiation of $\Delta\Psi$ collapse (absorbance increase on the upper curve) and of mitochondrial swelling (absorbance decrease on the lower curve).

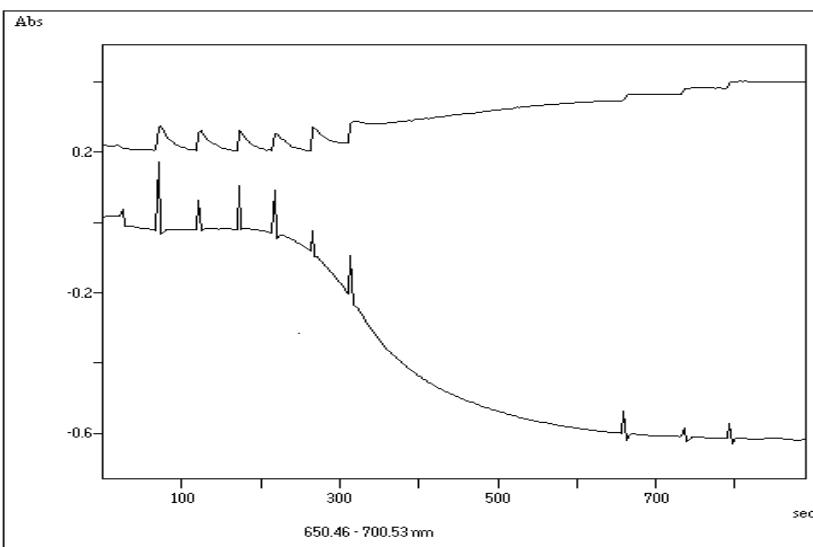


Fig. 3. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MSK medium. Note the fewer pulses of Ca^{2+} (than in the MS medium) to which the mitochondria resist.

FUNCTIONAL CORRELATIONS IN LIVER MITOCHONDRIA OF ETHANOL-FED RATS

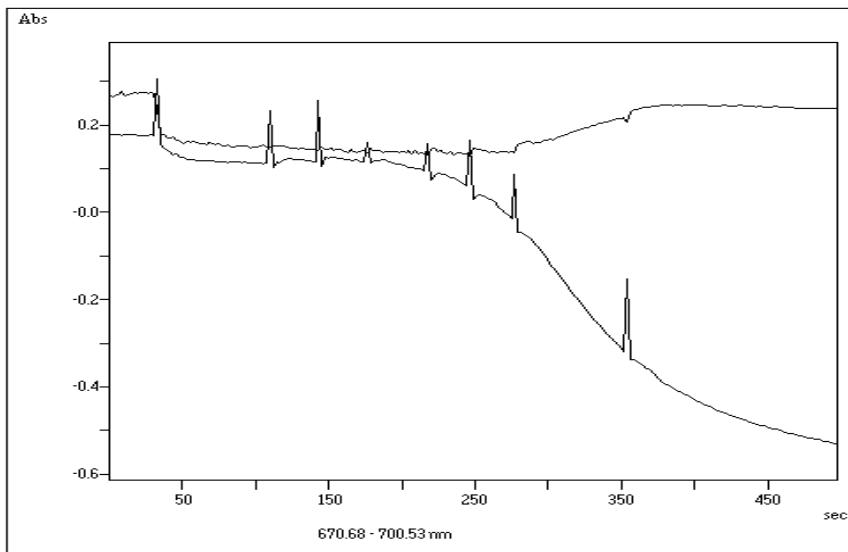


Fig. 4. Membrane potential (upper curve) and mitochondrial swelling (lower curve) in the MSK medium. Note that the swelling begins sooner than the initiation of $\Delta\Psi$ collapse.

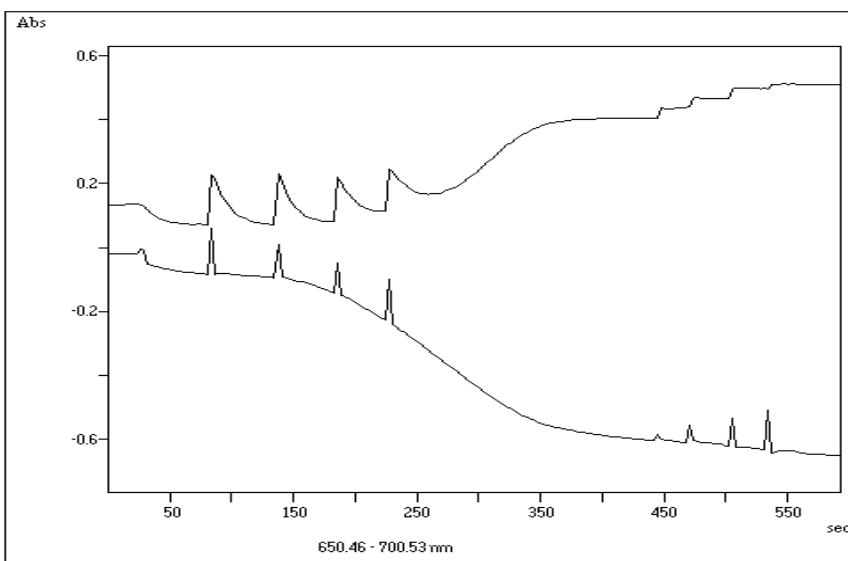


Fig. 5. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MSK-Mg medium. The massive calcium efflux (release) occurs easier than in the presence of Mg^{2+} .

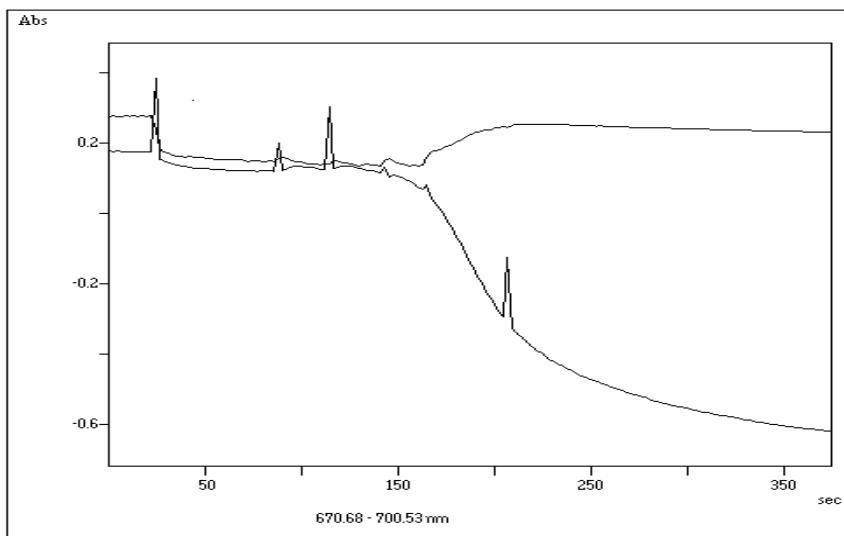


Fig. 6. Membrane potential (upper curve) and mitochondrial swelling (lower curve) in the MSK-Mg medium. Calcium pulses 3 (140 s) and 4 (160 s) are accompanied by very small electronic noises.

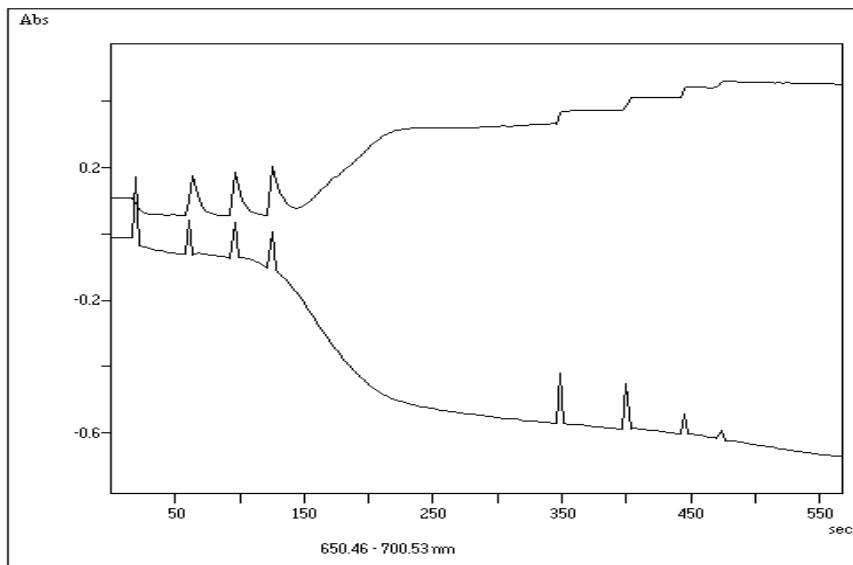


Fig. 7. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the KSW medium. Note that the swelling begins sooner than the release of Ca^{2+} .

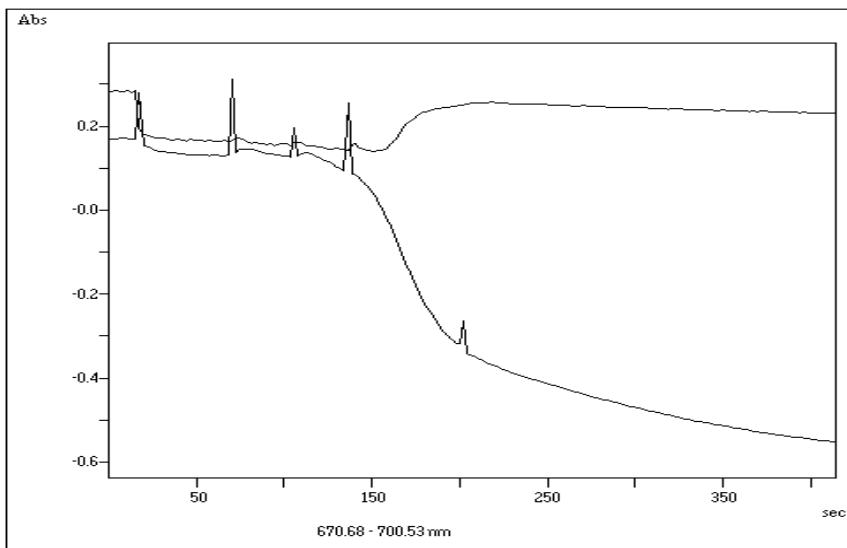


Fig. 8. Membrane potential (upper curve) and mitochondrial swelling (lower curve) in the KSW medium. Note the difference between the beginning of swelling and of the $\Delta\Psi$ collapse.

Despite a rather evident parallelism, at a closer analysis, one can observe that the swelling slightly precedes the other two phenomena. Thus, in Fig. 1, one can see that the swelling begins after the 6th calcium pulse while calcium release occurs after the 8th pulse. In Fig. 2, the swelling occurs after the 7th calcium pulse whereas $\Delta\Psi$ collapse begins only after the 8th pulse. From Figs. 5-6 one can observe that the swelling begins after the second pulse while $\Delta\Psi$ collapse and calcium release only after the 4th pulse. There are many discussions in the literature with regard to the succession of these events. Some researchers claim that the permeability transition (PT) may occur after $\Delta\Psi$ collapse while others that $\Delta\Psi$ can be preserved or restored after PT and cyt c release, leaving space for different scenarios regarding the exact relationship (conditioning) among these events (see [6, 14, 21, 31, 69, 70]). Our work proves that under relatively physiological conditions the first phenomenon to be triggered is the swelling, followed shortly by $\Delta\Psi$ collapse and calcium release, which proceed apparently in parallel with the moment of maximum rate of swelling (hence, PT), and there is no recovery. One should not forget, however, that the phenomena recorded by us represent a statistical result of a relatively variable behaviour of a mitochondrial population *in vitro*. Therefore, we can not exclude that under *in situ* physiological conditions some of the mitochondria, that have suffered less, may recover. This possibility probably depends also on the intensity and duration of the stress factor. As regards the relationship between $\Delta\Psi$ and calcium fluxes, this is more complicated and the relatively slow response of our instrument does not allow a clear-cut conclusion. In fact, we think that the problem has

already been solved more than two decades ago by Bernardi and Azzone [2], which showed that the massive calcium efflux begins when $\Delta\Psi$ decreases below 130 mV.

Mitochondria of ethanol-fed rats. Before presenting the spectrophotometric recordings, we have to make the observation that of the 8 treated animals only 6 behaved in a consistent manner, characteristic for alcohol-fed rats, while two of them had a totally different behaviour. The quantitative measurements that we are going to present later in this article will prove that the two (sub)groups differ significantly from a statistic point of view. While the group of 6 (designated A₁₋₆) showed a higher sensitivity to the calcium stress, the other group (designated A_{7,8}) demonstrated a behaviour somewhat similar to that of the control group (C), one of the rats even displaying a higher resistance to the calcium stress. The behaviour of mitochondria obtained from both subgroups of alcohol-fed rats is illustrated in Figs. 9-16.

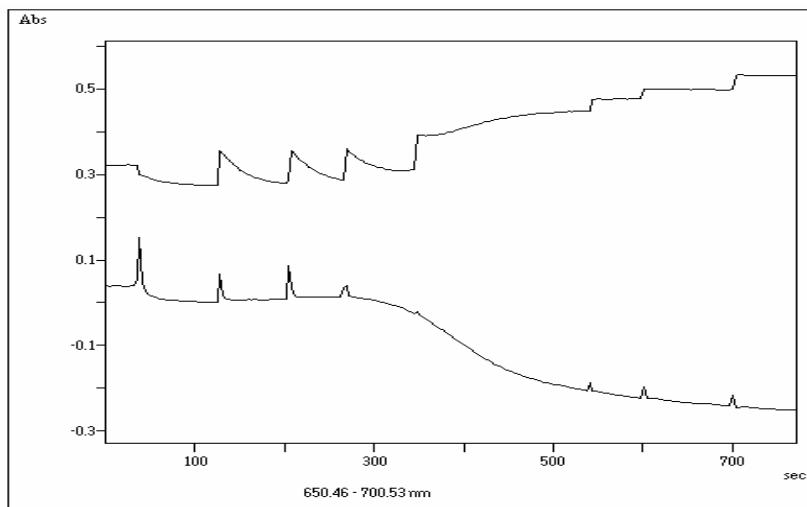


Fig. 9. Calcium fluxes (upper curve) and matrix swelling (lower curve), in mitochondria of an alcoholic rat with typical behaviour (A₁₋₆), recorded in the MS medium.

Figs. 9-12 present illustrations taken from the A₁₋₆ subgroup. The most striking observation is the increased sensitivity of mitochondria to calcium. For example, in the MS medium, where the resistance is the highest, PT occurs after the third calcium pulse (Figs. 9-10) while in the C group this happens after 6-7 pulses. The same is true for calcium release (Fig. 9) and $\Delta\Psi$ collapse (Fig. 10). The situation is worse in the ionic media. Thus, in the KSW medium, the phenomena discussed occur after a single pulse of calcium (Fig. 11). Moreover, as can be seen from Fig. 12, if the addition of calcium is a little delayed, PT and $\Delta\Psi$ collapse occur spontaneously. This demonstrates that the chronic ethanol feeding had a rather grave effect on the

liver of the animals, despite the fact that their average body weight was slightly higher than in the control animals, although not statistically significant. These results point to the fact that the mitochondrial membranes of the alcoholic rats are extremely fragile, with a high degree of ion permeability, ready to give up to slight metabolic stresses and release the components of the intermembrane space and probably even of the matrix, despite the apparent well-being of the animals.

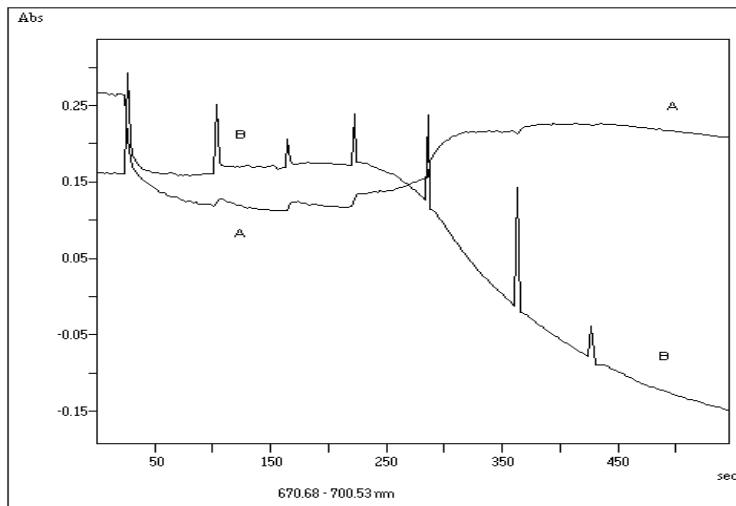


Fig. 10. Membrane potential (curve A) and matrix swelling (curve B), in mitochondria of an alcoholic rat with typical behaviour (A_{1-6}), recorded in the MS medium.

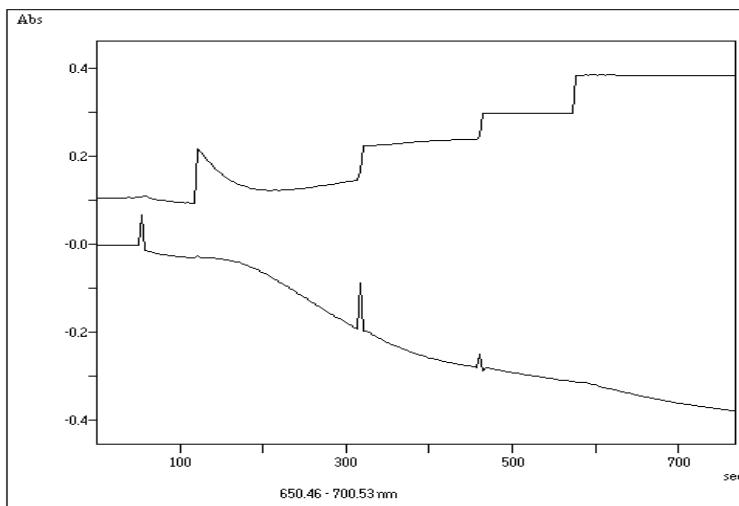


Fig. 11. Calcium fluxes (upper curve) and matrix swelling (lower curve), in the KSW medium, in mitochondria of an alcoholic rat with typical behaviour (A_{1-6}).

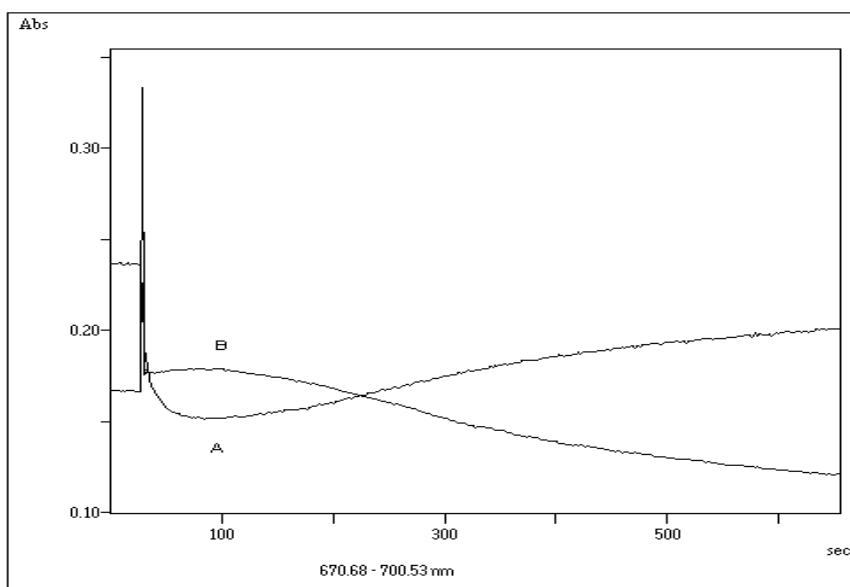


Fig. 12. Spontaneous dissipation of membrane potential (curve A) and matrix swelling (curve B), in the KSW medium, in mitochondria of an alcoholic rat with typical behaviour (A₁₋₆).

As regards the behaviour of mitochondria from atypical alcoholic rats (A_{7,8}), it is illustrated in Figs. 13-16. One can observe that in the basically nonionic medium (MS), the release of calcium and $\Delta\Psi$ collapse occur after 8 pulses of calcium and the swelling begins after 6 pulses (Figs. 13-14), while in ionic media devoid of Mg²⁺ these phenomena occur after 3-4 pulses of calcium (Figs. 15-16), similar to what we have seen in the control group. Taking into consideration the inter-individual variability in regard to the enzymes (and enzyme induction capacity) involved in the degradation of ethanol and acetaldehyde (the metabolite basically responsible for alcohol intolerance), such a behaviour is not totally unexpected. In fact, the phenomenon is known under the name of alcohol tolerance (see, for example, [16]).

Quantitative comparison regarding the sensitivity of mitochondria to calcium. As mentioned before, if we take into consideration the number of calcium pulses (12.5 μ M each or 12.5 nmols/mg protein) necessary to induce calcium release, the sensitivity of mitochondria can readily be quantified using relatively small integers or fractions equal to 0.5 and computing means that can be compared by the use of the Student *t* test or other more sophisticated tests. The first comparison of this kind that is presented in Table 1 proves that the alcoholic group (A) did not respond in a homogeneous manner to ethanol treatment and that there are statistically significant differences between the A₁₋₆ and A_{7,8} subgroups.

FUNCTIONAL CORRELATIONS IN LIVER MITOCHONDRIA OF ETHANOL-FED RATS

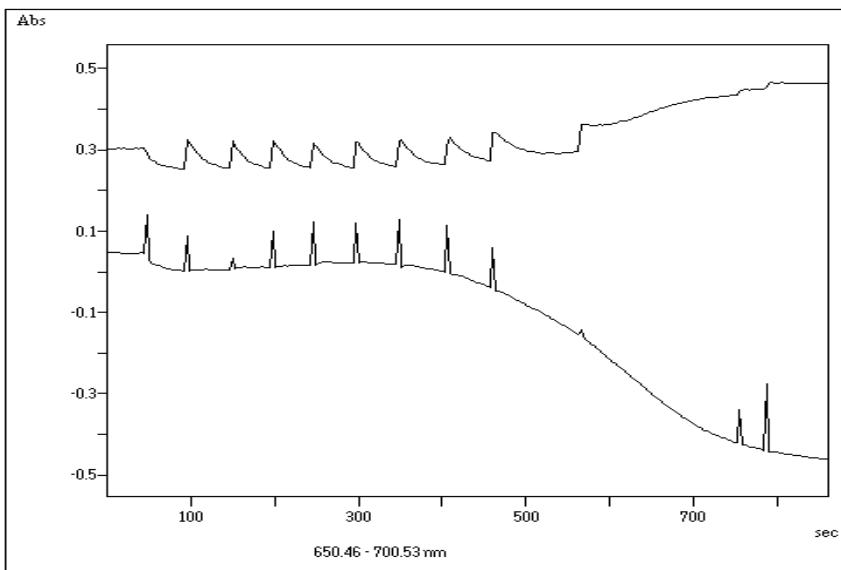


Fig. 13. Calcium fluxes (upper curve) and matrix swelling (lower curve), in mitochondria of an alcoholic rat with atypical behaviour ($A_{7,8}$), recorded in MS medium.

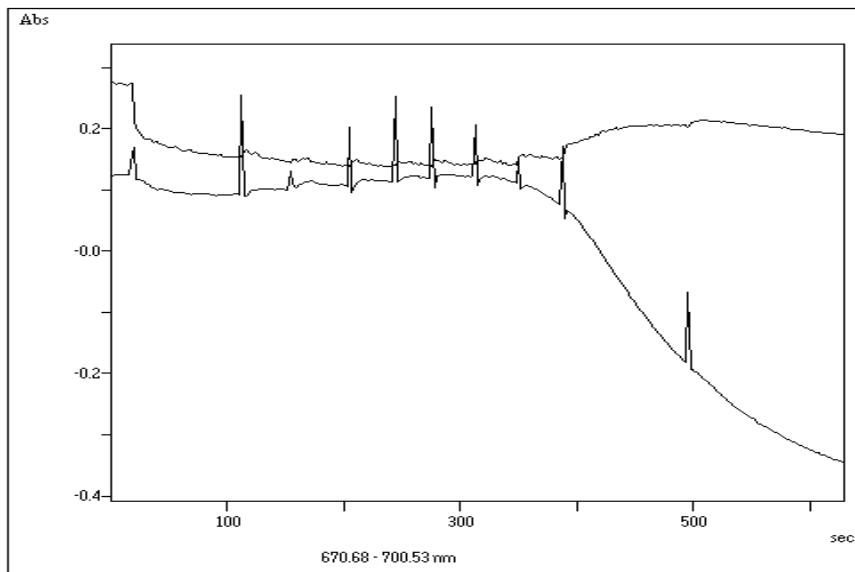


Fig. 14. Membrane potential (upper curve) and matrix swelling (lower curve), in mitochondria of an alcoholic rat with atypical behaviour ($A_{7,8}$), recorded in MS medium.

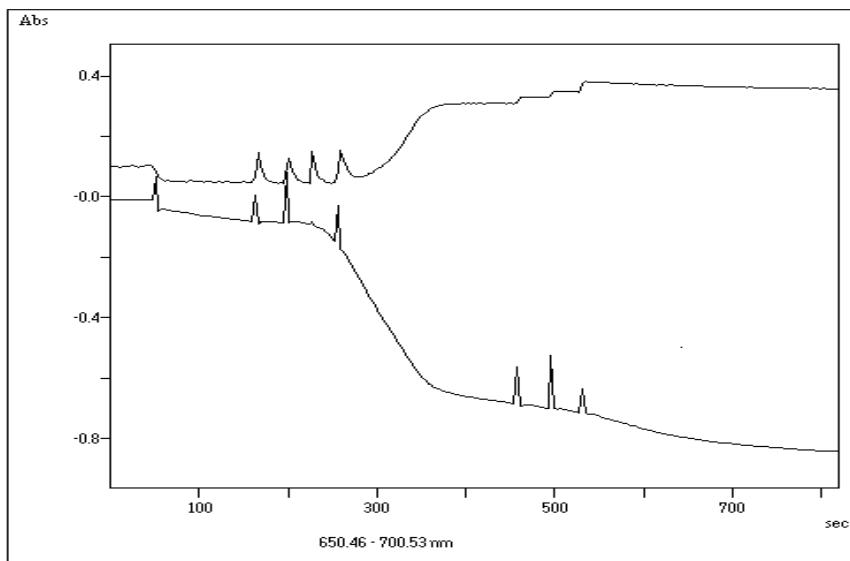


Fig. 15. Calcium fluxes (upper curve) and matrix swelling (lower curve), in mitochondria of an alcoholic rat with atypical behaviour ($A_{7,8}$), recorded in KSW medium.

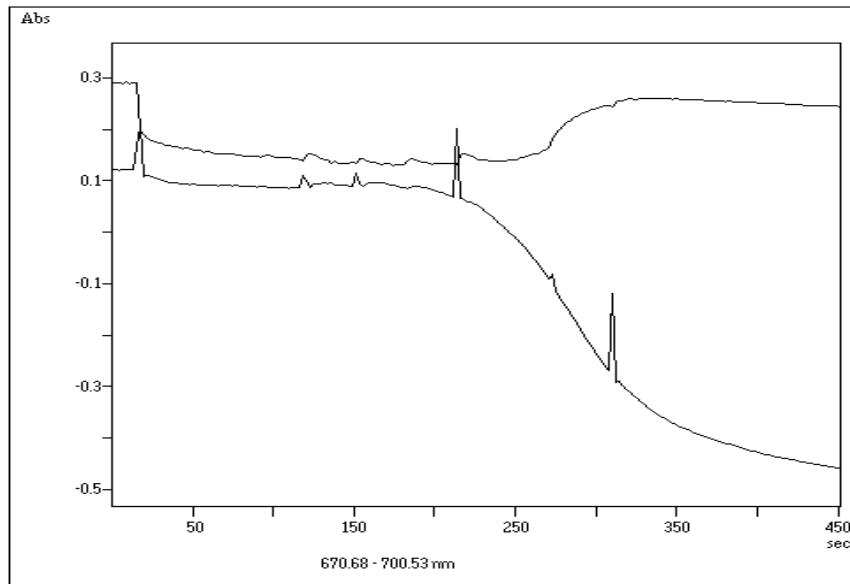


Fig. 16. Membrane potential (upper curve) and matrix swelling (lower curve), in mitochondria of an alcoholic rat with atypical behaviour ($A_{7,8}$), recorded in MSK-Mg medium.

Table 1
Comparison of mitochondrial sensitivity in different groups of alcohol-fed rats

Statistical parameter	$\bar{x} \pm SEM$		t	$n' = n_1 + n_2 - 2$	p
Suspending medium \ Group	A ₁₋₆	A _{7,8}			
MS	4.750 ± 0.750	10.000 ± 1.000	4.200	2	≥0.05
MSK	2.333 ± 0.105	5.500 ± 0.881	10.408	6	< 0.001
MSK-Mg	2.000 ± 0.129	5.000 ± 1.000	5.692	6	< 0.002
KSW	1.000 ± 0.000	3.500 ± 0.500	6.708	3	< 0.01

The fact that for the MS medium the differences, although very large, are only at the limit of significance can be explained by the small number of measurements performed in this medium. For the rest of the media, however, the differences between means are from very to highly significant. Under these circumstances it is recommended to consider the two subgroups as distinct and to compare them to the control. By doing this, we established that there is no statistical difference between A_{7,8} and C. In turn, between A₁₋₆ and C, with one exception, all the differences are statistically highly significant ($p < 0.001$), as can be seen from Table 2. The exception is again due to the small number of measurements performed in the MS medium.

Table 2
Mitochondrial sensitivity in typical alcohol-fed rats in comparison to the control

Statistical parameter	$\bar{x} \pm SEM$		t	$n' = n_1 + n_2 - 2$	p
Suspending medium \ Group	C	A ₁₋₆			
MS	7.750 ± 0.250	4.750 ± 0.750	3.795	2	≥0.05
MSK	5.214 ± 0.343	2.333 ± 0.105	7.487	11	< 0.001
MSK-Mg	3.286 ± 0.184	2.000 ± 0.129	5.520	11	< 0.002
KSW	3.143 ± 0.143	1.000 ± 0.000	9.488	8	< 0.01

These quantitative data confirm the conclusions drawn from comparing the spectrophotometric kinetic recordings, *i.e.*, the alcoholic rats are characterised by a high sensitivity to metabolic stress, specifically to that exerted by an increased calcium concentration. Even if the comparison is made with the entire alcoholic group (A), there is a very significant difference ($p < 0.01$) in the case of the most relevant medium (MSK), where the number of recordings is the largest.

We have also tried to see whether the extension of swelling at different calcium loads is also different for different groups and subgroups, because in a previous work [63] we observed a higher tendency to swelling for mitochondria of alcoholic rats, especially at high calcium loads (50-250 µM) and long incubation times (20-30 min). The conclusion is that although there are certain differences

they are in most cases not significant and not always in the expected direction. On the contrary, the amplitude of swelling is usually smaller in alcoholic rats, even though the differences are statistically significant only in a few cases, a fact that can be seen from Table 3, where our data are presented as percent of absorbance decrease with reference to the maximum absorbance (A_0), usually attained immediately after the addition of mitochondria to the complete incubation medium. We shall try an explanation of this apparent discrepancy with our previous observations in an accompanying paper [66].

Table 3
Amplitude of mitochondrial swelling (mean values) at different calcium loads and times (SEM = standard error of mean; n = number of animals tested;
 * p < 0.05; ** p < 0.01)

Group	Time (min)	1		5		10		20		30	
		[Ca ²⁺] (mM)	– ΔA/A ₀ (%) ± SEM (n)								
C	0	2.57±0.59 (8)	4.83±0.46 (8)	6.55±1.12 (8)	21.62±3.77 (8)	45.15±2.82 (6)					
	25	2.91±0.45 (8)	19.43±4.51 (8)	43.98±2.42 (8)	56.34±1.78 (8)	60.39±0.90 (5)					
	200	27.39±2.62 (7)	46.61±7.03 (7)	52.73±2.31 (7)	57.22±2.04 (7)	61.89±0.30 (2)					
A ₁₋₆	0	1.80±0.39 (6)	2.97±0.58 (6)*	3.66±0.87 (6)	17.35±4.23 (6)	45.75±4.16 (4)					
	25	4.65±1.90 (6)	22.84±6.65 (6)	40.33±4.01 (6)	49.27±3.34 (6)*	—					
	200	26.36±1.87 (6)	41.57±2.16 (6)	46.30±2.34 (6)	51.07±2.53 (6)	52.54±4.07 (4)					
A	0	1.80±0.29 (8)	3.25±0.46 (8)*	3.92±0.66 (8)	17.76±2.91 (8)	41.53±4.46 (6)					
	25	3.99±1.46 (8)	18.26±5.71 (8)	32.84±5.80 (8)	48.54±1.07(8)**	55.36±0.95(2)*					
	200	26.49±1.67 (8)	42.19±1.63 (8)	47.15±1.84 (8)	52.11±2.00 (8)	54.84±2.98 (6)					

As regards the aim of the present study, *i.e.*, trying to clarify the succession of the phenomena responsible for (or at least associated with) the release of the apoptotic factors from mitochondria, in the first place that of cyt c, necessary for establishing a correct causal relationship in the interpretation of the apoptotic process, we have succeeded in identifying indubitably the matrix swelling as the event that precedes both $\Delta\Psi$ collapse and calcium release and established that the latter events are triggered as the swelling approaches the maximum rate, *i.e.*, they are caused by the permeability transition and not vice-versa. Of course, one should not forget that our recordings are an expression of the statistical behaviour of a mitochondrial population, these phenomena stimulating each other, as more and more mitochondria suffer the permeability transition. The sigmoid aspect of the swelling curve is in fact suggestive of co-operative phenomena. Also, along with the loss of membrane potential there is a loss of mitochondrial ability to perform oxidative phosphorylation (see [10, 28, 61] and the references therein). Consequently, the lack of energy (under the form of ATP) enhances the metabolic stress and accentuates the permeability transition and all the associated phenomena, even being able to change the type of cell death, from apoptosis to necrosis (see, for example, [37, 47]). The mutual conditioning of these phenomena is proved also by the fact that when the metabolic stress is very intense, the association between the 3 phenomena studied by us is so tight that

the limited time resolution of our spectrophotometer does not allow their temporal dissociation. We observed this situation with very high concentrations of calcium, added in one pulse, or in the presence of other strong apoptotic inducers [65].

In conclusion, our present results are a promising preliminary step, which, hopefully, in corroboration with the electron microscopy and biochemical assays that we are going to perform on the material taken at different points in our experiments, will allow us to clarify other controversial problems of apoptosis, or at least of the cell death induced by alcohol, *i.e.*, of the general relationship between apoptosis and necrosis in the alcoholic liver disease.

R E F E R E N C E S

1. Adachi, M., Ishii, H., *Role of mitochondria in alcoholic liver injury*, "Free Radic. Biol. Med.", **32**, 2002, 487-491.
2. Bernardi, P., Azzone, G.F., *Regulation of Ca^{2+} efflux in rat liver mitochondria. Role of membrane potential*, "Eur. J. Biochem.", **134**, 1983, 377-383.
3. Bernardi, P., Petronilli, V., Di Lisa, F., Forte, M., *A mitochondrial perspective on cell death*, "Trends Biochem. Sci.", **26**, 2001, 112-117.
4. Bernardi, P., Pietrobon, D., *On the nature of P_i -induced Mg^{2+} -prevented Ca^{2+} release in rat liver mitochondria*, "FEBS Lett.", **139**, 1982, 9-12.
5. Bernardi, P., Scorrano, L., Colonna, R., Petronilli, V., Di Lisa, F., *Mitochondria and cell death. Mechanistic aspects and methodological issues*, "Eur. J. Biochem.", **264**, 1999, 687-701.
6. Bossy-Wetzel, E., Newmeyer, D.D., Green, D.R., *Mitochondrial cytochrome c release in apoptosis occurs upstream of DEVD-specific caspase activation and independently of mitochondrial transmembrane depolarization*, "EMBO J.", **17**, 1998, 37-49.
7. Cahill, A., Baio, D.L., Ivester, P., Cunningham, C.C., *Differential effects of chronic ethanol consumption on hepatic mitochondrial and cytoplasmic ribosomes*, "Alcoholism: Clin. Exp. Res.", **20**, 1996, 1362-1367.
8. Carafoli, E., *Calcium signaling:a tale for all seasons*, "Proc.Natl.Acad.Sci. USA", **99**, 2002, 1115-1122.
9. Crompton, M., *The mitochondrial permeability transition pore and its role in cell death*, "Biochem. J.", **341**, 1999, 233-249.
10. Cunningham, C.C., Coleman, W.B., Spach, P.I., *The effects of chronic ethanol consumption on hepatic mitochondrial energy metabolism*, "Alcohol Alcoholism", **25**, 1990, 127-136.
11. Deaciuc, I.V., D'Souza, N.B., Spitzer, J.J., *Tumor necrosis factor- α cell surface receptors of liver parenchymal and nonparenchymal cells during acute and chronic alcohol administration to rats*, "Alcoholism: Clin. Exp. Res.", **19**, 1995, 332-338.
12. Deaciuc, I.V., Fortunato, F., D'Souza, N.B., Hill, D.B., McClain, C.J., *Chronic ethanol exposure of rats exacerbates apoptosis in hepatocytes and sinusoidal endothelial cells*, "Hepatol. Res.", **19**, 2001, 306-324.
13. Deaciuc, I.V., Fortunato, F., D'Souza, N.B., Hill, D.B., Schmidt, J., Lee, E.Y., McClain, C.J., *Modulation of caspase-3 activity and Fas ligand mRNA expression in rat liver cells in vivo by alcohol and lipopolysaccharide*, "Alcoholism: Clin. Exp. Res.", **23**, 1999, 349-356.

14. Eskes, R., Antonsson, B., Osen-Sand, A., Montessuit, S., Richter, C., Sadoul, R., Mazzei, G., Nichols, A., Martinou, J.-C., *Bax-induced cytochrome c release from mitochondria is independent of the permeability transition pore but highly dependent on Mg²⁺ ions*, "J. Cell Biol.", **143**, 1988, 217-224.
15. Fernandez-Checa, J.C., Garcia-Ruiz, C., Ookhtens, M., Kapowitz, N., *Impaired uptake of glutathione by hepatic mitochondria from chronic ethanol-fed rats. Tracer kinetic studies in vitro and in vivo and susceptibility to oxidant stress*, "J. Clin. Invest.", **83**, 1991, 397-405.
16. Frank, J., Iqbal, J., Haas, M., Massaro, D., *New "rest period" protocol for inducing tolerance to high O₂ exposure in adult rats*, "Am. J. Physiol.", **257**, 1989, L226-L231.
17. French, S.W., *Effect of chronic ethanol ingestion on liver enzyme changes induced by thiamine, riboflavin, pyridoxine or choline deficiency*, "J. Nutr.", **88**, 1966, 365-378.
18. Galle, P.R., *Apoptosis in liver disease*, "J. Hepatol.", **27**, 1997, 405-412.
19. Garcia-Ruiz, C., Collel, A., Mari, M., Morales, A., Fernandez-Checa, J.C., *Direct effect of ceramide on the mitochondrial electron transport chain leads to generation of reactive oxygen species. Role of mitochondrial glutathione*, "J. Biol. Chem.", **272**, 1997, 11369-11377.
20. Garcia-Ruiz, C., Collel, A., Paris, R., Fernandez-Checa, J.C., *Direct interaction of GD3 ganglioside with mitochondria generates reactive oxygen species followed by mitochondrial permeability transition, cytochrome c release and caspase activation*, "FASEB J.", **14**, 2000, 847-858.
21. Gogvadze, V., Robertson, J.D., Zhivotovsky, B., Orrenius, S., *Cytochrome c release occurs via Ca²⁺-dependent and Ca²⁺-independent mechanisms that are regulated by Bax*, "J. Biol. Chem.", **276**, 2001, 19066-19071.
22. Goldstein, J.C., Waterhouse, N.J., Juin, P., Evan, G.I., Green, D.R., *The coordinated release of cytochrome c during apoptosis is rapid, complete and kinetically invariant*, "Nature Cell Biol.", **2** (3), 2000, 156-162.
23. Grijalba, M.T., Vercesi, A.E., Schreier, S., *Ca²⁺-induced lipid packing and domain formation in submitochondrial particles. A possible early step in the mechanism of Ca²⁺-stimulated generation of reactive oxygen species by the respiratory chain*, "Biochemistry", **38**, 1999, 13729-13287.
24. Halestrap, A.P., Woodfield, K.Y., Connern, C.P., *Oxidative stress, thiol reagents and membrane potential modulate the mitochondrial permeability transition by affecting nucleotide binding to the adenine nucleotide translocase*, "J. Biol. Chem.", **272**, 1997, 3346-3354.
25. Hartoft, W.S., Porta, E.A., *Alcohol, diet and experimental hepatic injury*, "Can. J. Physiol. Pharmacol.", **46**, 1968, 463-473.
26. Higuchi, H., Adachi, M., Miura, S., Gores, J.G., Ishii, H., *The mitochondrial permeability transition contributes to acute ethanol-induced apoptosis in rat hepatocytes*, "Hepatology", **34**, 2001, 320-328.
27. Hirsh, T., Susin, S.A., Marzo, I., Marchetti, P., Zamzami, M., Kroemer, G., *Mitochondrial permeability transition in apoptosis and necrosis*, "Cell Biol. Toxicol.", **14** (2), 1998, 141-145.
28. Hoek, J.B., *Mitochondrial energy metabolism in chronic alcoholism*, "Current Topics Bioenerg.", **17**, 1994, 197-241.
29. Ishii, H., Adachi, M., Fernandez-Checa, J.C., Cederbaum, A.I., Deaciuc, I.V., Nanji, A.A., *Role of apoptosis in alcoholic liver injury*, "Alcoholism: Clin. Exp. Res.", **27**, 2003, 1207-1212.

30. Johnson, D., Lardy, H., *Isolation of liver or kidney mitochondria*, in Colowick, S.P., Kaplan, N.O. (Eds.), *Methods in Enzymology*, Vol.10, pp.94-96, Acad. Press, New York, 1967.
31. Kim, T.-H., Zhao, Y., Barber, M.J., Kuharsky, D.K., Yin, X.-M., *Bid-induced cytochrome c release is mediated by a pathway independent of mitochondrial permeability transition pore and Bax*, "J. Biol. Chem.", **275**, 2000, 39474-39481.
32. Kroemer, G., Dallaporta, B., Resche-Rigon, M., *The mitochondrial death/life regulator in apoptosis and necrosis*, "Annu. Rev. Physiol.", **60**, 1998, 619-642.
33. Kroemer, G., Petit, P., Zamzami, N., Vayssiere, J.-L., Mignotte, B., *The biochemistry of programmed cell death*, "FASEB J.", **9**, 1995, 1277-1287.
34. Kroemer, Zamzami, N., Susin, S.A., *Mitochondrial control of apoptosis*, "Immunol. Today", **18**, 1997, 44-51.
35. Kukielka, E., Dicker, E., Cederbaum, A.I., *Increased production of reactive oxygen species by rat liver mitochondria after chronic ethanol treatment*, "Arch. Biochem. Biophys.", **309**, 1994, 377-386.
36. Kurose, I., Higuchi, H., Miura, S., Saito, H., Watanabe, N., Hokari, R., Hirokawa, M., Takashi, M., Zeki, S., Nakamura, T., Ebinuma, H., Kato, S., Ishii, H., *Oxidative stress mediated apoptosis of hepatocytes exposed to acute ethanol intoxication*, "Hepatology", **25**, 1997, 368-378.
37. Leist, M., Single, B., Castoldi, A.F., Kuhnle, S., Nicotera, P., *Intracellular adenosine triphosphate (ATP) concentration: a switch between apoptosis and necrosis*, "J. Exp. Med.", **185**, 1997, 1481-1486.
38. Lemasters, J.J.V., *Necrapoptosis and the mitochondrial permeability transition: shared pathways to necrosis and apoptosis*, "Am. J. Physiol.", **276**, 1999, G1-G6.
39. Lieber, C.S., *Role of oxidative stress and antioxidant therapy in alcoholic and non-alcoholic liver disease*, "Adv. Pharmacol.", **38**, 1997, 601-628.
40. Lieber, C.S., DeCarli, L.M., Rubin, E., *Sequential production of fatty liver, hepatitis and cirrhosis in subhuman primates fed ethanol with adequate diets*, "Proc. Natl. Acad. Sci. USA", **72**, 1975, 437-441.
41. Lieber, C.S., Jones, D.P., Nendelson, J., DeCarli, L.M., *Fatty liver, hyperlipemia and hyperuricemia produced by prolonged alcohol consumption despite adequate dietary intake*, "Trans. Assoc. Am. Phys.", **76**, 1963, 289-300
42. Marchetti, P., Castedo, M., Susin, S.A., Zamzami, N., Hirsch, T., Macho, A., Haeffner, A., Hirsh, F., Geuskens, M., Kroemer, G., *Mitochondrial permeability transition is a central coordinating event of apoptosis*, "J. Exp. Med.", **184**, 1996, 1155-1160.
43. Martinou, I., Desagher, S., Eskes, R., Antonsson, B., Andre, E., Fakan, S., Martinou, J.C., *The release of cytochrome c from mitochondria during apoptosis of NGF-deprived sympathetic neurons is a reversible event*, "J. Cell Biol.", **144**, 1999, 883-889.
44. McClain, C.J., Barve, S., Deaciuc, I., Kugelmas, M., Hill, D., *Cytokines in alcoholic liver disease*, "Semin. Liv. Dis.", **19**, 1999, 205-219.
45. Nanji, A.A., *Apoptosis and alcoholic liver disease*, "Semin. Liv. Dis.", **18**, 1998, 186-189.
46. Neuman, M.G., Shear, N.H., Cameron, R.G., Katz, G., Tiribelli, C., *Ethanol-induced apoptosis in vitro*, "Clin. Biochem.", **32**, 1999, 547-555.
47. Nicotera, P., Leist, M., *Energy supply and the shape of death in rat lymphoid cells*, "Cell Death Differ.", **4**, 1997, 435-442.
48. Orrenius, S., Zhivotovsky, B., Nicotera, P., *Regulation of cell death: The calcium-apoptosis link*, "Mol. Cell Biol.", **4**, 2003, 552-565.

49. Petit, P.X., Susin, S.A., Zamzami, N., Mignotte, B., Kroemer, G., *Mitochondria and the programmed cell death. Back to the future*, "FEBS Lett.", **396**, 1996, 7-13.
50. Petronilli, V., Penzo, D., Scorrano, L., Bernardi, P., Di Lisa, F., *The mitochondrial permeability transition, release of cytochrome c and cell death. Correlation with the duration of pore opening*, "J. Biol. Chem.", **276**, 2001, 12030-12034.
51. Pinton, P., Ferrari, D., Rapizzi, E., Di Virgilio, F., Pozzan, T., Rizzuto, R., *The Ca^{2+} concentration of the endoplasmic reticulum is a key determinant of ceramide-induced apoptosis: significance for the molecular mechanism of Bcl-2 action*, "EMBO J.", **20**, 2001, 2690-2701.
52. Reed, J.C., *Mechanisms of apoptosis*, "Am. J. Pathol.", **157**, 2000, 1415-1430.
53. Robertson, J.D., Orrenius, S., *Molecular mechanisms of apoptosis induced by cytotoxic chemicals*, "Crit. Rev. Toxicol.", **30**, 2000, 609-627.
54. Scarlett, J.L., Murphy, M.P., *Release of apoptogenic proteins from mitochondrial intermembrane space during the mitochondrial permeability transition*, "FEBS Lett.", **418**, 1997, 282-286.
55. Scorrano, L., *Divide et impera: Ca^{2+} signals, mitochondrial fission and sensitization to apoptosis*, "Cell Death Diff.", **10**, 2003, 1287-1289.
56. Skulachev, V.P., *Why are mitochondria involved in apoptosis?* "FEBS Lett." **397**, 1996, 7-10.
57. Susin, A.S., Zamzami, N., Castedo, M., Daugas, E., Wang, H.-G., Geley, S., Fassy, F., Reed, G.C., Kroemer, G., *The central executioner of apoptosis. Multiple links between protease activation and mitochondria in Fas/Apo-1/CD95- and ceramide induced apoptosis*, "J. Exp. Med.", **186**, 1997, 25-37.
58. Szalai, G., Krishnamurthy, R., Hajnoczky, G., *Apoptosis driven by IP-3 linked mitochondrial calcium signals*, "EMBO J.", **18**, 1999, 6349-6361.
59. Tarba, C., *Use of an optical probe to estimate membrane potential at subcellular level*, "Rev. Roum. Biochim.", **20**, 1983, 51-59.
60. Tarba, C., *Membrane potential in mitochondria estimated by absorption of a cyanine dye*, "Rev. Roum. Biochim.", **20**, 1983, 61-65.
61. Tarba, C., *Biomembrane, transport si energetică celulară*, Ed. Acad. Române, Bucuresti, 1996.
62. Tarba, C., *Apoptosis and necrosis: differences and similarities; the involvement of mitochondria*, "Bull. Mol. Med.", **7-8**, 2001, 85-96.
63. Tarba, C., Deaciuc, I.V., *Dose-dependent effects of bacterial endotoxin treatment on membrane permeability of ethanol-fed rats*, "Stud. Univ. Babeş-Bolyai, Biol.", **47** (2), 2002, 25-41.
64. Tarba, C., Orbai, P., *Ca-Mg-Gluconolactate treatment of allyl-alcohol intoxicated rats restores normal behaviour with respect to calcium fluxes*, "Stud. Univ. Babeş-Bolyai, Biol.", **31** (1), 1986, 38-45.
65. Tarba, C., Suărăşan, F., *Corelații între fluxurile de calciu, tranziția de permeabilitate și colapsarea potențialului de membrană în mitocondriile hepatice de şobolan: I.Efectul tratamentului cronic cu etanol*, "An. Soc. Naț. Biol. Cel.", **8** (1), 2003, 391-412.
66. Tarba, C., Suărăşan, F., *Effect of acute endotoxin administration to rats on membrane permeability and associated functional parameters of liver mitochondria*, "Stud. Univ. Babeş-Bolyai, Biol.", **49** (1), 2004, 105-128.
67. Thor, H., Hartzell, P., Orrenius, S., *Potentiation of oxidative cell injury in hepatocytes which have accumulated calcium*, "J. Biol. Chem.", **259**, 1984, 6612-6615.
68. Tsukamoto, H., Lu, S.C., *Current concepts in pathogenesis of alcoholic liver injury*, "FASEB J.", **15**, 2001, 1335-1349.

FUNCTIONAL CORRELATIONS IN LIVER MITOCHONDRIA OF ETHANOL-FED RATS

69. Von Ahsen, O., Renken, C., Perkins, G., Kluck, R.M., Bossy-Wetzel, E., Newmeyer, D.D., *Preservation of mitochondrial structure and function after Bid- or Bax-mediated cytochrome c release*, "J. Cell Biol.", **150**, 2000, 1027-1036.
70. Waterhouse, N.J., Goldstein, J.C., von Ahsen, O., Shuler, M., Newmeyer, D.D., Green, D.R., *Cytochrome c maintains mitochondrial transmembrane potential and ATP generation after mitochondrial membrane permeabilization during the apoptotic process*, "J. Cell Biol.", **153**, 2001, 319-328.
71. Xu, D., Thiele, G.M., Beckenhauer, J.L., Klassen, L.W., Sorrell, M.F., Tuma, D.J., *Detection of circulation antibodies to malondialdehyde-acetaldehyde adducts in ethanol-fed rats*, "Gastroenterol.", **115**, 1998, 686-692.
72. Zoccarato, F., Rugolo, M., Siliprandi, D., Siliprandi, N., *Correlated effluxes of adenine nucleotides, Mg²⁺ and Ca²⁺ induced in rat liver mitochondria by external Ca²⁺ and phosphate*, "FEBS Lett.", **114**, 1981, 195-199.

EFFECT OF ACUTE ENDOTOXIN ADMINISTRATION TO RATS ON MEMBRANE PERMEABILITY AND RELATED FUNCTIONAL PARAMETERS OF LIVER MITOCHONDRIA

CORNELIU TARBA* and FELICIA SUĂRĂŞAN*

SUMMARY. — Male white Wistar rats were maintained in our animal facility for 14–18 weeks and the evolution of their body weight was assessed periodically. 24 hrs before sacrificing, part of the control animals were injected with either 0.5 or 2 mg bacterial endotoxin/kg body weight. Another group of rats was chronically fed ethanol (1.5 ml of 48% alcohol/100 g body weight, daily) and some of these alcoholic rats also received 0.5 or 2 mg endotoxin/kg body weight 24 hrs prior to sacrifice. The liver of the animals was used for mitochondrial preparations and in some cases also for samples of electron microscopy. Mitochondrial membrane permeability was assessed indirectly from the matrix swelling in the presence of different concentrations of calcium. In addition, membrane potential ($\Delta\Psi$) and calcium fluxes were also monitored spectrophotometrically, as described in our previous paper (Tarba and Suărăşan [46]). The capacity of mitochondria to resist to different calcium loads was estimated qualitatively from such spectrophotometric recordings and also quantified by counting the number of calcium pulses (of 12.5 μM each) needed for the triggering of the permeability transition, $\Delta\Psi$ collapse and calcium release. Predictably, all the phenomena tested occur faster (at shorter times and/or lower calcium concentrations) in ionic media, especially in those that lack magnesium. However, statistically, no significant differences were observed between the results obtained with the two doses of endotoxin. Also, a certain degree of similarity could be observed between the ethanol-fed and the endotoxin-treated rats. On the other hand, very significant differences were observed between the mitochondria of the endotoxin-treated rats (as a single group) and of the control rats. Thus, mitochondria from endotoxin-treated animals are more sensitive to calcium loads than those of the (non-alcoholic) control, they approaching the behaviour of the mitochondria from alcohol-fed rats, although the sensitivity to calcium is somewhat lower. This lower sensitivity can be seen in ionic media, especially in the swelling medium (KSW), where the differences are highly significant ($p<0.001$). The mitochondria of double-treated animals (endotoxin-injected alcohol-fed rats) show a more complex behaviour and less homogeneity. Their sensitivity to metabolic stress is in general higher than that of the simply endotoxin-treated animals, but comparable to that of the simply ethanol-fed rats, although in ionic suspending media membrane permeability actually tends to be lower, a fact that must be attributed to the endotoxin treatment. A certain degree of dissociation of the $\Delta\Psi$ collapse from the matrix swelling and calcium release was also observed. The fact that $\Delta\Psi$ collapse occurs apparently at a later time than the matrix swelling and calcium release is interpreted by us as an indication of a differential sensitivity of different types of hepatic mitochondria to the metabolic stress factors, such as chronic ethanol feeding and acute LPS administration.

*Babeş-Bolyai University, Department of Experimental Biology, 400006 Cluj-Napoca, Romania.
E-mail: cntarba@biochem.ubbcluj.ro

As discussed in a more extensive way in previous articles [44, 46], chronic ethanol administration has been generally associated with specific hepatic structural and functional alterations known as the alcoholic liver disease (ALD). However, the exact mechanism of this disease is not entirely known. For decades, ALD has been attributed to necrotic events associated, among other things, with the production of proinflammatory cytokines, such as interleukin-1 (IL-1) and tumour necrosis factor- α (TNF- α) [6, 10, 25, 47]. However, relatively recent data indicate that under certain circumstances both the acute and chronic ethanol administration is associated with a process of liver apoptosis [1, 7, 14, 20, 30, 34] which, paradoxically, can facilitate latter necrotic events leading to ALD [1, 7, 8, 16, 26, 33, 47]. What exactly determines the type of cell death (apoptosis or necrosis) is not entirely understood, although it seems to be related finally to the amount of ATP produced by mitochondria, as described elsewhere [21, 22, 31]. Different properties, such as the rate and the intensity of the triggering agents, as well as the presence/absence of certain factors having sensitising, modulating or even protective effects determine the final outcome. Among other factors, for example, the excessive alcohol consumption increases the absorption from the gut of bacterial endotoxin or lipopolysaccharide (LPS) [3, 27, 32, 37], a toxin produced by Gram-negative bacteria, which is associated with the stimulation of the cytokine secreting cells (the hepatic macrophages or Kupffer cells) [5, 8, 16, 25, 26]. LPS seems to exert its effects mainly through TNF-related events [6, 8, 13, 16, 25, 47], although contradictory results are sometimes reported by different authors (see [23, 28, 33]). On the other hand, as anticipated above, progresses made during the last decade in the study of apoptosis have placed mitochondria in a central position in controlling this very complicated process [1, 2, 10, 16, 18, 19, 39]. It seems that the main event in the control of apoptosis by mitochondria is the formation of the so-called permeability transition pore (PTP), which represents the structural basis for a drastic change in membrane permeability associated with the release of the apoptogenic factors, including cytochrome c [4, 14, 18, 19, 24, 35, 36, 39]. Many studies have tried lately to link alcohol consumption and/or endotoxicosis to changes in mitochondrial membrane permeability, suggesting an increased permeability [14, 20, 26, 33, 34], but, in our opinion, the results are controversial [38, 43-46].

The present article is part of a larger study aimed at establishing the order of events leading to the release of the apoptogenic factors from the mitochondrial intermembrane space and uses alcohol and endotoxin as metabolic stress factors for the induction of apoptosis and/or necrosis in animals kept under more natural conditions (see *Material and Methods*) than usually reported in the literature. In a previous article [46] we dealt mainly with the effect of alcohol, whereas the present study is dedicated to the effect of acute endotoxin administration, both in control (*i.e.*, non-alcoholic) and alcoholic rats. Two preliminary reports on our results have also appeared [38, 45].

Material and methods. *Animal treatments and preparation of mitochondria.*

Male white Wistar rats were kept under normal conditions in our animal facility for 14-18 weeks, starting from an average weight of 120 g/individual, while the evolution of their weight was assessed periodically. The rats were kept on a normal diet (a premix containing all the ingredients of the Larsen diet), with free access to water. One group served as control (C), while in another group (A) each rat was supplemented daily with 1.5 ml of 48% ethanol/100 g body weight. In both groups, 24 hrs before the sacrifice, part of rats were injected with either 0.5 or 2 mg LPS/kg body weight (b.w.), administered in the caudal vein. All the rats were fasted for the last 24 hrs and sacrificed by decapitation after a slight anaesthesia. Small pieces of liver were taken in some cases and prepared for electron microscopy, while the rest of the liver was used for preparation of mitochondria, essentially according to Johnson and Lardy [17], in a medium containing 200 mM mannitol, 70 mM sucrose, 5 mM Hepes-KOH (pH 7.37) and 0.5 mM Na-EDTA. The washing medium (which also served as a preserving medium) lacked the chelating agent (EDTA).

Spectrophotometric measurements. Using the principles and methodology described in our previous paper [46], a diode-array spectrophotometer (Specord S 100B, Analytik Jena, Germany) was employed for measurements of membrane potential and/or calcium fluxes in parallel with the matrix swelling, while the extension (amplitude) of swelling was quantitatively assessed on separate recordings, where the reaction was triggered by the addition of mitochondrial protein itself. The suspending media used for incubations are designated as follows: MS (210 mM mannitol, 70 mM sucrose, 1.5 mM MgCl₂); MSK (110 mM mannitol, 40 mM sucrose, 65 mM KCl, 1.5 mM MgCl₂); MSK-Mg (the same as MSK, but without Mg). All these media also contained 5 mM Hepes (pH 7.37) and 1 mM KP_i. Mg (as MgCl₂) and phosphate were usually added directly to the cuvettes. In addition, a special swelling medium (KSW) was also used. It contained 100 mM KCl, 50 mM sucrose, 10 mM Hepes and 5 mM KP_i. Different amounts of CaCl₂ were either added gradually to the spectrophotometer cuvettes, where the incubation took place, or in one pulse, up to the desired final concentration. Since succinate (2.5 mM) was used as a respiratory substrate, 8 µM rotenone was also present in the cuvettes. Although we made some attempts at obtaining precise values of membrane potential amplitude and calcium concentrations (as described elsewhere [44]), the most reliable quantitative result was reached by counting the number of calcium pulses (of 12.5 µM each) necessary to be added to the mitochondrial suspension in order to induce the calcium release, a phenomenon which is always more-or-less associated with the permeability transition (as monitored by matrix swelling) and ΔΨ collapse. The statistical significance of the differences between groups was computed by the Student *t* test.

Results and discussion. *Mitochondria of non-alcoholic LPS-injected rats.*

Based on previous experiments in which we observed a dose-dependent response to acute LPS administration [43], we selected two different concentrations of endotoxin to be tested: 0.5 mg LPS/kg body weight (b.w.) and 2 mg LPS/kg b.w. Although there are clear differences in comparison with the non-alcoholic group (control, C), the

differences between the two LPS-injected subgroups are in general small. The spectrophotometric results obtained in several media with the two subgroups (designated as 0.5-LPS and 2-LPS, respectively) are compared in Figs. 1-10.

From Figs. 1 and 2 we can compare the behaviour of mitochondria belonging to the two subgroups, as regards the calcium fluxes, whereas from Figs. 3 and 4 the corresponding aspects of the membrane potential can be compared. It can be seen that not only the number of calcium pulses taken up by mitochondria before the beginning of calcium release and $\Delta\Psi$ collapse are very close (if not identical), but the general aspects of the recordings are similar, including the lower curves which represent the matrix swelling. At the same time, if we compare these recordings with those of the control rats (see our previous paper [46]), a higher sensitivity to calcium can be observed in the present experiment.

In general, the similarity of behaviour of the two LPS subgroups is also preserved in media with a higher ionic content (Figs. 5-10), where the number of calcium pulses taken up by mitochondria is smaller, but certain subtle differences also become apparent, as can be best seen from comparing Figs. 7 and 8. An obvious dissociation between the initiation of the permeability transition (absorbance decrease on the lower curve) and the beginning of $\Delta\Psi$ collapse (absorbance increase on the upper curve) can be observed in Fig. 8. This type of behaviour is in general more evident for mitochondria of the 2-LPS subgroup, especially in media lacking Mg^{2+} (*i.e.*, MSK-Mg and KSW). Figs. 9 and 10 show the similarity of behaviour of the two subgroups with respect to calcium fluxes.

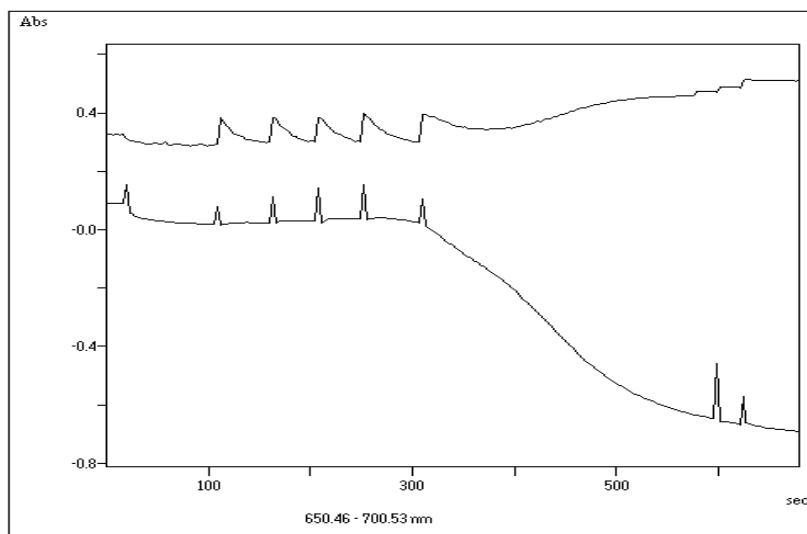


Fig. 1. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MS medium, in a preparation belonging to a rat from the 0.5-LPS subgroup. In all figures, the first spike is associated with the injection of succinate (2.5 mM), the rest being associated with $CaCl_2$ pulses of 12.5 μM each, except those added after Ca^{2+} release (or $\Delta\Psi$ collapse).

ENDOTOXIN TREATMENT AND FUNCTIONAL PARAMETERS OF LIVER MITOCHONDRIA

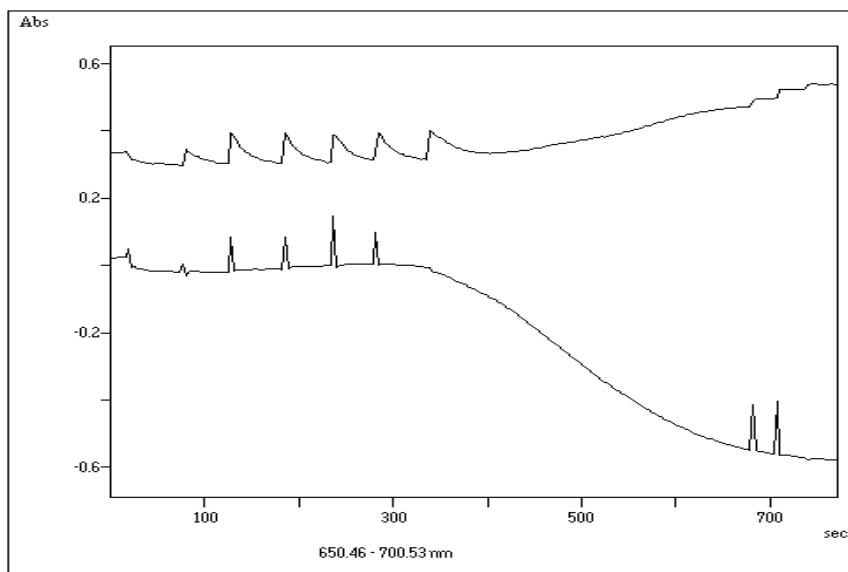


Fig. 2. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MS medium, in a preparation belonging to a rat from the 2-LPS subgroup.

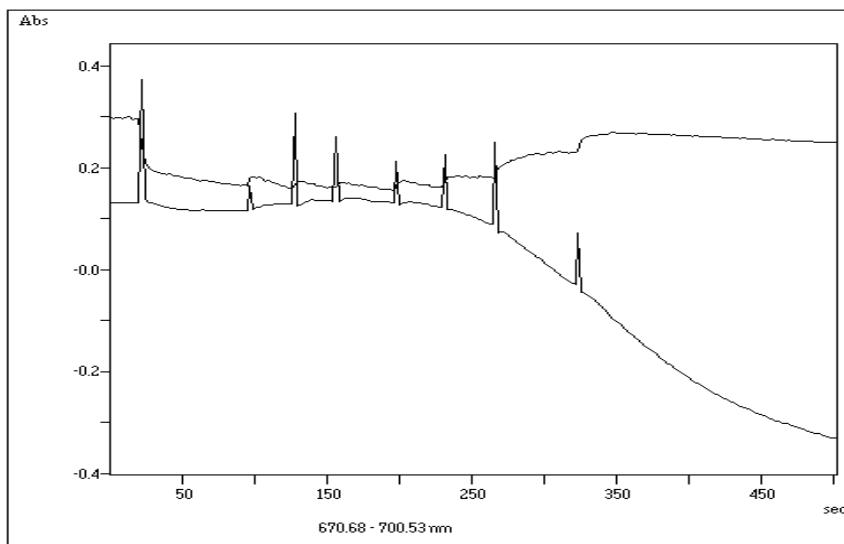


Fig. 3. Membrane potential (upper curve) and mitochondrial swelling (lower curve) in the MS medium, in a preparation belonging to a rat from the 0.5-LPS subgroup. Note the relatively close association between the initiation of $\Delta\Psi$ collapse (absorbance increase on the upper curve) and of the mitochondrial swelling (absorbance decrease on the lower curve).

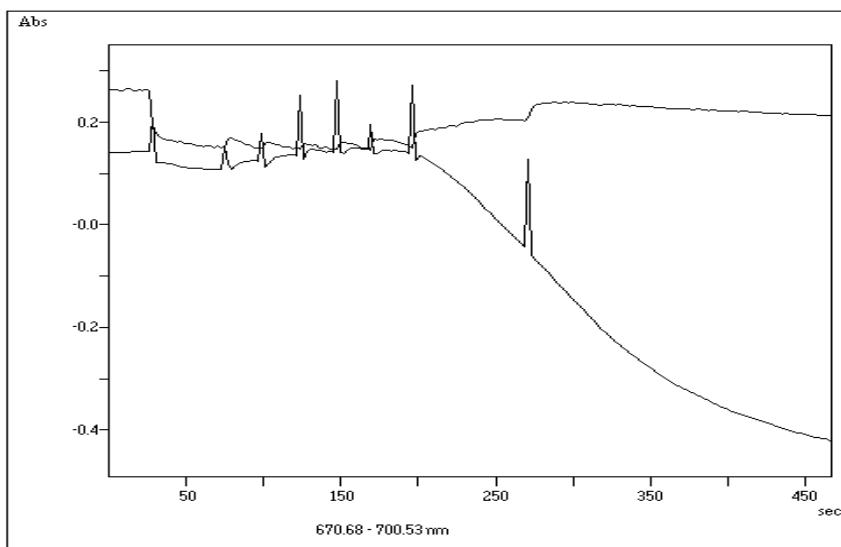


Fig. 4. Membrane potential (upper curve) and mitochondrial swelling (lower curve) in the MS medium, in a preparation belonging to a rat from the 2-LPS subgroup.

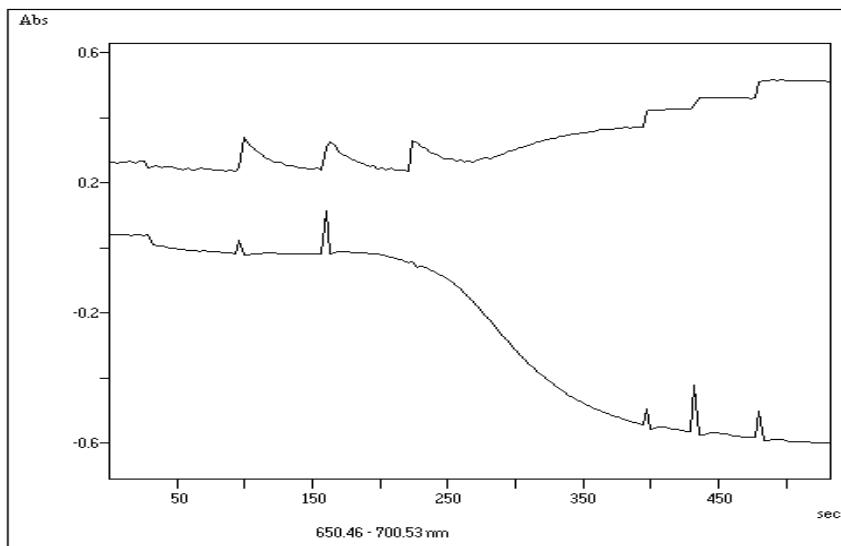


Fig. 5. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MSK medium, in a preparation belonging to a rat from the 0.5-LPS subgroup.

Note the smaller number of calcium pulses (than in the MS medium) to which mitochondria can resist (cf. Fig. 1).

ENDOTOXIN TREATMENT AND FUNCTIONAL PARAMETERS OF LIVER MITOCHONDRIA

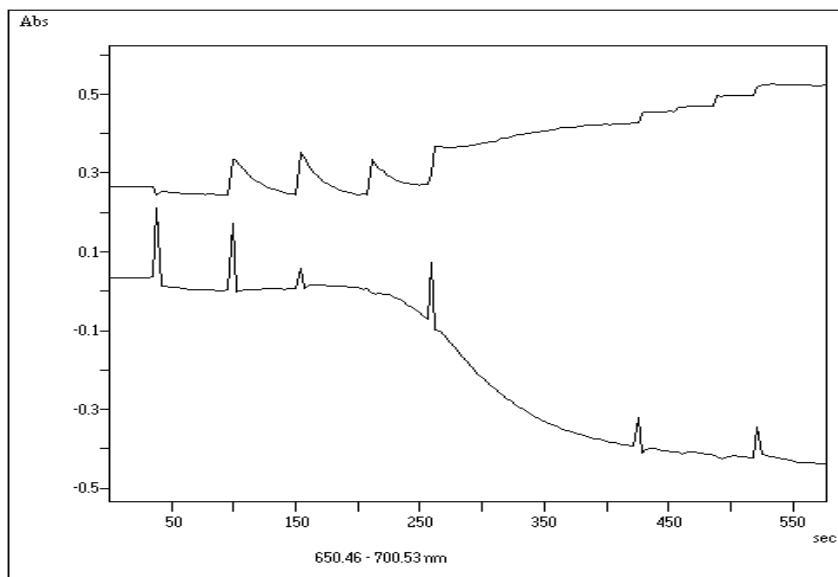


Fig. 6. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MSK medium, in a preparation belonging to a rat from the 2-LPS subgroup.

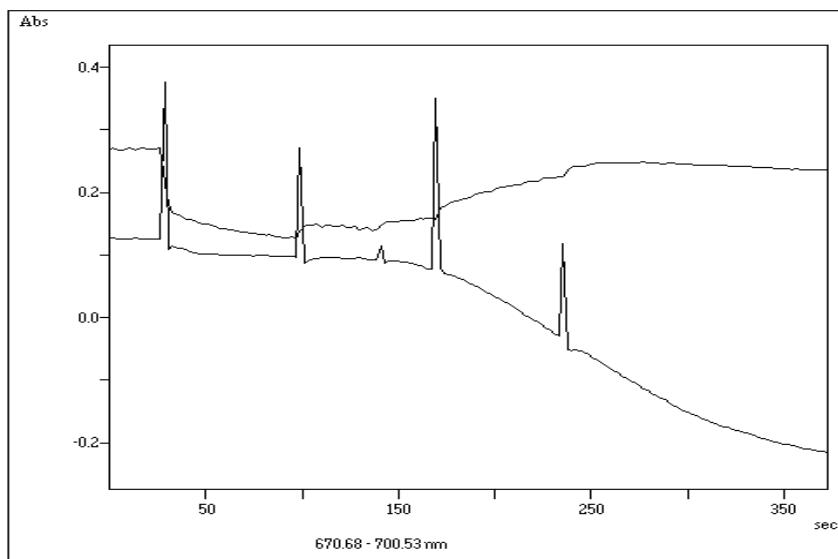


Fig. 7. Membrane potential (upper curve) and mitochondrial swelling (lower curve) in the MSK medium, in a preparation belonging to a rat from the 0.5-LPS subgroup.

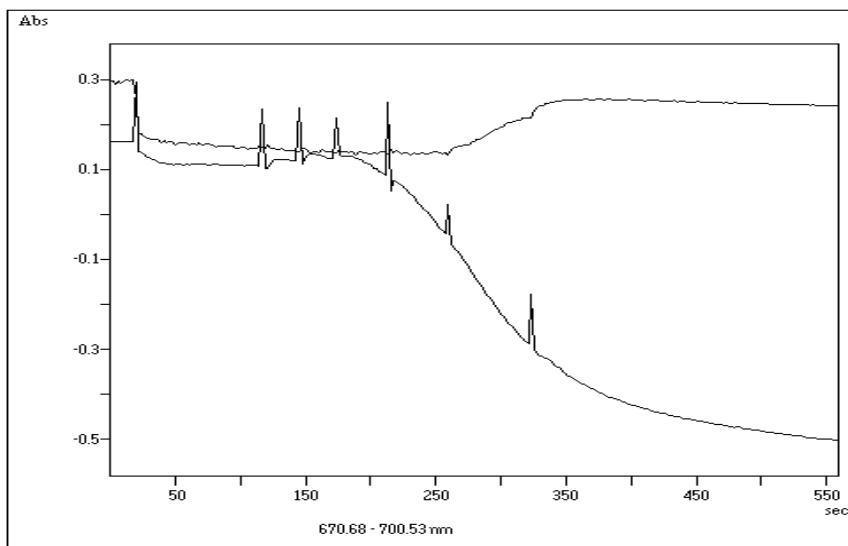


Fig. 8. Membrane potential (upper curve) and mitochondrial swelling (lower curve) in the MS medium, in a preparation belonging to a rat from the 2-LPS subgroup. Note the dissociation between the initiation of swelling and the beginning of $\Delta\Psi$ collapse.

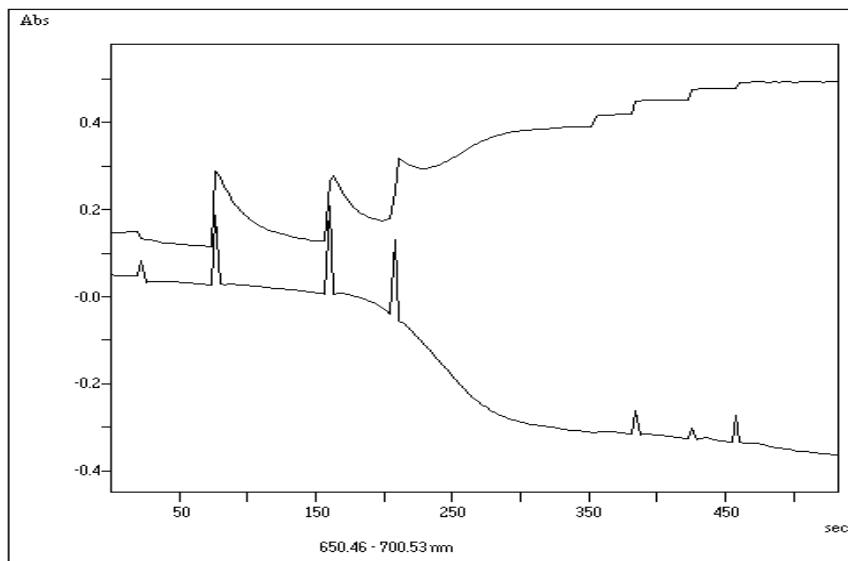


Fig. 9. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MSK-Mg medium, in a preparation belonging to a rat from the 0.5-LPS subgroup.

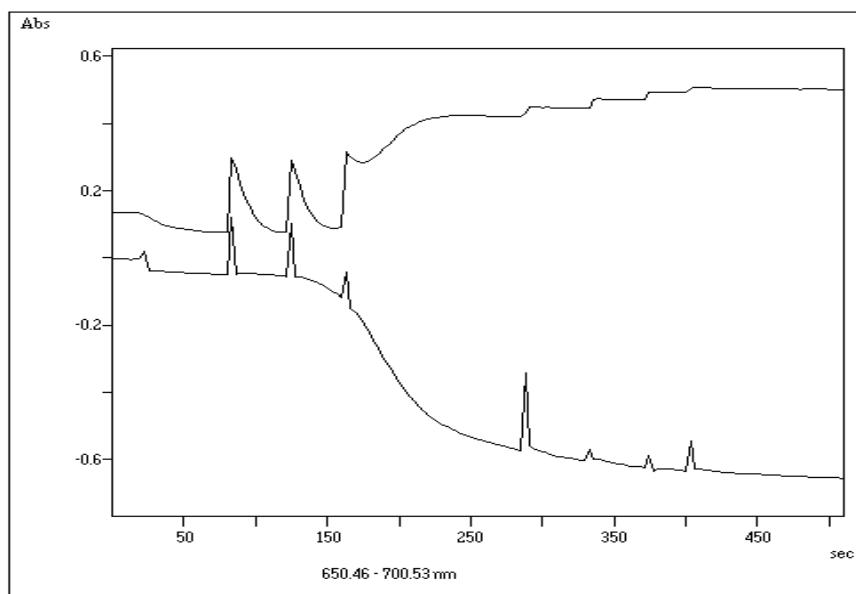


Fig. 10. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MSK-Mg medium, in a preparation belonging to a rat from the 2-LPS subgroup.

If a strictly quantitative comparison is made, in terms of the number of calcium pulses to which mitochondria resist before starting to swell, the differences between the two LPS subgroups are not statistically significant. Therefore, for comparisons with other groups, we pooled together the two LPS subgroups. Statistically, very significant differences are obtained in this way if the comparison is made with the control group (C), as can be observed from Table 1, whereas if the comparison is made with the alcoholic group (A) the differences are smaller and significant only in MSK and KSW media, as can be seen from Table 2. It must be stressed, however, that the differences are negative in comparison with the control and positive in comparison to the alcohol-fed rats, meaning that the effect of the acute LPS administration, even though obvious, is less strong than that of the chronic ethanol feeding.

Table I
Mitochondrial sensitivity of the LPS-treated group as compared to the control

Statistical parameter	$\bar{x} \pm SEM$		T	$n' = n_1 + n_2 - 2$	p
Suspending medium \ Group	C	LPS			
MS	7.750 ± 0.250	5.250 ± 0.339	3.981	6	< 0.01
MSK	5.214 ± 0.343	3.125 ± 0.190	7.567	19	< 0.001
MSK-Mg	3.286 ± 0.184	2.300 ± 0.200	3.567	10	< 0.01
KSW	3.143 ± 0.143	2.083 ± 0.091	6.122	11	< 0.001

Table 2
Mitochondrial sensitivity of the LPS-treated group as compared to the alcohol-fed group

Statistical parameter	$\bar{x} \pm SEM$		T	$n' = n_1 + n_2 - 2$	p
Suspending medium \ Group	A ₁₋₆	LPS			
MS	4.750 ± 0.750	5.250 ± 0.339	0.750	6	>> 0.1
MSK	2.333 ± 0.105	3.125 ± 0.190	4.683	18	< 0.001
MSK-Mg	2.000 ± 0.129	2.300 ± 0.200	1.303	9	> 0.1
KSW	1.000 ± 0.000	2.083 ± 0.091	8.878	7	< 0.001

Approximately the same conclusions can be drawn if the amplitude of swelling (calculated as percent of absorbance decrease relative to the maximum absorbance) is compared at different times, especially for higher calcium loads (25 and 200 μ M). From Table 3, one can see significant differences in comparison to the control (C), whereas the differences are in general not significant if comparison is made with the ethanol-fed group (not presented).

Table 3
Amplitude of mitochondrial swelling (mean values) at different calcium loads and times (SEM = standard error of mean; n = number of animals tested; * p < 0.05; ** p < 0.01)

Time (min)	Group	[Ca ²⁺] (mM)	1	5	10	20	30
			$-\Delta A/A_0$ (%) ±SEM (n)				
C	0	2.57±0.59 (8)	4.83±0.46 (8)	6.55±1.12 (8)	21.62±3.77 (8)	45.15±2.82 (6)	
	25	2.91±0.45 (8)	19.43±4.51 (8)	43.98±2.42 (8)	56.34±1.78 (8)	60.39±0.90 (5)	
	200	27.39±2.62 (7)	46.61±7.03 (7)	52.73±2.31 (7)	57.22±2.04 (7)	61.89±0.30 (2)	
LPS	0	2.23±0.22 (14)	3.44±0.50 (14)	5.81±1.41 (14)	26.25±2.46 (14)	41.83±2.99 (6)	
	25	2.83±0.47 (13)	20.63±4.20 (13)	34.53±3.69 (13)	45.10±2.79 (11)**	56.60±1.29 (4)*	
	200	22.28±2.07 (14)	39.02±2.10 (14)	43.64±2.38 (14)*	48.77±2.27 (14)*	54.84±2.98 (6)	

Mitochondria of alcohol-fed LPS-injected rats. The spectrophotometric results with this type of mitochondria are presented in Figs. 11-22. The alcoholic rats injected with 0.5 mg LPS/kg b.w. and 2 mg LPS/kg b.w. are designated as A+0.5LPS and A+2LPS, respectively. Although there are certain small differences between the two subgroups, they turned out to be statistically not significant. However, as in the case of simply alcohol-fed rats (see [46]), we observed (in both subgroups injected with LPS) the presence of rats with atypical responses. Therefore, this time, the figures are so arranged as to make obvious the differences between the rats with typical (A+LPS)_t and atypical (A+LPS)_a responses, for each subgroup. Differences can be observed in all the media and for all the parameters monitored or measured.

Illustrations from the typical and atypical (A+0.5LPS) subgroups are presented in Figs. 11-16. Figs. 11-12 and 13-14 present aspects of the calcium fluxes and membrane potential, respectively, each associated with the matrix swelling, in the MSK medium, while Figs. 15-16 present aspects of the calcium fluxes and matrix swelling of the same subgroups in KSW medium. It can be seen that the calcium load tolerated by the mitochondria of the atypical subgroup is higher than that tolerated by the typical subgroup. In addition, a certain degree of dissociation can be observed between the moments of initiation of matrix swelling and $\Delta\Psi$ collapse, especially in ionic media, for mitochondria of the atypical subgroup (see Fig. 14). There is also a difference in the extent of matrix swelling at relatively low calcium loads, as can be seen from Figs. 15-16, the amplitude of swelling being larger in mitochondria of the atypical subgroup.

Figs. 17-22 present illustrations taken from the two (A+2LPS) subgroups. From Figs. 17-20, one can see the really large differences, recorded in the MS medium, existing between the typical and atypical subgroups, regarding the number of calcium pulses that mitochondria are able to take up before the triggering of the matrix swelling, $\Delta\Psi$ collapse and calcium release. Figs. 21-22 demonstrate that such differences are also evident in ionic media devoid of Mg^{2+} , such as MSK-Mg, although they are not so impressive as in a basically non-ionic medium (MS).

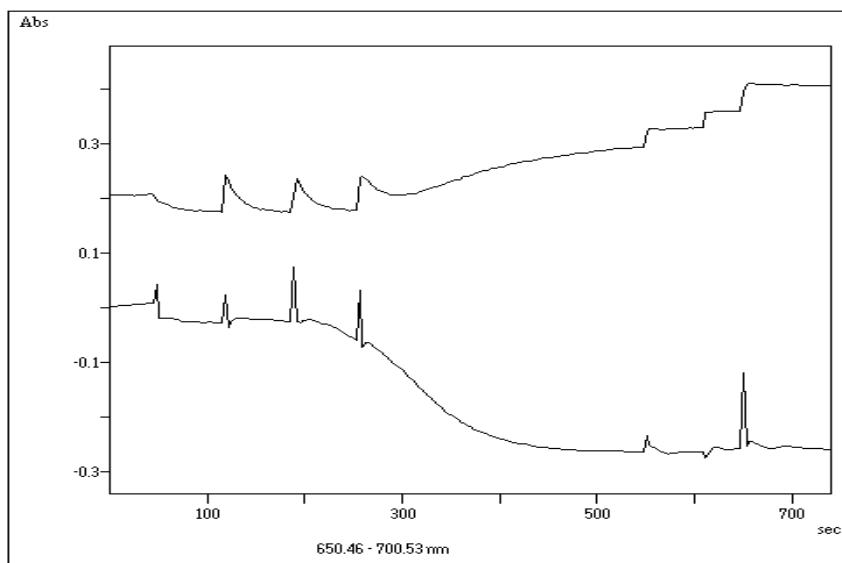


Fig. 11. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MSK medium, in a preparation belonging to a rat from the (A+0.5LPS), subgroup.

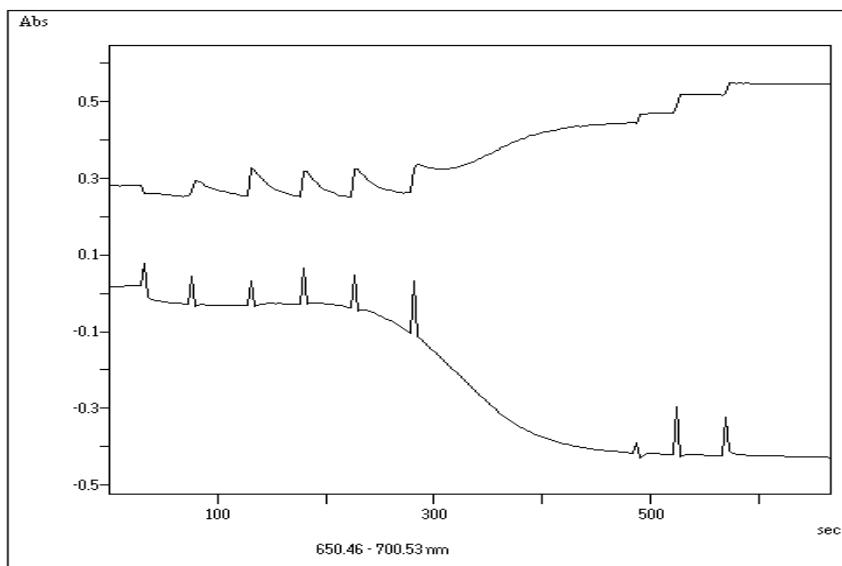


Fig. 12. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MSK medium, in a preparation belonging to a rat from the $(A+0.5LPS)_a$ subgroup.

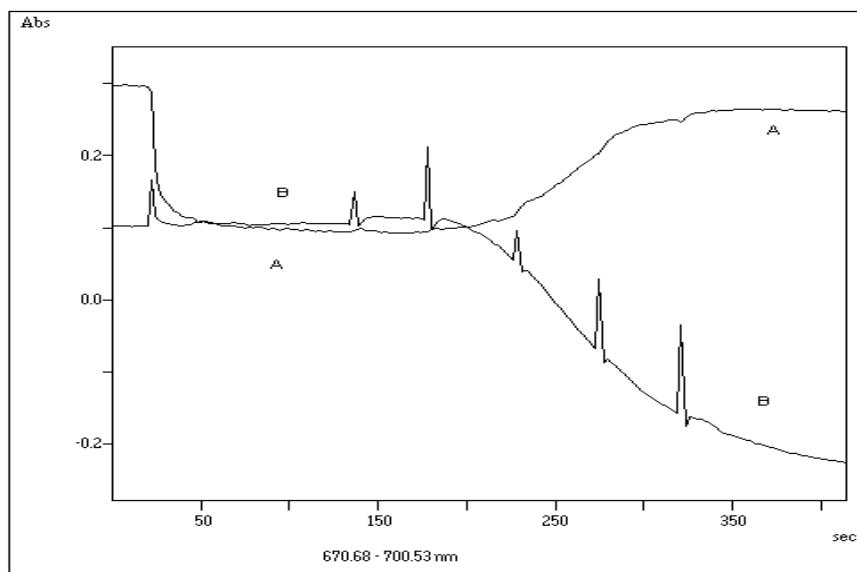


Fig. 13. Membrane potential (curve A) and mitochondrial swelling (curve B) in the MSK medium, in a preparation belonging to a rat from the $(A+0.5LPS)_t$ subgroup.

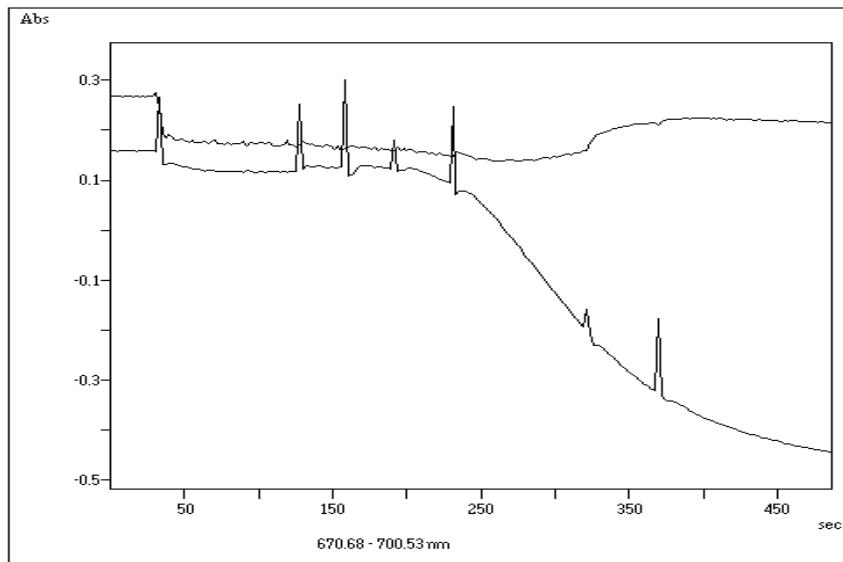


Fig . 1 4 . Membrane potential (upper curve) and mitochondrial swelling (lower curve) in the MSK medium, in a preparation belonging to a rat from the (A+0.5LPS)_a subgroup. Note the dissociation between the initiation of swelling and the beginning of $\Delta\Psi$ collapse.

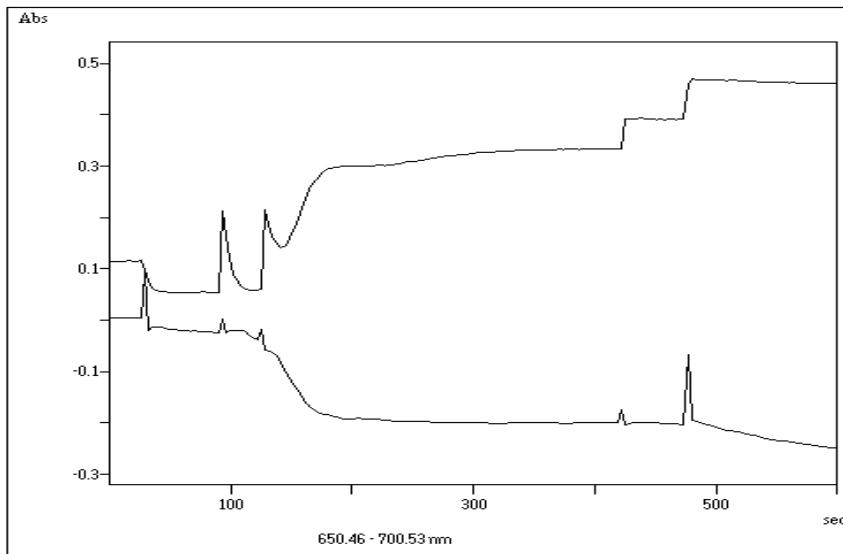


Fig . 1 5 . Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in KSW medium, in a preparation belonging to a rat from the (A+0.5LPS)_b subgroup. Note the small number of Ca^{2+} pulses to which the mitochondria resist and a rapid but shallow swelling, until further Ca^{2+} is added.

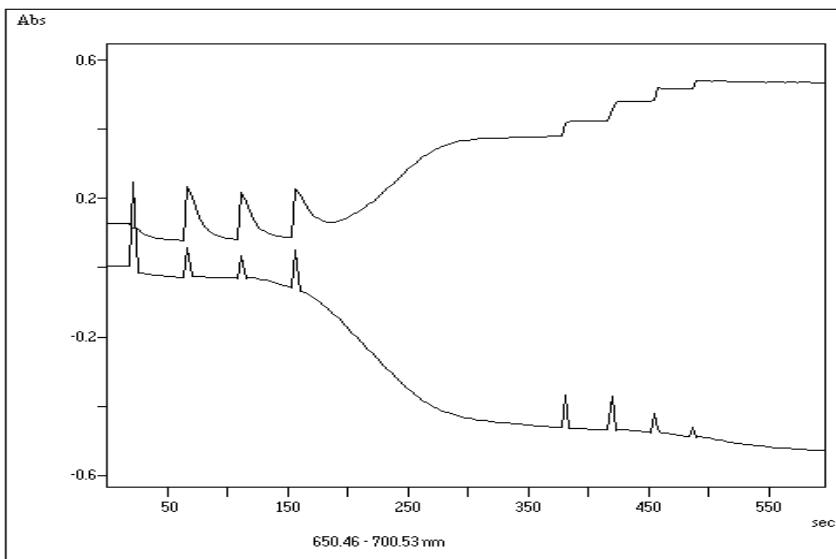


Fig. 16. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in KSW medium, in a preparation belonging to a rat from the (A+0.5LPS)_a subgroup.

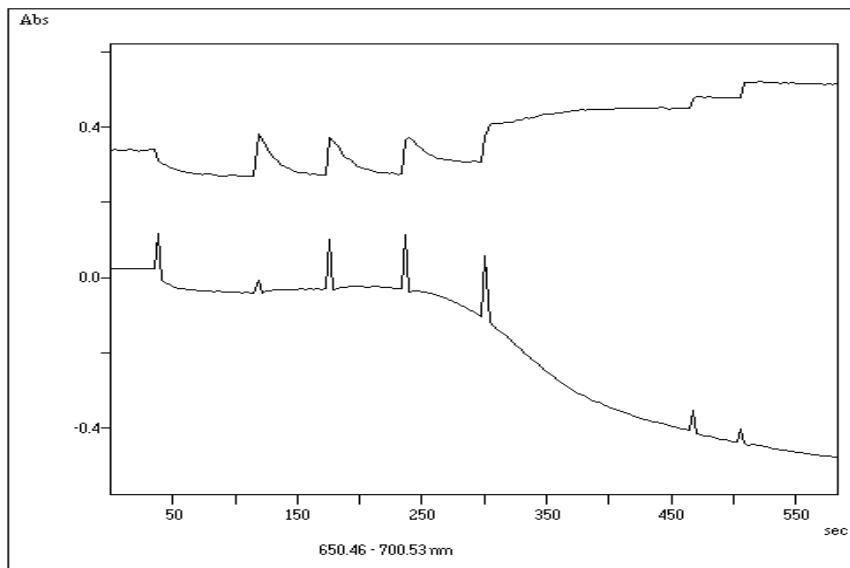


Fig. 17. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MS medium, in a preparation belonging to a rat from the (A+2LPS)_t subgroup.

ENDOTOXIN TREATMENT AND FUNCTIONAL PARAMETERS OF LIVER MITOCHONDRIA

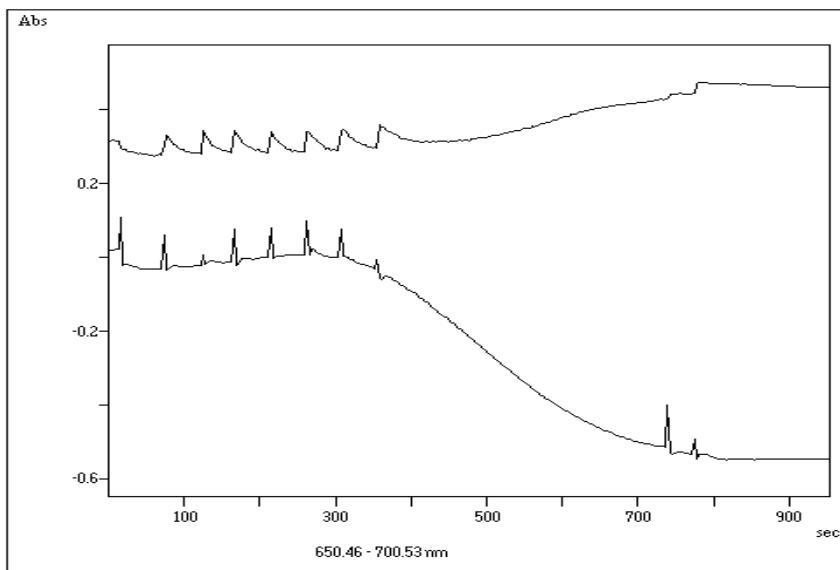


Fig. 18. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MS medium, in a preparation belonging to a rat from the $(A+2LPS)_a$ subgroup. Note the high resistance of mitochondria to calcium loading.

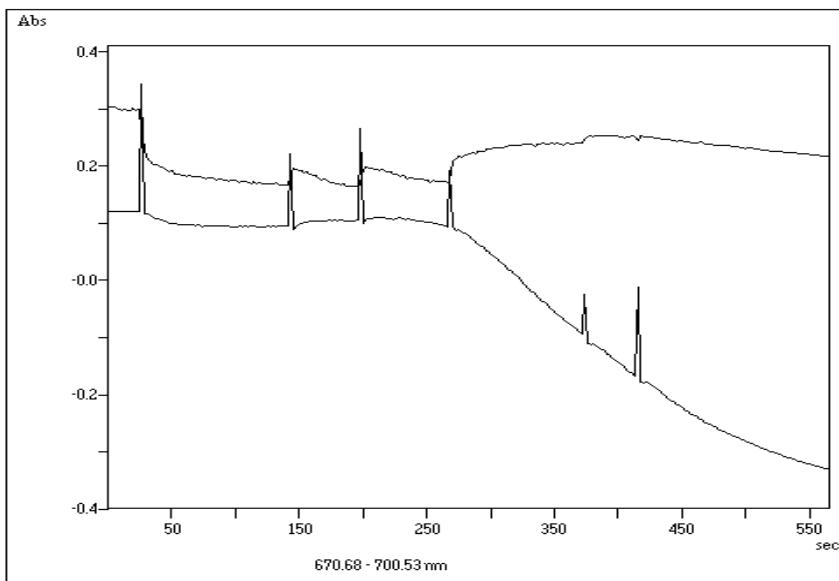


Fig. 19. Membrane potential (upper curve) and mitochondrial swelling (lower curve) in the MS medium, in a preparation belonging to a rat from the $(A+2LPS)_t$ subgroup.

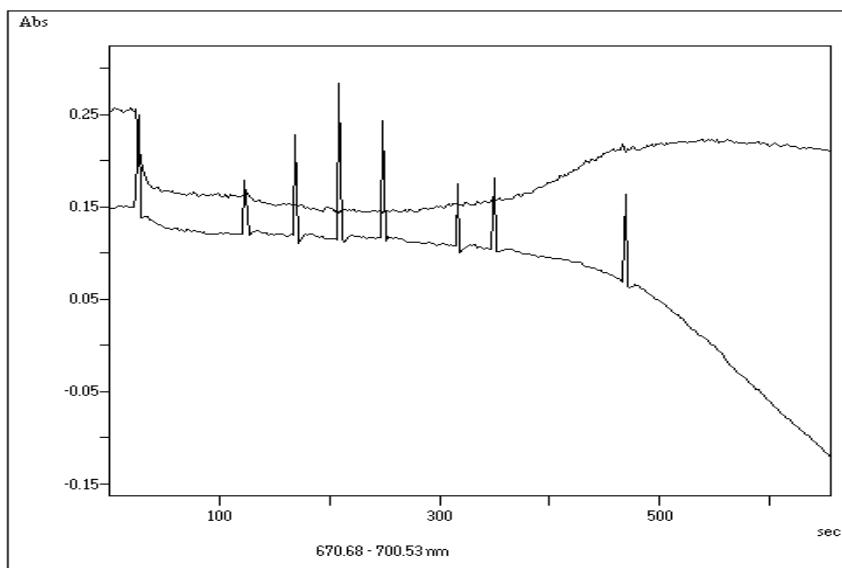


Fig. 20. Membrane potential (upper curve) and mitochondrial swelling (lower curve) in the MS medium, in a preparation belonging to a rat from the (A+2LPS)_a subgroup.

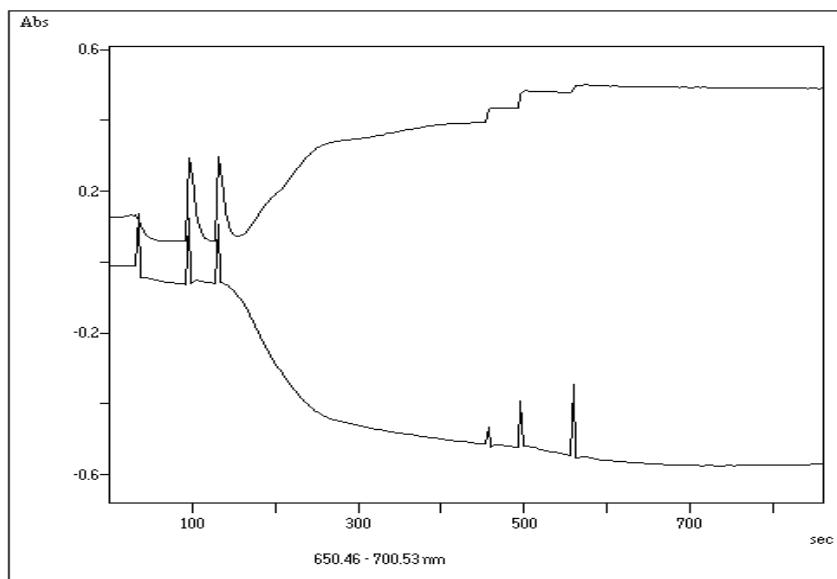


Fig. 21. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MSK-Mg medium, in a preparation belonging to a rat from the (A+2LPS)_t subgroup.

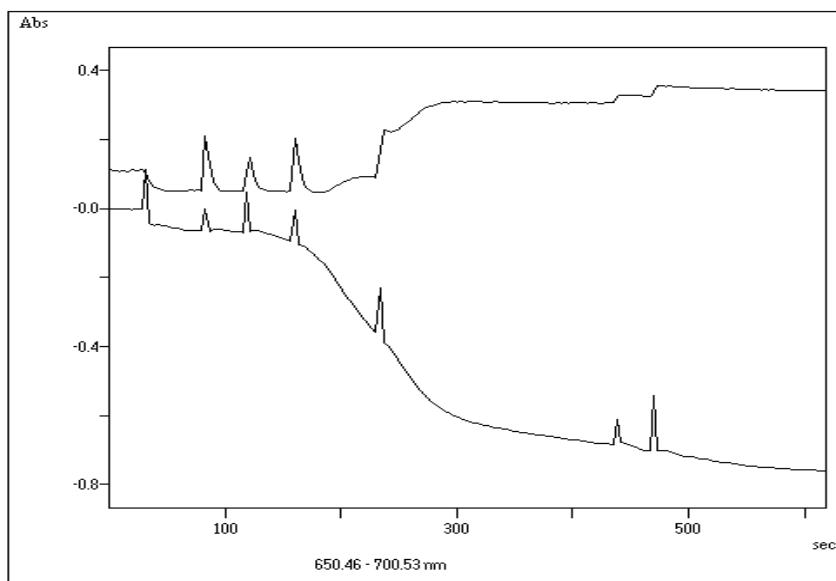


Fig. 22. Calcium fluxes (upper curve) and mitochondrial swelling (lower curve) in the MSK-Mg medium, in a preparation belonging to a rat from the (A+2LPS)_a subgroup.

Quantitative comparisons regarding the number of calcium pulses (of 12.5 μM each), taken up by mitochondria before the initiation of calcium release, are presented for the (A+0.5LPS) and (A+2LPS) subgroups in Tables 4 and 5, respectively. As can be seen, the differences between the typical and atypical subgroups are statistically significant or even very significant in all the media tested. On the other hand, despite apparent small differences, no significant statistical differences could be established between the typical subgroups themselves (or the atypical subgroups). Therefore, in further comparisons, the typical subgroups were pooled together in one group.

Table 4
Comparison between the typical and atypical A+0.5LPS groups

Statistical parameter	$\bar{x} \pm \text{SEM}$		t	$n' = n_1 + n_2 - 2$	p
Suspending medium \ Group	(A+0.5LPS) _t	(A+0.5LPS) _a			
MS	3.750 ± 0.250	5.500 ± 0.500	3.615	4	< 0.05
MSK	2.313 ± 0.237	4.250 ± 0.250	5.012	4	< 0.01
MSK-Mg	2.000 ± 0.000	3.000 ± 0.000	$\rightarrow \infty$	3	$\rightarrow 0$
KSW	1.750 ± 0.250	3.000 ± 0.000	3.333	4	< 0.05

Table 5
Comparison between the typical and atypical A+2LPS groups

Statistical parameter	$\bar{x} \pm SEM$		t	$n' = n_1 + n_2 - 2$	p
Suspending medium \ Group	(A+2LPS) _t	(A+2LPS) _a			
MS	3.833 ± 0.183	7.500 ± 0.500	5.389	3	< 0.02
MSK	2.833 ± 0.167	5.000 ± 0.000	1.073	3	< 0.01
MSK-Mg	2.000 ± 0.000	3.250 ± 0.354	4.743	3	< 0.02
KSW	2.000 ± 0.000	3.000 ± 0.000	$\rightarrow \infty$	3	$\rightarrow 0$

In order to establish in a more precise way the contribution of each of the two stress factors used in this experiment (chronic ethanol feeding and acute LPS injection), we compared the typical A+LPS group in turn with the typical alcoholic group (A_{1-6}) and with the (non-alcoholic) LPS-injected group. The results are presented in Tables 6 and 7, respectively. As can be seen from Table 6, with one exception (the KSW medium), there are no significant statistical differences between the typical A+LPS group and the typical alcoholic group. On the other hand, also with one exception (the MSK-Mg medium), there are very significant differences between the typical A+LPS group and the one treated with LPS only, as shown in Table 7.

Table 6
Comparison of the typical A+LPS group with the alcohol-fed group

Statistical parameter	$\bar{x} \pm SEM$		t	$n' = n_1 + n_2 - 2$	p
Suspending medium \ Group	A_{1-6}	(A+LPS) _t			
MS	4.750 ± 0.750	3.786 ± 0.214	1.820	7	> 0.1
MSK	2.333 ± 0.105	2.536 ± 0.176	0.946	11	> 0.1
MSK-Mg	2.000 ± 0.129	2.000 ± 0.000	$\rightarrow 0$	10	>>> 0.1
KSW	1.000 ± 0.000	1.857 ± 0.091	3.794	8	< 0.01

Table 7
Comparison of the typical A+LPS group with the LPS-treated group

Statistical parameter	$\bar{x} \pm SEM$		t	$n' = n_1 + n_2 - 2$	p
Suspending medium \ Group	LPS	(A+LPS) _t			
MS	5.250 ± 0.339	3.786 ± 0.214	4.965	11	< 0.001
MSK	3.125 ± 0.190	2.536 ± 0.176	3.134	19	< 0.01
MSK-Mg	2.300 ± 0.200	2.000 ± 0.000	1.662	9	> 0.1
KSW	2.083 ± 0.091	1.857 ± 0.091	2.274	11	< 0.02

This observation tells us that of the two stress factors used in our experiments, the chronic ethanol feeding is by far the most important one. This does not mean that the effect of LPS is not visible. In fact, the swelling amplitude (in KSW medium) of the mitochondria obtained from double-treated rats is significantly smaller at almost all times and for all calcium loads, whereas in the case of either (simply) alcoholic or LPS-injected animals this is true only for a few selected combinations of incubation time-calcium load (see Table 3 in [46] and the present work).

General discussion. In accord with previous results obtained with rats kept under tightly controlled conditions and fed a liquid diet [12, 43], we find that the acute LPS administration in small doses also affects the membrane permeability and related parameters of liver mitochondria from animals raised under the present conditions and that it modulates the response to chronic ethanol feeding, by decreasing the membrane permeability, even though the effect is not so obvious as before. In fact, it is very likely that the chronic ethanol feeding for such a long time (4 months) is associated with a constant penetration of LPS through the intestinal wall of the rats, as shown in the literature for similar circumstances [3, 32, 37]. Under these conditions, it is expected that the effect of alcohol should be modulated by the endogenous LPS and that the rats have already gained a certain tolerance to endotoxin. This might explain also why the LPS dose-dependent effects reported previously [43] are not so obvious in the present experiments. It must be stressed, however, that in our previous experiment we monitored only the matrix swelling in KSW medium, while the present work deals with more parameters and several different incubating media, for which we do not have a comparing possibility. Nevertheless, there is one study that can be partly corroborated with ours, showing an increased mitochondrial respiratory capacity after endotoxin pretreatment *in vivo* [12] and several other reports demonstrating protection against oxygen toxicity and an increased synthesis of superoxide dismutase by tracheal insufflation of endotoxin or other similar treatments [9, 11, 29, 40]. These observations and others lend support to the idea that LPS partly counteracts the effects of ethanol by controlling the production of the so-called reactive oxygen species (ROS), which have been demonstrated to be involved in ALD [1, 15, 16, 20, 47].

The problem with our alcohol-fed rats, however, seems to be complicated by the presence of animals with (what we call) atypical response, *i.e.*, rats with very low sensitivity to alcohol, that can be observed even after the LPS treatment of ethanol-fed animals. We found that about 30% of the rats can be included in this category. Of course, considering our model of ethanol administration, one could suspect that the rats insensitive to alcohol (designated as atypical) were in fact rats that somehow managed to avoid taking their share of ethanol. This can be ruled out, however, by the fact that atypical animals were also discovered among the double-treated (alcohol-fed, LPS-injected) animals, whereas no such animals were identified among the non-alcoholic (control) rats injected with LPS. If the effect is due to alcohol avoidance, the mitochondria of such rats should have a similar behaviour to

that of the LPS-injected control. However, this is generally not the case, since the mitochondria of atypical double-treated rats show less sensitivity to metabolic stress than those of the LPS-injected rats.

Another problem that should be addressed is the origin and explanation of the partial dissociation between the initiation of the matrix swelling and the beginning of $\Delta\Psi$ collapse in mitochondria of either alcohol-fed or LPS-injected rats, but especially in mitochondria of double-treated rats. Is this real or is it an artefact? And, if it is real, what is its significance? Since this dissociation phenomenon is not so obvious for the calcium release, which normally should accompany $\Delta\Psi$ collapse, an artefact seems a likely candidate. However, if one understands the response mechanism of the potential-sensitive probe, alternative explanations can be put forward. The mechanism of response of the cyanine dye used in the present experiments was unravelled many years ago [41, 42]. The most important feature of the dye is its moderate hydrophobicity, which makes it membrane permeant, but at the same time membrane bound. When a membrane potential difference is built up, the cationic dye, which is initially symmetrically distributed, simply concentrates on the negative face of the membrane. This surface concentration change leads to a concentration-dependent dimerisation associated with a change in the absorption spectrum. We assume that our preparation is a heterogeneous population of mitochondria, in which some organelles are more sensitive than others to different treatments (ethanol and/or LPS), either by their different origin and/or by their different position in the liver. If this is true, then some of the mitochondria will start depolarising and releasing calcium sooner than others. Since the probe is in a limited amount (2.5 μM), the dye of the inert mitochondria (those that have suffered the permeability transition) can now redistribute among the still active mitochondria, which can concentrate and dimerise or polymerise it, thus keeping the absorbance relatively unchanged until the majority of mitochondria suffer depolarisation. The calcium probe, on the other hand, is in much higher concentration (30 μM) and, being much more polar, it also has a different mechanism of response. Therefore, the release of calcium becomes apparent sooner after the triggering of the matrix swelling. In our opinion, the apparent dissociation of the membrane potential collapse from the other two phenomena studied, is a strong indication that mitochondria from different liver cells and/or different places have a differential sensitivity to apoptotic/necrotic stress factors (in our case ethanol and LPS). We hope that our electron microscopy samples, at least to some extent, will confirm our supposition. However, in order to have a clear-cut response, it is necessary to prepare mitochondria from different liver cells (parenchymal, Kupffer, stellate cells etc.) and find more subtle ways to monitor the activity of these cells *in situ*.

Conclusions. From the results obtained in the present work and partly from those reported in our accompanying paper [46], we can draw four groups of conclusions.

1. For all the experimental groups of rats (hence, for all the treatments), from the three membrane-associated phenomena studied, the one that is triggered first is the matrix swelling. Calcium release and $\Delta\Psi$ collapse occur more or less in parallel, as a consequence of the permeability transition and not vice-versa.

2. In general, in all media used for incubation, the chronic alcohol-feeding increases the sensitivity of mitochondria to acute stress factors (such as LPS and calcium), although the reactivity of the animals vary, some of them displaying alcohol tolerance.

3. The acute administration of bacterial endotoxin (LPS) to control rats, in sublethal doses, has an effect similar to, but less grave than, that of the chronic ethanol-feeding. A certain degree of dissociation of $\Delta\Psi$ collapse from the other two phenomena studied can be observed.

4. LPS administration to alcoholic rats leads to an intensification of the dissociation phenomenon mentioned above, a fact which may point to a differential sensitivity of different types of hepatic mitochondria. At the same time, a decrease in mitochondrial membrane permeability can be observed in the swelling medium, when moderate concentrations of calcium are used, suggesting that LPS may contribute to the phenomenon of alcohol tolerance.

R E F E R E N C E S

1. Adachi, M., Ishii, H., *Role of mitochondria in alcoholic liver injury*, "Free Radic. Biol. Med.", **32**, 2002, 487-491.
2. Bernardi, P., Petronilli, V., Di Lisa, F., Forte, M., *A mitochondrial perspective on cell death*, "Trends Biochem. Sci.", **26**, 2001, 112-117.
3. Bode, C., Kugler, V., Bode, J.C., *Endotoxemia in patients with alcoholic and non-alcoholic cirrhosis and in subjects with no evidence of chronic liver disease following acute ethanol excess*, "J. Hepatol.", **4**, 1987, 8-14.
4. Crompton, M., *The mitochondrial permeability transition pore and its role in cell death*, "Biochem. J.", **341**, 1999, 233-249.
5. Deaciuc, I.V., D'Souza, N.B., De Villiers, W.J.S., Burikhanov, R., Sraphie, T.G., Hill, D.B., McClain, C.J., *Inhibition of caspases in vivo protects the rat liver against alcohol-induced sensitization to bacterial lipopolysaccharide*, "Alcoholism: Clin. Exp. Res.", **25**, 2001, 935-943.
6. Deaciuc, I.V., D'Souza, N.B., Spitzer, J.J., *Tumor necrosis factor- α cell surface receptors of liver parenchymal and nonparenchymal cells during acute and chronic alcohol administration to rats*, "Alcoholism: Clin. Exp. Res.", **19**, 1995, 332-338.
7. Deaciuc, I.V., Fortunato, F., D'Souza, N.B., Hill, D.B., McClain, C.J., *Chronic ethanol exposure of rats exacerbates apoptosis in hepatocytes and sinusoidal endothelial cells*, "Hepatol. Res.", **19**, 2001, 306-324.
8. Deaciuc, I.V., Fortunato, F., D'Souza, N.B., Hill, D.B., Schmidt, J., Lee, E.Y., McClain, C.J., *Modulation of caspase-3 activity and Fas ligand mRNA expression in rat liver cells in vivo by alcohol and lipopolysaccharide*, "Alcoholism: Clin. Exp. Res.", **23**, 1999, 349-356.
9. Frank, J., Iqbal, J., Haas, M., Massaro, D., *New "rest period" protocol for inducing tolerance to high O₂ exposure in adult rats*, "Am. J. Physiol.", **257**, 1989, L226-L231.

10. Galle, P.R., *Apoptosis in liver disease*, "J. Hepatol.", **27**, 1997, 405-412.
11. Guidot, D.M., *Endotoxin treatment increases lung mitochondrial scavenging of extramitochondrial superoxide in hyperoxia-exposed rats*, "Arch. Biochem. Biophys.", **326**, 1996, 266-270.
12. Guidot, D.M., *Endotoxin pre-treatment in vivo increases the mitochondrial respiratory capacity in rat hepatocytes*, "Arch. Biochem. Biophys.", **354**, 1998, 9-17.
13. Hansen, J., Cherwitz, D.L., Allen, J.I., *The role of tumor necrosis factor-alpha in acute endotoxin-induced hepatotoxicity in ethanol-fed rats*, "Hepatology", **20**, 1994, 461-474.
14. Higuchi, H., Adachi, M., Miura, S., Gores, J.G., Ishii, H., *The mitochondrial permeability transition contributes to acute ethanol-induced apoptosis in rat hepatocytes*, "Hepatology", **34**, 2001, 320-328.
15. Hoek, J.B., *Mitochondrial energy metabolism in chronic alcoholism*, "Current Topics Bioenerg.", **17**, 1994, 197-241.
16. Ishii, H., Adachi, M., Fernandez-Checa, J.C., Cederbaum, A.I., Deaciuc, I.V., Nanji, A.A., *Role of apoptosis in alcoholic liver injury*, "Alcoholism: Clin. Exp. Res.", **27**, 2003, 1207-1212.
17. Johnson, D., Lardy, H., *Isolation of liver or kidney mitochondria*, in Colowick, S.P., Kaplan, N.O. (Eds.), *Methods in Enzymology*, Vol.10, pp.94-96, Acad. Press, New York, 1967.
18. Kroemer, G., Dallaporta, B., Resche-Rigon, M., *The mitochondrial death/life regulator in apoptosis and necrosis*, "Annu. Rev. Physiol.", **60**, 1998, 619-642.
19. Kroemer, Zamzami, N., Susin, S.A., *Mitochondrial control of apoptosis*, "Immunol. Today", **18**, 1997, 44-51.
20. Kurose, I., Higuchi, H., Miura, S., Saito, H., Watanabe, N., Hokari, R., Hirokawa, M., Takashi, M., Zeki, S., Nakamura, T., Ebinuma, H., Kato, S., Ishii, H., *Oxidative stress mediated apoptosis of hepatocytes exposed to acute ethanol intoxication*, "Hepatology", **25**, 1997, 368-378.
21. Leist, M., Single, B., Castoldi, A.F., Kuhnle, S., Nicotera, P., *Intracellular adenosine triphosphate (ATP) concentration: a switch between apoptosis and necrosis*, "J. Exp. Med.", **185**, 1997, 1481-1486.
22. Lemasters, J.J.V., *Necrapoptosis and the mitochondrial permeability transition: shared pathways to necrosis and apoptosis*, "Am. J. Physiol.", **276**, 1999, G1-G6.
23. Manna, S.K., Aggarwal, B.B., *Lipopolysaccharide inhibits TNF-induced apoptosis: role of nuclear factor- κ B activation and reactive oxygen intermediates*, "J. Immunol.", **162**, 1999, 1510-1518.
24. Marchetti, P., Castedo, M., Susin, S.A., Zamzami, N., Hirsch, T., Macho, A., Haeffner, A., Hirsh, F., Geuskens, M., Kroemer, G., *Mitochondrial permeability transition is a central coordinating event of apoptosis*, "J. Exp. Med.", **184**, 1996, 1155-1160.
25. McClain, C.J., Barve, S., Deaciuc, I., Kugelmas, M., Hill, D., *Cytokines in alcoholic liver disease*, "Semin. Liv. Dis.", **19**, 1999, 205-219.
26. Nanji, A.A., *Apoptosis and alcoholic liver disease*, "Semin. Liv. Dis.", **18**, 1998, 186-189.
27. Nanji, A.A., Ketthry, U., Sadrzadek, S.M.H., Yamanaka, T., *Severity of liver injury in experimental alcoholic liver disease. Correlation with plasma endotoxin, prostaglandin E₂, leukotriene B₄, and thromboxane B₂*, "Am. J. Pathol.", **142**, 1993, 367-373.

28. Nelson, S., Bagby, G., Summer, W., *Alcohol suppresses lipopolysaccharide-induced tumor necrosis factor activity in serum and lung*, "Life Sci.", **44**, 1989, 673-676.
29. Nethery, D., DiMarco, A., Stofan, D., Supinski, G., *Sepsis increases contraction-related generation of reactive oxygen species in the diaphragm*, "J. Appl. Physiol.", **87**, 1999, 1279-1286.
30. Neuman, M.G., Shear, N.H., Cameron, R.G., Katz, G., Tiribelli, C., *Ethanol-induced apoptosis in vitro*, "Clin. Biochem.", **32**, 1999, 547-555.
31. Nicotera, P., Leist, M., *Energy supply and the shape of death in rat lymphoid cells*, "Cell Death Differ.", **4**, 1997, 435-442.
32. Nolan, J.P., *Intestinal endotoxins as mediators of hepatic injury – an idea whose time has come again*, "Hepatology", **10**, 1989, 887-891.
33. Pastorino, J.G., Hoek, J.B., *Ethanol potentiates tumor necrosis factor- α cytotoxicity in Hepatoma cells and primary rat hepatocytes by promoting the induction of the mitochondrial permeability transition*, "Hepatology", **31**, 1999, 1141-1152.
34. Pastorino, J.G., Marcineviciute, A., Cahill, A., Hoek, J.B., *Potentiation by chronic ethanol treatment of the mitochondrial permeability transition*, "Biochem. Biophys. Res. Commun.", **265**, 1999, 405-409.
35. Petronilli, V., Penzo, D., Scorrano, L., Bernardi, P., Di Lisa, F., *The mitochondrial permeability transition, release of cytochrome c and cell death. Correlation with the duration of pore opening*, "J. Biol. Chem.", **276**, 2001, 12030-12034.
36. Scarlett, J.L., Murphy, M.P., *Release of apoptogenic proteins from mitochondrial intermembrane space during the mitochondrial permeability transition*, "FEBS Lett.", **418**, 1997, 282-286.
37. Schafer, C., Parlesak, A., Bode, J.C., Bode, C., *Endotoxin plasma levels and endotoxin-neutralizing factors in patients with alcoholic liver disease*, in Faist, E. (Ed.), *4th Int. Congr. on 'The Immune Consequences of Trauma, Shock and Sepsis'*, pp. 59-63, Monduzzi Editore, Bologna, 1997.
38. Suărășan, F., Tarba, C., *Corelații între fluxurile de calciu, tranziția de permeabilitate și colapsarea potențialului de membrană în mitocondriile hepatice de șobolan: II. Efectul tratamentului acut cu endotoxină (LPS)*, "An. Soc. Naț. Biol. Cel.", **8**, 2003, 413-429.
39. Susin, A.S., Zamzami, N., Castedo, M., Daugas, E., Wang, H.-G., Geley, S., Fassy, F., Reed, G.C., Kroemer, G., *The central executioner of apoptosis. Multiple links between protease activation and mitochondria in Fas/Apo-1/CD95- and ceramide induced apoptosis*, "J. Exp. Med.", **186**, 1997, 25-37.
40. Tang, G., Berg, J.T., White, J.E., Lumb, P.D., Lee, C.Y., Tsan, M.F., *Protection against oxygen toxicity by tracheal insufflation of endotoxin: role of MnSOD and alveolar macrophages*, "Am. J. Physiol.", **266**, 1994, L38-L45.
41. Tarba, C., *Use of an optical probe to estimate membrane potential at subcellular level*, "Rev. Roum. Biochim." **20**, 1983, 51-59.
42. Tarba, C., *Membrane potential in mitochondria estimated by absorption of a cyanine dye*, "Rev. Roum. Biochim." **20**, 1983, 61-65.
43. Tarba, C., Deaciuc, I.V., *Dose-dependent effects of bacterial endotoxin treatment on membrane permeability of ethanol-fed rats*, "Stud. Univ. Babeș-Bolyai, Biol.", **47** (2), 2002, 25-41.

44. Tarba, C., Suărăşan, F., *Corelaţii între fluxurile de calciu, tranzitia de permeabilitate şi colapsarea potenţialului de membrană în mitocondriile hepatice de şobolan: I .Efectul tratamentului cronic cu etanol*, “An. Soc. Naț. Biol. Cel.”, **8**, 2003, 391-412.
45. Tarba, C., Suărăşan, F., *Corelaţii între fluxurile de calciu, tranzitia de permeabilitate şi colapsarea potenţialului de membrană în mitocondriile hepatice de şobolan: III. Efectul combinat al alcoolizării cornice şi administrării acute de endotoxină (LPS)*, “An. Soc. Naț. Biol. Cel.”, **8**, 2003, 430-446.
46. Tarba, C., Suărăşan, F., *Functional correlations between the permeability transition, membrane potential collapse and calcium release by rat liver mitochondria of ethanol-fed rats*, “Stud. Univ. Babeş-Bolyai, Biol.”, **49** (1), 2004, 83-103.
47. Tsukamoto, H, Lu, S.C., *Current concepts in pathogenesis of alcoholic liver injury*, “FASEB J.”, **15**, 2001, 1335-1349.

UTILIZAREA INDICATORULUI ENZIMATIC AL CALITĂȚII NĂMOLULUI ÎN SCOPUL VALORIZĂRII ȘI PROTECȚIEI LACURILOR SALINE

MIHAIL DRĂGAN-BULARDA*, CĂTĂLIN EMIL GRIGORE**
și DANIEL ȚURA*

SUMMARY. – Utilisation of the Enzymatic Indicator of Mud Quality for Exploitation and Protection of Salt Lakes. We have determined six enzymatic activities (phosphatase, urease, protease, catalase, actual dehydrogenase, potential dehydrogenase) and three nonenzymatic catalytic activities (H_2O_2 splitting in autoclaved samples, TTC-[2,3,5-triphenyltetrazolium chloride] reduction in autoclaved samples, without or with glucose addition) in the muds of the salt lakes Techirghiol, Nuntași and Costinești, during four seasons (summer and autumn 2002, winter and spring 2003, respectively). The enzymatic and the nonenzymatic catalytic activities varied depending on the season, lake and the collection zone. The highest values were obtained in the mud of the Techirghiol lake, being followed by the Costinești and the Nuntași lakes. Generally, no evident seasonal variations were registered.. Based on the enzymatic indicator of mud quality (EIQM), we could determine the quality of the mud in the three analysed lakes. The mud of the Techirghiol lake is a very active one, the EIQM values being between 0.5 and 0.59. The mud of the Costinești lake had an EIQM between 0.4 and 0.6, while the mud of the Nuntași lake had the EIQM of 0.3-0.45. Taking into consideration the biological stability of the mud through the variability coefficients, we were able to establish that the muds of the Nuntași and Costinești lakes indicate an evolution towards high quality muds. By means of the EIQM we could estimate the evolution of the mud in the three lakes, in comparison to the situation registered 20 years ago. In the case of Techirghiol lake the value of EIQM increased with 0.4 units, in the case of Costinești lake it increased with 0.2 units and in the Nuntași lake with 0.1 units. The EIQM may be used for determining the quality of the muds, in order to use them in balneotherapy or in detecting of their some negative modifications – having this way the possibility to protect the lakes.

Este bine cunoscut că, în sedimentele acvatice, procesele microbiene stau la baza mineralizării materiei organice, precum și la baza degradării poluanților organici.

* Universitatea “Babeş-Bolyai”, Catedra de Biologie Experimentală, 400084 Cluj-Napoca, România.

E-mail: draganb@bioge.ubbcluj.ro

** Inspectoratul Școlar Județean Ialomița, Slobozia, România

Cataliza enzimatică, sub controlul direct sau indirect al microorganismelor, mediază acest proces complex ce asigură desfășurarea ciclurilor biologice ale elementelor C, N, P și S.

Determinarea activităților enzimatice în sedimentele acvatice, de altfel ca și în sol, constituie un instrument de cercetare pentru a evalua diversitatea funcțională a microbiotei, cu alte cuvinte a proceselor biochimice din aceste medii naturale și pentru găsirea unor indicatori ai calității solului [11], respectiv ai calității sedimentelor acvatice [10].

Nămolul din lacurile saline este produsul proceselor microbiologice ce au loc într-un ecosistem acvatic în strânsă conexiune cu condițiile de mediu [15]. Potențialul enzimatic al nămolului reflectă în mod direct sau indirect activitatea microbiotei, influența diferenților factori fizici, chimici, antropogeni asupra microbiotei, precum și asupra procesului de acumulare a enzimelor și chiar a intensității activităților enzimatice. Funcționarea unui ecosistem nu poate fi înțeleasă fără participarea activă a proceselor enzimatice [11].

Primele cercetări care au abordat determinarea activității enzimatice a unui sediment acvatic au fost cele realizate de Léonhardt [8, 9] la sedimentele unor râuri din Africa de Sud, ca mai târziu cercetările să se extindă și la sedimentele unor lacuri saline, apreciindu-se potențialul enzimatic al nămolului printr-un indicator enzimatic al calității obținut pe baza determinării mai multor activități enzimatice și catalitice neenzimatice [1-3, 5-7, 10, 12-14]. Prima clasificare a lacurilor saline din țara noastră pe baza activităților enzimatice și catalitice neenzimatice ale nămolului terapeutic a fost publicată în 1986 de către Küss et al. [6], fiind comparate valorile indicatorului enzimatic al calității nămolului din 37 lacuri saline, iar apoi s-a efectuat o a doua clasificare a 56 lacuri saline de către Muntean și colab. în 1996 [10], datele obținute ulterior fiind comparate cu acest din urmă sistem de clasificare a lacurilor saline.

In lucrarea de față ne-am propus să urmărim valorile indicatorului enzimatic al calității nămolului la trei lacuri saline din zona litoralului românesc, determinate în anii 2002 și 2003, în vederea aprecierii calității nămolului, ca factor de bază în valorificarea acestuia în balneoterapie, respectiv în vederea stabilirii unor măsuri de protecție a lacurilor în cazul când apar influențe antropogene nedorite.

Materiale și metode. Au fost studiate sedimentele a 3 lacuri saline din zona litoralului românesc și anume Techirghiol, Nuntași și Costinești. Prelevarea probelor de nămol s-a realizat în vara și toamna anului 2002, respectiv în iarna și primăvara anului 2003. În cazul lacului Techirghiol, probele de nămol s-au recoltat din 3 puncte (Debarcader, Coada lacului, Băile reci), din lacul Nuntași colectarea s-a realizat din 4 puncte (Camping, Istria 1, Istria 2, E-lacului), iar din lacul Costinești din 3 puncte (Debarcader, Zona Obelisc, Camping).

Au fost determinate următoarele 9 activități enzimatice și catalitice neenzimatice: fosfatazică, ureazică, proteazică, catalazică, catalitică neenzimatică, reducerea TTC în probe de nămol neautoclavate, fără glucoză (activitate dehidrogenazică

actuală), respectiv cu glucoză (activitate dehidrogenazică potențială), reducerea TTC în probe de nămol autoclavate, fără glucoză, respectiv cu adaos de glucoză. În vederea determinării activităților catalitice neenzimatiche, părți din probele de nămol s-au autoclavat la 121°C , timp de o oră, în trei zile consecutive.

Activitatea fosfatazică s-a determinat prin metoda lui Kramér și Erdélyi [4], în amestecuri de reacție ce au constat din 2,5 g sediment + 2 mltoluen (ca agent antiseptic) + 10 ml substrat enzimatic (fenilfosfat disodic, soluție 0,5 %) sau 10 ml apă distilată. Activitatea ureazică s-a determinat prin metoda Sunmér [4], în amestecuri de reacție constând din 2,5 g sediment + 2 mltoluen + 5 ml soluție tampon fosfat (pH 6,7) + 5 ml soluție de uree 3%. Activitatea proteazică s-a determinat prin metoda Beck [4], în amestecuri de reacție conținând 2,5 g sediment + 2 mltoluen + 20 ml soluție de gelatină 2%. O tehnică bazată pe metoda lui Kapppelen [4] a fost aplicată pentru determinarea activității catalazice, în amestecuri de reacție cu 1,5 g sediment + 2 ml H_2O_2 3% + 10 ml apă distilată. Aceeași tehnică a fost utilizată și pentru determinarea activității catalitice neenzimatiche, dar probele de nămol au fost inactivate termic prin autoclavare. Reducerea TTC în probe de nămol neautoclavate (activitatea dehidrogenazică) și în probe de nămol autoclavate s-a determinat conform metodei lui Casida și colab. [4]. Compoziția amestecurilor de reacție a fost următoarea: 0,5 g nămol + 1 ml apă distilată sau 1 ml soluție de glucoză 3% + 0,5 ml soluție de clorură de 2,3,5-trifeniltetrazoliu (TTC) sau 0,5 ml apă distilată.

Activitățile au fost exprimate astfel: activitatea fosfatazică în mg fenol/ 2,5 g sediment uscat/ 24 ore, la 37°C ; activitatea ureazică în mg NH_4^+ / 2,5 g sediment uscat/24 ore, la 37°C ; activitatea proteazică în mg $\text{NH}_2\text{-N}/2,5$ g sediment uscat/24 ore, la 37°C ; activitatea catalazică și catalitică neenzimatică în mg $\text{H}_2\text{O}_2/1,5$ g sediment uscat/1 oră, la 20°C ; activitățile de reducere a TTC în mg trifeniiformazan/0,5 g sediment uscat/24 ore, la 37°C . Pentru fiecare activitate enzimatică și catalitică neenzimatică a fost calculat coeficientul de variație [4].

Rezultate și discuții. Rezultatele obținute sunt prezentate în Tabelele 1-5, respectiv în Fig. 1-5.

Tabelul 1 prezintă datele privind activitățile enzimatiche și catalitice neenzimatiche ale probelor de nămol colectate în vara anului 2002 din cele 3 lacuri litorale studiate (Techirghiol, Nuntași și Costinești). Se poate constata, din analiza rezultatelor, că au fost evidențiate toate cele 6 activități enzimatiche (fosfatazică, ureazică, proteazică, catalazică, dehidrogenazică actuală, dehidrogenazică potențială) și cele 3 activități catalitice neenzimatiche (scindarea catalitică a H_2O_2 , reducerea neenzimatică a TTC în probe fără glucoză, reducerea neenzimatică a TTC în probe cu glucoză).

Valorile activităților determinate au oscilat în funcție de lacul studiat, respectiv în funcție de locul de colectare. Lacul Techirghiol, unul din cei mai importanți producători de nămol sapropelic din Europa, se situează pe primul loc în privința potențialului enzimatic al nămolului, valorile activităților enzimatiche, în special, fiind cele mai ridicate, comparativ cu cele determinate în celelalte două lacuri (Nuntași și Costinești). În privința locului de colectare, cele mai ridicate valori s-au înregistrat

în zona Băile reci, la toate cele 6 activități enzimaticice, urmate de zonele Debarcader și Coada lacului. În ceea ce privește activitățile catalitice neenzimatice, nu se constată deosebiri marcante între cele 3 zone de colectare, valori ușor mai ridicate fiind obținute în zona Debarcader, respectiv în cea de la Coada lacului. Subliniem că lacul Techirghiol are un nămol negru onctuos, cu un grad ridicat de peloidizare, care posedă un potențial enzimatic ridicat.

Din analiza Tabelului 1 reiese, de asemenea, că după lacul Techirghiol, valorile activităților enzimaticice sunt destul de însemnate și în nămolul din lacul Costinești, dar mai reduse comparativ cu cele întâlnite la toate zonele lacului Techirghiol. Lacul Costinești, se cunoaște, are o salinitate mai scăzută, iar biocenozele sunt mai sărace, fapt ce se reflectă și în potențialul enzimatic. Nămolul este tot unul negru, dar cu un grad mai scăzut de onctuoitate. În cazul acestui lac, nămolul din zona Obelisc este cel mai activ pe plan enzimatic, prezentând valorile cele mai ridicate atât la activitățile enzimaticice cât și la cele neenzimatice.

Nămolul din lacul Nuntași prezintă cele mai scăzute valori ale activităților enzimaticice, chiar și ale activităților catalitice neenzimatice, cu excepția scindării H_2O_2 , unde la toate punctele de prelevare nămolul a prezentat valori mai ridicate decât probele colectate din celelalte două lacuri. Între cele 4 zone de colectare, există deosebiri în funcție de activitatea enzimatică determinată, dar nămolul din punctele Istria 1 și Istria 2 s-a dovedit a fi cel mai activ pe plan enzimatic. Subliniem că lacul Nuntași are un nămol cenușiu, ușor onctuos sau lipsit de onctuoitate, cu un grad mai mic de peloidizare, care s-a demonstrat a avea și un potențial enzimatic mai scăzut. Lacul Nuntași prezintă o salinitate mai mică, probabil și biocenozele sunt de alt tip, ceea ce se răsfrângă și asupra procesului de peloidogeneză, respectiv asupra potențialului biologic (enzimatic) al nămolului.

In Fig. 1 se prezintă evoluția potențialului enzimatic al nămolului din cele 3 lacuri litorale, în perioada de vară, din toate punctele de colectare, pe baza indicatorului enzimatic al calității nămolului (IECN). Se poate constata mai pregnant diferențierea între cele 3 lacuri litorale, precum și între punctele de colectare de la fiecare lac în parte. Se confirmă faptul că pe baza acestui indicator global al activității enzimaticice, lacurile Techirghiol și Costinești se situează în categoria lacurilor cu sedimente active din punct de vedere enzimatic, având valori ale IECN cuprinse între 0,5 și 1, pe când lacul Nuntași are nămol de calitate mai redusă din punct de vedere enzimatic, valorile IECN fiind cuprinse între 0,3 și 0,5. De asemenea, este evidentă diferențierea și între punctele de colectare, fiecare lac prezentând o zonă cu nămol mai activ din punctul de vedere al potențialului enzimatic.

După cum se poate vedea, în cazul lacului Techirghiol, diferențierile între cele 3 zone sunt mai pregnante, în sensul că zona de la Coada lacului se distanțează la o diferență de 0,2 unități, de zona Băile reci, pe când la celelalte două lacuri nămolul este mai omogen, diferența între punctele de colectare fiind doar de circa 0,1 unități.

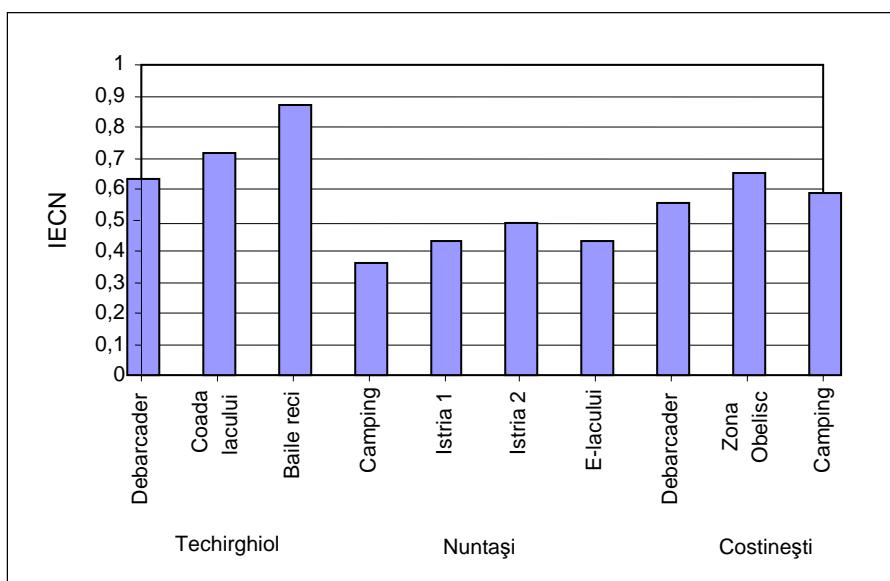
INDICATORUL ENZIMATIC AL NĂMOLULUI DIN LACURI SALINE

Tabel 1

Valorile activităților enzimatici și catalitice neenzimatice ale probelor de nămol prelevate vara din lacurile litorale studiate

Lacul	Locul de prelevare	Activitatea				Activitatea catalitică	Reducerea TTC				IECN*
		fosfatazică	ureazică	proteazică	Enzi-matică		Neenzi-matică	Nămol neautoclavat	Nămol autoclavat		
		Fără glucoză	Cu glucoză	Fără glucoză	Cu glucoză		Fără glucoză	Cu glucoză			
Techirghiol	Debarcader	5,925	1,450	1,925	41,560	24,680	4,500	5,825	0,920	0,985	0,635
	Coadă lacului	4,116	1,160	2,524	48,440	22,450	4,025	4,940	1,100	1,200	0,715
	Băile reci	7,220	2,350	3,286	46,200	21,600	5,116	6,200	1,090	1,195	0,868
Nunăști	Camping	0,900	0,490	0,526	27,222	23,670	2,160	2,905	0,526	0,520	0,364
	Istria 1	1,270	0,600	0,724	28,324	28,953	2,612	3,100	0,620	0,700	0,435
	Istria 2	1,925	0,702	0,803	29,966	29,440	2,928	3,625	0,725	0,800	0,488
	E-lacului	0,830	0,395	0,640	31,445	30,166	1,900	2,400	0,832	0,790	0,431
Costinești	Debarcader	3,250	0,936	0,983	32,160	23,100	3,100	4,100	0,940	1,030	0,555
	Zona Obelisc	4,500	1,025	1,124	39,260	24,525	3,900	4,600	1,095	1,200	0,652
	Camping	3,800	0,986	1,090	35,166	23,900	3,125	4,250	0,990	1,050	0,586

*IECN - Indicatorul enzimatic al calității nămolului.



F i g . 1. Indicatorul enzimatic al calității nămolului – vara.

In Tabelul 2 sunt trecute rezultatele obținute la determinările activităților enzimaticice și catalitice neenzimaticice la probele de nămol colectate în toamna anului 2002. Se poate observa că există deosebiri, dar nu de ordin global, ci mai mult oscilații ale valorilor obținute la diferite activități enzimaticice. Se menține poziția lacurilor din punctul de vedere al potențialului enzimatic, cu specificarea că probele din lacul Techirghiol prezintă chiar activități ușor mai ridicate, pe când nămolorile din lacurile Nuntași și Costinești au furnizat valori cu ceva mai scăzute decât în perioada de vară, ceea ce se poate remarcă și pe baza analizei Fig. 2. Se constată că se menține și poziția locurilor de colectare, cu o singură excepție, în cazul lacului Techirghiol, unde s-au înregistrat valori mai crescute la nămolul din zona Coada lacului, acestea fiind foarte apropiate de valorile obținute la cel din zona Debarcader. Se poate afirma că în mediul acvatic nu se semnalează variații sezoniere decât de mică intensitate, la toate cele 3 lacuri litorale luate în studiu.

Tabel 2

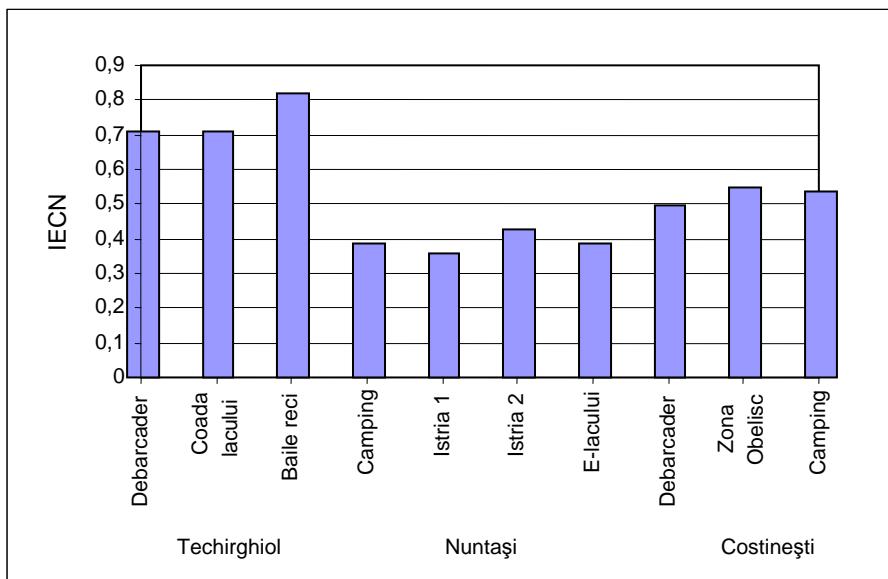
Valorile activităților enzimaticice și catalitice neenzimaticice ale probelor de nămol prelevate toamna din lacurile litorale studiate

Lacul	Locul de prelevare	Activitatea fosfatazică			Activitatea ureazică			Activitatea proteazică			Activitatea catalitică			Reducerea TTC		
		Enzi-	Neenzi-	Nămol	Nămol	IECN*										
						Fără	Cu	glucoză	Fără	Cu	glucoză	glucoză				
Techirghiol	Debarcader	6,240	1,829	2,146	46,606	21,096	6,379	9,200		1,725	1,650		0,707			
	Coadă lacului	4,600	1,700	2,788	45,305	20,366	5,925	8,311		1,402	1,495		0,708			
	Băile reci	8,200	2,688	3,929	43,200	23,725	6,800	9,025		1,200	1,290		0,822			
Nuntași	Camping	1,320	0,700	0,680	29,708	26,600	2,400	3,100	0,400	0,525	0,525		0,388			
	Istria 1	1,425	0,925	0,932	28,609	24,425	2,800	3,600	0,325	0,613	0,613		0,359			
	Istria 2	1,640	0,990	0,998	31,405	28,306	3,100	4,000	0,788	0,780	0,780		0,428			
	E-lacului	1,304	0,650	0,720	29,983	23,405	2,900	3,300	0,902	0,844	0,844		0,387			
Costinești	Debarcader	3,996	1,126	1,090	34,507	24,705	3,820	4,222	0,980	1,010	1,010		0,494			
	Zona Obelisc	5,200	1,300	1,485	25,900	25,900	4,129	4,900	0,991	1,115	1,115		0,546			
	Camping	4,326	1,250	1,325	26,100	26,100	4,016	4,725	0,940	0,975	0,975		0,537			

*IECN - Indicatorul enzimatic al calității nămolului.

Rezultatele obținute la determinările efectuate asupra probelor de nămol colectate în iarna anului 2003 sunt cuprinse în Tabelul 3, respectiv în Fig. 3. Se poate observa că, practic, oscilațiile valorilor activităților enzimaticice și ale activităților catalitice neenzimaticice sunt foarte restrâns, menținându-se constatarea că influența sezonieră este mică în cazul lacului Techirghiol și foarte mică în cazul lacurilor Nuntași și Costinești (lacuri cu adâncimi mai mici, comparativ cu lacul Techirghiol). Este de semnalat că la nămolorile din lacurile Nuntași și Costinești chiar s-au obținut valori ale IECN mai ridicate în perioada de iarnă decât în perioada de toamnă, la toate punctele de colectare, ceea ce întărește afirmația că în sedimentele acvatice ale lacurilor saline se face mai puțin simțită influență sezonieră, datorată în primul rând factorilor climaterici.

INDICATORUL ENZIMATIC AL NĂMOLULUI DIN LACURI SALINE



F i g . 2. Indicatorul enzimatic al calității nămolului – toamna.

Tabel 3

Valorile activităților enzimaticice și catalitice neenzimatice ale probelor de nămol prelevate iarna din lacurile litorale studiate

Lacul	Locul de prelevare	Activitatea			Activitatea			Activitatea			Reducerea TTC			
		fosfatazică	ureazică	proteazică	catalitică	Enzi-	Neenzi-	Nămol	Nămol	Nămol	Nămol	IECN*		
								neautoclavat	autoclavat	Fără	Cu			
Techirghiol	Debarcader	3,200	0,995	1,100	38,006	25,506	2,125	2,800	0,560	0,610	0,583			
	Coada lacului	2,100	0,800	1,500	36,305	24,425	3,000	3,926	0,912	0,980	0,728			
	Băile reci	3,800	1,326	2,100	40,126	26,100	3,600	3,728	0,990	1,012	0,878			
Nuntași	Camping	0,840	0,330	0,310	22,448	26,516	1,740	2,305	0,415	0,418	0,393			
	Istria 1	0,990	0,460	0,450	25,556	27,412	1,925	2,604	0,306	0,350	0,420			
	Istria 2	1,100	0,520	0,560	28,312	29,500	2,010	2,925	0,525	0,600	0,503			
	E-lacului	0,930	0,280	0,380	26,405	31,040	1,500	2,040	0,280	0,300	0,389			
Costinești	Debarcader	2,115	0,720	0,700	30,802	24,940	2,117	3,116	0,726	0,800	0,590			
	Zona Obelisc	3,090	0,940	0,910	32,516	25,518	2,924	3,041	0,980	0,990	0,714			
	Camping	2,240	0,880	0,830	33,400	22,416	2,200	2,900	0,800	0,860	0,620			

*IECN - Indicatorul enzimatic al calității nămolului.

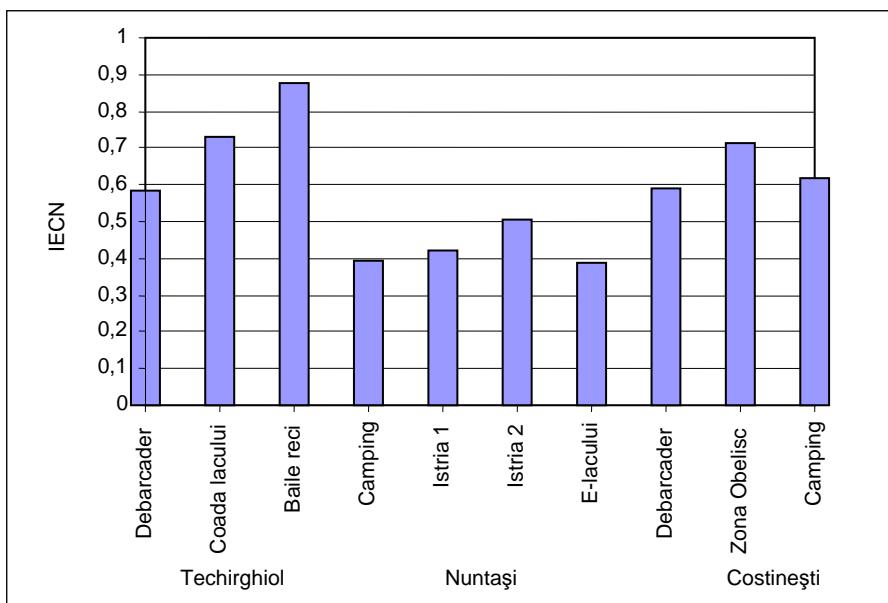


Fig. 3. Indicatorul enzimatic al calității nămolului – iarna.

Determinările efectuate în perioada de primăvară 2003, la nămolurile din cele 3 lacuri studiate, sunt trecute în Tabelul 4, respectiv în Fig. 4. Acestea au confirmat concluzia noastră că schimbările sezoniere în sedimentele celor 3 lacuri analizate sunt de mică importanță, comparativ cu sedimentele acvatice din apele curgătoare și în contrast cu cele semnalate de mulți cercetători în cazul solurilor din zonele temperate. Mai mult, la lacurile Nuntași și Costinești, chiar se înregistrează valori mai mici ale IECN în perioada de primăvară, comparativ cu cele calculate în celelalte 3 anotimpuri.

Tabel 4
Valorile activităților enzimatice și catalitice neenzimatice ale probelor de nămol prelevate primăvara din lacurile litorale studiate

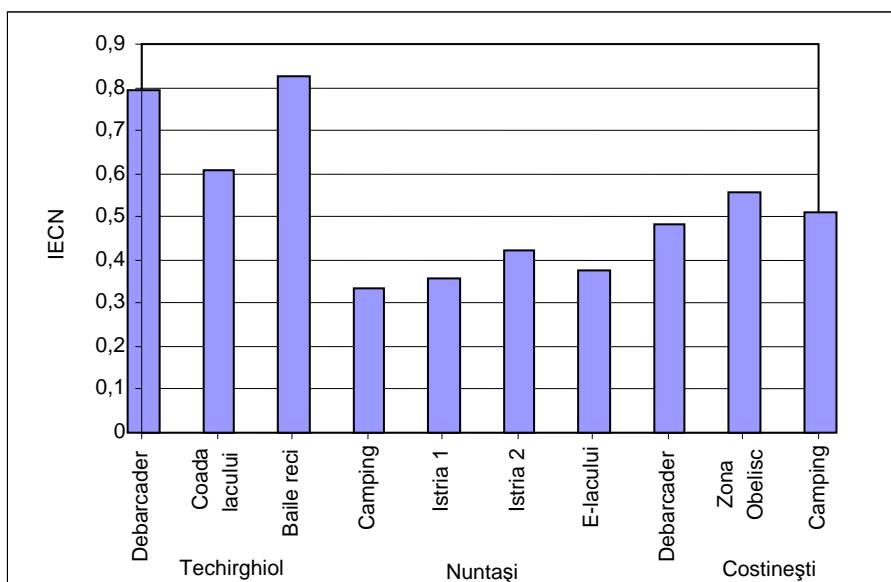
Lacul	Locul de prelevare	Activitatea	Activitatea	Activitatea	Activitatea	Reducerea TTC		IECN*			
		fosfatazică	ureazică	proteazică	catalitică	Nămol neautoclavat	Nămol autoclavat				
		Enzi- matică	Neenzi- matică	Nămol Fără glucoză	Nămol Cu glucoză	Fără glucoză	Cu glucoză				
Techirghiol	Debarcader	8,659	1,955	2,800	48,365	20,345	6,328	9,306	2,900	2,980	0,795
	Coada lacului	5,275	1,624	3,112	40,200	22,284	5,210	6,800	1,480	1,600	0,609
	Băile reci	7,935	2,985	3,560	46,121	24,386	7,251	10,245	2,100	2,300	0,824
Nuntași	Camping	2,112	0,825	0,805	30,900	25,300	2,700	3,400	0,500	0,504	0,335
	Istria 1	2,614	0,991	1,220	29,300	26,525	3,112	3,700	0,380	0,400	0,359
	Istria 2	2,912	1,200	1,405	31,225	27,627	3,345	4,114	0,804	0,860	0,420
	E-lacului	2,557	0,850	0,980	30,011	24,428	3,091	3,900	0,900	0,990	0,377

INDICATORUL ENZIMATIC AL NĂMOLULUI DIN LACURI SALINE

Tabel 4 (continued)

Lacul	Locul de prelevare	Activitatea	Activitatea	Activitatea	Activitatea	Reducerea TTC		IECN*			
		fosfatazică	ureazică	proteazică	catalitică	Enzi-	Neenzi-				
		matică	matică	Fără	Nămol	neautoclavat	Nămol				
Costinești	Debarcader	4,925	1,400	1,800	37,500	25,624	3,990	4,466	0,900	0,906	0,484
	Zona	6,129	1,890	1,925	42,829	26,120	4,500	5,290	1,050	1,012	0,555
	Obelisc										
	Camping	5,300	1,750	1,550	41,925	24,928	4,346	5,016	0,968	0,900	0,512

*IECN - Indicatorul enzimatic al calității nămolului.



F i g . 4. Indicatorul enzimatic al calității nămolului – primăvara.

Trebuie să subliniem că în determinările noastre, 4 dintre activitățile enzimatic se datorează enzimelor acumulate (fosfatază, urează, protează, catalază) care sunt în mai mică măsură influențate de factorii climaterici în mediul acvatic, comparativ cu activitățile enzimatic datorate microbiotei proliferante (dehidrogenaza actuală și cea potențială), care pot suferi oscilații în funcție de dezvoltarea microbiotei, dar fără a modifica semnificativ valoarea IECN. De asemenea, nici activitățile catalitice neenzimatic nu sunt influențate în mod vizibil de influențele sezoniere, ceea ce a dus la obținerea unor valori ale IECN destul de apropiate, indiferent de anotimpul în care s-au colectat probele de nămol pentru analiză (Fig. 5).

Datele din Tabelul 5 ne furnizează informații despre stabilitatea biologică a nămolorilor din cele 3 lacuri, pe baza coeficientilor de variabilitate calculați. Se înregistrează diferențe în privința valorilor coeficientilor, în funcție de activitatea

Tabel 5

Coeficienții de variabilitate sezonieră și anuală ai activităților enzimaticе și catalitice neenzimaticе în nămolul lacurilor litorale studiate (%)

Lacul	Sezonul	Activitatea	Activitatea	Activitatea	Activitatea	Reducerea TTC			
		fosfatazică	ureazică	proteazică	catalitică	Enzi-	Neenzi-	Nămol	Nămol
						matică	matică	neautoclavat	autoclavat
Techirghiol	Primăvara	24,400	32,400	11,700	7,600	9,000	16,300	20,000	32,900 30,000
	Vara	27,100	37,500	26,400	7,700	6,900	12,000	11,400	9,700 9,900
	Toamna	28,400	25,900	30,500	3,800	8,100	6,800	5,100	18,300 12,200
	Iarna	28,400	25,500	32,100	5,000	3,300	25,500	16,900	25,700 14,600
	Anual	24,075	30,320	18,580	6,020	6,500	15,150	13,350	21,650 16,450
Nunăști	Primăvara	12,900	17,700	30,200	2,700	2,300	8,700	6,100	38,000 40,400
	Vara	31,100	24,300	17,100	6,400	6,000	19,100	16,800	19,500 18,400
	Toamna	10,800	11,800	39,300	4,700	8,500	10,500	10,100	46,100 22,300
	Iarna	11,300	28,000	23,500	9,500	7,100	12,600	15,400	31,400 29,200
	Anual	16,520	20,450	27,520	5,820	5,970	12,720	12,100	33,750 27,570
Costinești	Primăvara	10,600	15,000	10,800	6,900	2,300	6,100	9,200	7,700 6,700
	Vara	16,200	4,500	6,900	7,100	2,900	13,400	5,900	7,700 7,500
	Toamna	13,800	7,300	15,200	9,500	2,900	3,900	7,700	2,700 4,200
	Iarna	19,000	13,400	13,600	4,000	6,700	18,300	3,600	10,900 15,000
	Anual	14,900	10,050	11,620	6,870	3,700	10,420	6,600	7,250 8,350

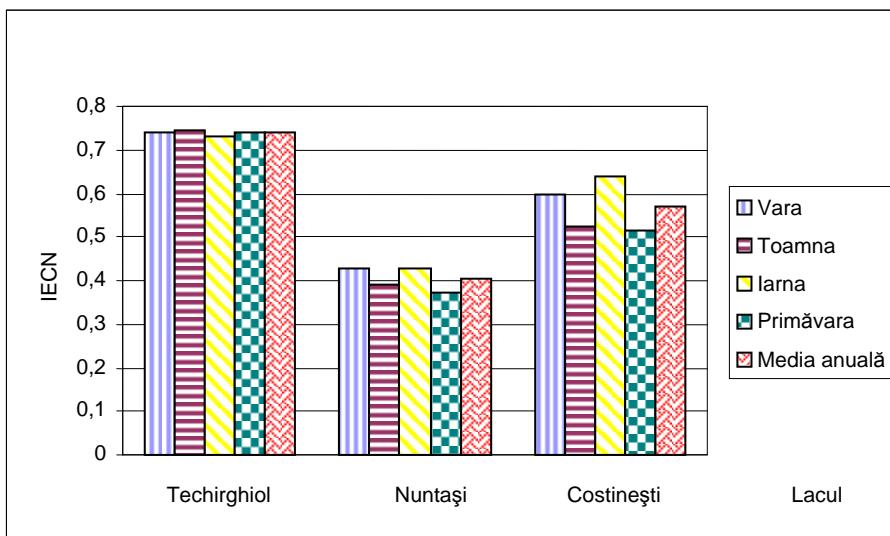


Fig. 5. Indicatorul enzimatic global.

enzimatică determinată, de data colectării probelor și de proveniența lor în funcție de lacul studiat. Valori mai ridicate ale acestora denotă stabilitate biologică mai mică și invers. Comparând lacurile între ele, din acest punct de vedere, putem constata că nămolul din lacul Costinești prezintă cea mai bună stabilitate biologică, urmat de nămolul din lacul Nuntași și apoi de cel din lacul Techirghiol. Aceste rezultate ne permit să afirmăm că nămolul din cele 3 lacuri prezintă o bună stabilitate biologică, ceea ce denotă evoluția acestora spre nămoluri cu potențial microbiologic (enzimatic) ridicat, intrând în categoria nămolurilor valoroase pentru balneoterapie. Determinarea activităților descrise mai sus, la perioade mai scurte de timp, vor putea sesiza modificările apărute în potențialul biologic al nămolului, impunându-se măsuri de protecție.

Concluzii. 1. In nămolurile colectate din lacurile litorale Techirghiol, Nuntași și Costinești au fost evidențiate 6 activități enzimaticice (fosfatazică, ureazică, proteazică, catalazică, dehidrogenazică actuală, dehidrogenazică potențială) și 3 activități catalitice neenzimaticice (scindarea catalitică a H_2O_2 , reducerea neenzimatică a TTC în probe fără glucoză, respectiv reducerea neenzimatică a TTC în probe cu glucoză), în toate cele 4 anotimpuri (vara și toamna anului 2002, respectiv iarna și primăvara anului 2003).

2. Activitățile enzimaticice și cele catalitice neenzimaticice au oscilat în funcție de anotimp, de lac și de zona de colectare. Cele mai ridicate valori ale activităților enzimaticice determinate s-au înregistrat în nămolul lacului Techirghiol, urmat de cel din lacul Costinești și apoi de cel din lacul Nuntași. La toate lacurile nu s-au semnalat variații sezoniere, decât de mică intensitate, ceea ce ne-a condus la constatarea că în mediul acvatic, în special în lacurile saline, influența factorilor climaterici asupra activității microbiologice a sedimentului este de mai mică importanță.

3. Pe baza indicatorului enzimatic al calității nămolului (IECN) s-a putut aprecia valoarea nămolului din cele 3 lacuri, pe primul loc situându-se cel din lacul Techirghiol, urmat de cel din lacul Costinești și apoi de cel din lacul Nuntași. Lacul Techirghiol dispune de un nămol foarte activ din punct de vedere enzimatic, cu un potențial microbiologic ridicat (IECN cuprins între 0,5 și 0,9), încadrat pe baza datelor obținute între cele mai active nămoluri terapeutice de la noi din țară. Cel din lacul Costinești se încadrează în categoria nămolurilor active (IECN cuprins între 0,4 și 0,6), iar cel din lacul Nuntași în categoria nămolurilor mai puțin active (IECN cuprins între 0,3 și 0,45).

4. Pe baza coeficienților de variabilitate calculați s-a putut aprecia stabilitatea biologică a sedimentului, care s-a dovedit a fi mai bună la nămolurile din lacurile Costinești și Nuntași, permitând evoluția acestora spre un nămol de calitate superioară.

5. Cu ajutorul IECN s-a putut aprecia evoluția nămolului din cele 3 lacuri, cel din lacul Techirghiol crescând cu 0,4 unități, față de perioada 1988, iar nămolul din celelalte lacuri, cu o rată mai mică. Pe baza valorilor IECN se poate monitoriza evoluția unui nămol terapeutic, luându-se măsuri de protecție, în anumite situații, respectiv se poate aprecia valoarea acestuia pentru utilizarea lui în balneoterapie.

REFERENCES

1. Bulgăreanu, V.-A.C., Kiss, S., Pașca, D., Drăgan - Bularda, M., Crișan, R., Pintea, H., Zborovschi, E., *The relations between the limnogeology and enzymology of sediments from some brackish and salt lakes*, "Rev. Roum. Géol., Géophys. Géogr.", Sér. Géol.", **31**, 1987, 107-114.
2. Bulgăreanu, V.- A., Kiss, S., Rădulescu, D., Drăgan - Bularda, M., Hanisch, D., *Influence of limnogeological conditions on the enzymatic and nonenzymatic catalytic activities in the sediments of salt lakes in Sovata*, "Stud. Univ. Babeș-Bolyai, Biol.", **26** (1), 1981, 68-72.
3. Crișan, R., Pașca, D., Muntean, V., Zborovschi, E., Drăgan - Bularda, M., Kiss, S., *Enzymological study of muds from salt lakes in Turda (Cluj district)*, "Stud. Univ. Babeș-Bolyai, Biol.", **33** (2), 1988, 76-86.
4. Drăgan - Bularda, M., *Lucrări practice de Microbiologie generală*, p. 67-68, 175-191, Univ. Babeș-Bolyai, Cluj-Napoca, 2000.
5. Drăgan - Bularda, M., Kiss, S., Rădulescu, D., Crișan, R., Pintea, H., Kolozsi, E., Bulgăreanu, V.- A., *Studierea enzimologică a nămolurilor din lacurile de la Ocna Sibiului*, "Lucr. celui de al 3-lea Simp. Microbiol. Ind." (București, 1981), 1982, 700-709.
6. Kiss, S., Drăgan - Bularda, M., Pașca, D., *Environmental enzymology*, "Stud. Univ. Babeș-Bolyai, Biol.", **31** (1), 1986, 3-13.
7. Kiss, S., Rădulescu, D., Drăgan - Bularda, M., Bulgăreanu, V.- A., Nicula, G., *Contributions to the enzymological study of therapeutic muds*, "Stud. Univ. Babeș-Bolyai, Biol.", **24** (2), 1979, 54-63.
8. Lenhard, G., du Plooy, A., *Studies on bottom deposits of the Vaal River System*, "Hydrobiologia", **26**, 1965, 271-291.
9. Lenhard, G., Ross, W.R., du Plooy, A., *A study of methods for the classification of bottom deposits of natural waters*, Hydrobiologia", **20**, 1962, 223-240.
10. Muntean, V., Crișan, R., Pașca, D., Kiss, S., Drăgan - Bularda, M., *Enzymological classification of salt lakes in Romania*, "Int. J. Salt Lake Res.", **5**, 1996, 35-44.
11. Nannipieri, P., Badalucco, L., *Biological processes*, în Bendix, D.K., Nieder, R. (Eds.), *Handbook of Processes and Modeling in the Soil-Plant System*, pp.57-82, Haworth Press, Binghamton, New York, 2002.
12. Pașca, D., Kiss, S., Pintea, H., *Cercetări enzimologice la nămolul din lacul Băile de la Cojocna (jud.Cluj)*, "Actualitate și perspectivă în biologie. Structuri și funcții în ecosisteme terestre și acvatice", 1985, 265-274, Centr. Cerc.Biol., Cluj-Napoca.
13. Pașca, D., Muntean, V., Crișan, R., Pintea, H., Zborovschi, E., Iliescu, M., Iliescu, C., Drăgan - Bularda, M., Kiss, S., *Evolution of the enzymatic potential of mud in the Techirghiol salt lake*, "Evol. Adapt." (Cluj-Napoca), **4**, 1991, 119-123.
14. Rădulescu, D., Kiss, S., Drăgan - Bularda, M., *Studii enzimologice asupra nămolului terapeutic de la Băile I Mai – Oradea*, "Stud. Univ. Babeș-Bolyai, Biol.", **15** (2), 1970, 145-149.
15. Tuculescu, I., *Dinamica lacului Techirghiol. Biocenozele și geneza nămolului*, Ed. Acad. Rom., București, 1965.

CONTRIBUȚII LA STUDIEREA DROJDILOR INDUSTRIALE

LETIȚIA OPREAN*

SUMMARY. - Contributions to the Study of Industrial Yeasts. Twenty yeast strains have recently been isolated in pure cultures from industrial sources and identified based mainly on the morphological properties of their cells and cultures on solid and liquid nutrient media. Majority of the strains (15) are alcohologenic belonging to the genus *Saccharomyces* and comprising two brewer's (beer) yeast strains (*S. carlsbergensis* = *S. uvarum* A and B), two baker's yeast strains (*S. cerevisiae* CA and CP), one spirit yeast strain (*S. cerevisiae* CF) and ten wine yeast strains (*S. cerevisiae* var. *ellipsoideus* 1, 3, 4, 6, 8 and 9; *S. bayanus* var. *oviformis* 2, 5 and 7; and *S. uvarum* 10). The other, 5 yeast strains belong to different species: *Kloeckera apiculata*, *Candida mycoderma* (*Mycoderma vini*), *Pichia membranaefaciens*, *Rhodotorula glutinis* and *Torulopsis holmii*, respectively.

Examinarea caracterelor morfologice celulare și culturale ale tulpinilor de drojdie izolate în culturi pure prezintă o importanță deosebită în caracterizarea și identificarea acestora [1, 2, 4].

În lucrarea de față au fost examinatele caracterele morfologice ale tulpinilor de drojdie izolate în culturi pure prin examen microscopic, caracterul culturilor pe mediu solid și lichid, capacitatea de formare a sporilor și aspectul coloniilor gigant.

Materiale și metode. Am întreprins un studiu asupra caracterelor morfologice ale unui număr de 20 tulpi de drojdie, majoritatea cu aplicații în industria fermentativă, dintre care 15 tulpi de drojdie alcooligene aparținătoare genului *Saccharomyces*: două de bere (*S. carlsbergensis* A și B), două de panificație (*S. cerevisiae* CA și CP), una de spirt (*S. cerevisiae* CF), 10 de vin (*S. cerevisiae* var. *ellipsoideus* 1, 3, 4, 6, 8 și 9, *S. bayanus* var. *oviformis* 2, 5 și 7 și *S. uvarum* 10) și 3 tulpi de drojdie de contaminare (*Candida mycoderma*, *Pichia membranaefaciens* și *Rhodotorula glutinis*) au fost izolate sub formă de culturi pure, identificate și selecționate în cadrul Laboratorului de microbiologie al Catedrei de Biotehnologii alimentare, Universitatea "Lucian Blaga" Sibiu. Am supus tulpinele de drojdie, izolate în culturi pure, unei examinări biologice complexe, determinând caracterele lor morfologice celulare și culturale prin examenul microscopic al celulei vegetative de drojdie și a sporilor (în cazul drojdilor ascosporigene), prin examenul caracterelor culturilor pe mediu solid și lichid și a aspectului coloniilor gigant [1, 3].

Pe preparate umede și frotiuri colorate Gram am examinat la microscop (cu grosismenit 900X și, respectiv, 450X) forma și mărimea celulelor, formarea mugurilor, locul apariției lor pe celula mamă, formarea de pseudomiceliu (abundente ramificații). Dimensiunile celulelor au fost măsurate cu un ocular sau cu o lamă micrometrică [3].

* Universitatea "Lucian Blaga", Catedra de Biotehnologii alimentare, 2400 Sibiu, România.
E-mail: letitia_oprean@yahoo.com

Sporularea a fost urmărită prin cultivarea tulpinilor de drojdii pe mediul Gorodkova sau acetat-agar. Examinarea caracterelor culturale (macroscopice) ale tulpinilor de drojdii izolate în culturi pure și dezvoltate în/pe medii lichide și solide specifice a permis caracterizarea și diferențierea lor. Observațiile s-au efectuat după 3 și 7 zile de incubare în termostat la 25°C și după 30 zile de menținere la 17°C.

Dezvoltarea tulpinilor de drojdii pe medii solide a permis caracterizarea coloniilor obținute după formă, dimensiuni, profil, aspect, consistență și culoare. Sunt evidențiate și caracterele coloniei gigant pe mediul adecvat. Am utilizat următorul procedeu care a dat rezultate foarte bune. Cu ajutorul unei pipete Pasteur am lăsat să cadă o picătură din cultura de drojdie în centrul unei plăci Petri cu mediul adecvat. Incubarea s-a efectuat în termostat 7 zile la 25 °C și 30 de zile la 17 °C. Am obținut și culturi în strii pe mediul înclinat adecvat [3].

În cazul drojdiilor de vin s-a studiat și aspectul coloniilor pe must de struguri gelificat, precum și puterea de lichefiere a gelatinei (proteoliză). Ultimul caracter reprezintă un criteriu de diferențiere a celor două tulpini de drojdii predominante în mustul de fermentație (*S. cerevisiae* var. *ellipsoideus*) și vin Tânăr (*S. bayanus* var. *oviformis*) [3].

Rezultate. Caracterele morfologice (macroscopice și microscopice) ale tulpinilor de drojdii izolate și obținute sub formă de culturi pure în laborator sunt prezentate în continuare.

Tulpinile de drojdii *Saccharomyces carlsbergensis* (sin.: *S. uvarum*) A, B (bere) și 10 (vin) formează pe mediul must de bere-geloză, după 3 zile de incubare la 25°C, colonii de tip S (smooth), cu diametrul de 2 - 3,5 mm, cu aspect neted și mat, de culoare alb-crem, cu perimetru circular, uneori lobat triangular, cu profil bombat sau lenticular și consistență cremoasă. După 7 zile, coloniile ajung la diametrul de 3,5 – 8 mm și își mențin caracterele morfologice. După 30 zile, coloniile gigant se colorează de la alb-crem la crem închis sau maroniu și prezintă un perimetru lobat cu suprafața netedă sau pliată, verucoasă, restrânsă sau împrăștiată, profil plat sau bombat și margini ondulate, festonate sau lobiforme. Pe must cu gelatină, coloniile nu lichefiază gelatina.

Culturile în strii la *S. uvarum* 10 (vin), pe must de bere sau extract de malț-geloză, sunt mate cu suprafața netedă, de consistență cremoasă, de culoare bej, uniforme și fără relief; la *S. carlsbergensis* A și B (bere), culturile în strii sunt albe lucioase, mai puțin strălucitoare în centru.

Culturile de drojdii în mediul lichid (must de bere), după 24 ore de incubare la 25°C, produc o tulbureală opacă, densă, la suprafață formează o spumă slab persistentă, iar după 3 zile un sediment compact aderent (depozit flocunos). După 7 zile, aceste drojdii nu formează peliculă la suprafața mediului. *S. uvarum* 10 (vin), în mustul de bere, după 7 zile, formează la suprafață, ocazional, un inel incomplet, care se păstrează și după 30 zile.

Microscopic, după 3 zile de termostatare la 25°C, în must de bere, celulele au formă ovală sau sferică, globoasă, cu dimensiuni: (3-7) x (4-10) µm și dispuse singular sau în perechi, uneori în lanțuri scurte sau ciorchini mici. S-au observat și

celule înmugurite cu muguri dispuși polar. În mustul de struguri, celulele sunt ovale, puțin alungite, cu dimensiuni mari: (4-8) x (5-12) μm . Pe mediul Gorodkova sau acetatagar, se formează 1 - 4 ascospori într-un asc (de obicei 2 sau 4), prin partenogeneză sau copulare.

Tulpinile de drojdii *Saccharomyces cerevisiae* CA și CP (panificație) și CF (spirt) formează pe must de bere-geloză, după 3 zile de incubare la 25°C, colonii de tip S, cu diametrul de 1 - 2 mm, cu perimetru circular, uneori lobat triangular, profil convex, consistență cremoasă, aspect neted, lucios și de culoare alb-crem (*S. cerevisiae* CF) sau alb-cenușiu (*S. cerevisiae* CA și CP). După 7 zile, coloniile ajung la diametrul de 2,5 - 5 mm și își mențin caracterele morfologice. După 30 zile, coloniile gigant devin plate, ușor bombate, de culoare crem și maroniu închis, cu suprafața netedă sau pliată, zbârcită.

Culturile în strii pe must de bere-geloză sau extract de malțagar sunt lucioase de culoare albă-bej (*S. cerevisiae* CF) sau alb-cenușiu (*S. cerevisiae* CA și CP), uniforme, fără relief.

Culturile de drojdii în mediul lichid (must de bere), după 24 ore de incubare la 25°C, produc tulbureală, iar la suprafață formează o spumă fină, cu aspect mat. După 3 zile, se formează un sediment compact, aderent (depozit flocunos), ceea ce demonstrează că drojdiile de panificație și de spirt, ca și cele de bere, sunt drojdii floculanți. Ocazional, unele tulpini (*S. cerevisiae* CF) formează și un inel incomplet la suprafața mediului. Nu se formează peliculă nici în culturile bătrâne (peste 30 zile).

Microscopic, după 3 zile, celulele crescute pe must de struguri au formă, în general, eliptică, cu dimensiuni: (5-9) x (4-8) μm , unele fiind mai alungite, altele mai rotunjite. Ele sunt dispuse izolat, în perechi sau lanțuri. S-au observat și celule cu muguri dispuși multipolar. Uneori, după incubare prelungită pe extract de malțagar, precum și pe mediul Gorodkova, s-au format ascospori. Ascii provin din celule vegetative, când sporii s-au format prin partenogeneză. Un asc conține, de cele mai multe ori, 3 - 4 ascospori sferici sau ovali cu suprafața netedă .

Tulpinile de drojdii *Saccharomyces cerevisiae* var. *ellipsoideus* 1, 3, 4, 6, 8 și 9 formează pe must de struguri-geloză, după 3 zile, colonii de tip S, cu diametrul de 2-4 mm cu perimetru circular, slab lobat, uneori triangular, cu profil convex, de consistență umedă, cremoasă, mai puțin netedă sau granuloase, lucioase de culoare albă sau alb-cenușiu. După 7 zile, coloniile ajung la diametrul de 4-6 mm, unele colonii devin proeminente, țuguiate, iar restul caracterelor morfologice se mențin constante. După 30 zile, coloniile gigant prezintă în funcție de tulpini aspecte multiple. Sunt bine dezvoltate, mai plate sau ușor bombate, de culoare crem-maroniu, cu suprafața netedă sau pliată, cu margini bine dezvoltate ondulate, larg festonate.

Culturile în strii pe must de malț-geloză sau must de struguri-geloză sunt lucioase, de culoare albă, uneori alb-gălbui, uniforme, fără relief. Pe must cu gelatină, coloniile sunt mate sau strălucitoare, de dimensiuni mari (3-5 mm diametru), în general destul de groase, deseori proeminente, țuguiate. Prezintă proprietatea de a lichefia gelatina.

Culturile în mediul lichid (mustul de struguri), după 24 ore de incubare la 25°C, produc tulbureală, o spumă fină, cu aspect mat, iar după 72 ore, formează un sediment granular, aderent (depozit flocunos), ceea ce demonstrează că și aceste tulpini sunt drojdi floculante. Separarea rapidă la sfârșitul fermentației prin floculare reprezintă o proprietate tehnologică importantă, după care sunt selecționate în laborator drojdiile de vin. Caracterele de sedimentare sunt variabile de la o tulpină la alta. Pe culturile mai în vîrstă (după 30 zile), nu se remarcă o peliculă la suprafață, dar poate rămâne pe eprubetă un inel format de drojdiile ridicate de spumă în timpul fermentației active.

Microscopic, după 3 zile de cultivare la 25°C, în must de struguri, celulele au o formă elipsoidală, globoasă sau elipsoidală lungită, cu dimensiuni de (5-8) x (4-8) µm, dispuse singular, în perechi sau lanțuri. S-au observat și celule înmugurite, cu muguri dispuși, de obicei, polar și care au rămas alipiți de celula mamă. Pe mediul Gorodkova sau acetat-agar, *S. cerevisiae* var. *ellipsoideus* sporulează ușor. După o lună s-au observat asce ce conțin 1-4 ascospori rotunzi sau ușor ovali. Formarea sporilor este precedată de copulare.

Tulpinile de drojdi *Saccharomyces bayanus* var. *oviformis* 2, 5 și 7 formează pe mediul must de bere-geloză sau must de struguri-geloză, după 3 zile de incubare la 25°C, colonii de tip S, cu diametrul de 2-3,5 mm, cu contur circular sau lobat triangular, cu profil bombat lenticular, de consistență cremoasă, cu suprafață netedă, lucioasă și de culoare albă sau alb-cenușie (se asemănă cu coloniile de *S. cerevisiae* var. *ellipsoideus*). După 7 zile, colonia ajunge la diametrul de 2,5-5 mm; unele colonii devin țuguiate, iar restul caracterelor morfologice se mențin la fel. După 30 zile, coloniile gigant prezintă aspecte variate: se asemănă și se confundă cu cele ale lui *S. cerevisiae* var. *ellipsoideus*. Unele colonii devin mai plate, altele cu suprafață zbârcită, cutată, cu margini ondulate și festonate de culoare mai închisă (maroniu).

Culturile în strii pe must de malț-geloză și must de struguri-geloză sunt destul de abundente și compacte uneori netede și lucioase, de culoare albă. Cultivate pe must cu gelatină, aceste drojdi nu lichefiază gelatina (nu produc proteoliză).

Culturile de drojdi în mediul lichid (must de struguri), după 24 ore de incubare la 25°C, produc o slabă opalescență și rareori o spumă fină cu aspect mat, iar după 3 zile, se formează un sediment pulverulent, granular, neaderent, ușor de îndepărtat. După 7 zile, se poate forma o peliculă la suprafața vinurilor, uneori se formează un inel superficial.

Microscopic, celulele de *S. bayanus* var. *oviformis* crescute în must de struguri, după 3 zile, au formă ovală, eliptică, sunt uneori aproape rotunde, cu dimensiuni de (4-7) x (5-9) µm, dispuse singular, în perechi sau lanțuri. Pe must de struguri-geloză, celulele sunt mai alungite, uneori foarte alungite. Pe unele celule se observă muguri dispuși polar, care în majoritate rămân alipiți de celula mamă. Pe must-geloză, celulele sunt mai puțin înguste și mai alungite în mediul solid: (3-6) x (5-9) µm, cu câteva celule foarte alungite. Pe mediul Gorodkova sau acetat-agar, aceste drojdi formează celule mai alungite și sporulează mai greu decât *S. cerevisiae* var. *ellipsoideus*; formează 2 - 3 spori rotunzi într-un asc.

Kloeckera apiculata formează pe must de struguri-geloză sau pe must de malț-geloză, după 3 zile la 25°C, colonii de tip S cu diametrul de 1,5-2,5 mm, aproape semisferice, slab lobate, cu margini circulare, aspect neted și lucios, de culoare alb-crem. După 7 zile, coloniile ajung la diametrul de 3-6 mm, unele prezintând un aspect rugos, puțin încrețit; restul caracterelor morfologice se mențin. După 30 zile, majoritatea coloniilor gigant devin gri sau de culoare gălbui spre crem. Coloniile sunt de obicei catifelate, lucioase, uneori încrețite cu marginea întreagă până la neregulată. Majoritatea coloniilor au o regiune centrală joasă și fine striații radiale, ocazional observându-se și crateri mici.

Culturile în strii pe must de malț-geloză sunt lucioase, subțiri, de culoare crem, uniforme, fără relief. Pe must-gelatină, coloniile lichefiază gelatina.

Cultura de drojdie în mediul lichid (must de struguri), după 3 zile de incubare, nu produce tulbureală. Lichidul este limpede, cu urme de inele superficiale. Nu se formează peliculă la suprafața mediului.

Microscopic, celulele de *K. apiculata*, crescute pe must de struguri, după 3 zile la 25°C, au formă fie elipsoidală, subțiri, mici: (3-1,5) µm sau groase, mai mari: (5-6) x (3-4) µm, fie apiculată cu dimensiuni de (3-6) x (7-10) µm, datorită mugurilor terminali bipolari. Pe malț-geloză, celulele sunt foarte mici și de aceeași formă. Pe mediul Gorodkova, nu formează spori.

Candida mycoderma (Mycoderma vini) crescută pe must de struguri-geloză formează, după 3 zile, colonii de tip S/R cu un perimetru slab lobat, neregulat cu o proeminență în partea centrală, margini ondulate și cu diametrul de 2,5-5,8 mm. Suprafața coloniei are un aspect mat, neted lucios sau cutat de culoare alb-cenușiu, uneori alb-crem. După 7 zile, coloniile ajung la diametrul de 4,5-7 mm și își mențin caracterele morfologice. După 30 zile, coloniile gigant devin mai plate și prezintă o suprafață zbârcită sau încrețită cu perimetru lobat și neregulat. În mediu cu gelatină, nu lichefiază gelatina.

Culturile în strii pe must de malț-geloză sunt mate, uniforme, fără relief, de culoare alb-crem.

Cultura în mediu lichid (Rieder) formează tulbureală, iar în scurt timp o peliculă albă sau galbenă la suprafața mediilor slab-alcoolice și subțire ce devine uscată și încrețită după 3-4 zile. Cu timpul se îngroașă și o parte din ea se destramă și se depune pe fundul vasului sub formă de sediment pulverulent. Formarea peliculei reprezintă un criteriu taxonomic important în identificarea drojdilor oxidative (peliculare).

Microscopic, celulele de *C. mycoderma* crescute pe must de struguri-geloză, timp de 3 zile, sunt de formă elipsoidală, alungită, cilindrică, cu extremitățile rotunjite sau slab ascuțite și cu dimensiuni cuprinse între (2-4) x (6-13) µm. Prezintă una sau două granulații refringente. S-au observat și celule cu muguri dispuși multipolar. În culturile mai vechi (14 zile), această drojdie formează pseudomiceliu cu blastospori (muguri grupați sub formă de ciorchine ce rămân alipite de celula mamă). Formarea de pseudomiceliu reprezintă un criteriu taxonomic de bază în identificarea acestei drojdi. În must de struguri, celulele de la suprafață sunt mai alungite, decât cele care cad la fund. Pe mediul Gorodkova, nu se formează spori.

Pichia membranaefaciens formează pe mediu must de bere-geloză, după 3 zile, colonii de tip S/R cu diametrul de 2,5-4 mm, cu perimetru slab lobat sau chiar neregulat, margini ondulate sau lobiforme, aspect mat, neted, catifelat sau zbârcit, rugos, de culoare albă sau alb-cenușie. După 7 zile, coloniile ajung la diametrul de 4-5,5 mm și își mențin caracterele morfologice. După 30 zile, colonia gigant este albă-gălbui spre maro-gălbui, moale, catifelată, parțial sau total zbârcită. Însămânțată pe mediul cu făină de porumb și cartof, celulele formează pseudomiceliu bine dezvoltat cu aspect de capac sau poate fi absent.

Culturile în strii pe must de bere-geloză sunt mate, prăfoase, de culoare albă, uniforme, fără relief. În anumite cazuri, coloniile pe must gelificat au lichefiat gelatina.

În mustul de struguri după 3 zile, se formează tulbureală, iar la suprafața mediului apare o peliculă groasă, plisată, de culoare alb-gălbui, mată, relativ friabilă, care se depune producând un sediment pulverulent. Formarea peliculei reprezintă un criteriu taxonomic important care stă la baza clasificării acestei drojdii, deoarece prezintă un metabolism predominant oxidativ.

Microscopic, celulele de *P. membranaefaciens*, crescute în must de struguri, sunt ovale alungite, cu capetele ușor ascuțite sau rotunjite, cu dimensiuni de (3-10) x (2-4) μm și dispuse în lanțuri mai mult sau mai puțin lungi sau ramificate. S-au observat și celule cu muguri dispuși pe toată suprafața celulei. Formează în culturile mai vechi (30 zile) pseudomicelii cu blastospori rezultați prin înmugurire multilaterală sau hife adevarate (cu pori în septum). Pe mediul solid (must de malț-geloză), celulele sunt mai rotunjite, izolate sau câte două. Pe mediul Gorodkova, se formează spori de dimensiuni variabile cu forme diferite: sferici sau hemisferici, saturn sau de pălărie, cu suprafață netedă sau rugoasă care se eliberează ușor din asce. Într-un asc se găsesc 1 - 4 spori, uneori 8. Prezența unui număr mare de spori dă culturii culoare maro.

Rhodotorula glutinis formează pe mediul extract de malț sau must de bere-geloză, după 3 zile, colonii de tip S/R, cu diametrul de 1,5 - 3 mm, cu perimetru circular, margini drepte, profil convex, cu aspect neted lucios sau consistență mucoasă, de culoare roșu pastelat. Prezența pigmentelor carotenoizi de culoare roz și roșu este foarte caracteristică genului *Rhodotorula* și reprezintă un criteriu taxonomic delimitativ al acestui gen de drojdii. După 7 zile, coloniile ajung la diametrul de 3,5-7 mm și își mențin caracterele morfologice. După 30 zile, coloniile gigant devin roz spre portocaliu. Suprafața coloniei variază de la catifelat, adesea cu striații fine transversale, până la zbârcită și cu aspectul foarte lucios sau semilucios. Consistența variază de la mucoasă până la ușor dură. Profilul este plat spre larg convex cu marginea neregulată spre întreagă. Uneori în culturile vechi (30 zile), se diferențiază și un pseudomiceliu. În culturile crescute pe mediul cu făină de porumb, pseudomiceliul, în general, lipsește sau este rudimentar. În ocazii rare, se pot forma pseudohife bine dezvoltate sau hife adevarate. Uneori se produc clamidospori mari pigmentați puternic.

Culturile în strii pe extract de malț sau must de bere-geloză sunt lucioase, de culoare roz-roșie, uniforme, fără relief.

În mustul de struguri, drojdia nu fermenteaază. După 30 zile, mediul rămâne limpede și se depune un sediment fin, iar la suprafață se formează o peliculă subțire (drojdie strict oxidativă).

Microscopic, celulele de *R. glutinis*, crescute pe mediul extract de malț, timp de 3 zile, sunt globulare, puternic vacuolate, dispuse izolat sau în perechi cu dimensiuni medii: (4-5,5) x (3-3,5) μm . S-au observat și muguri pe întreaga suprafață a celulei. Unele celule au formă alungită. Se pot dezvolta pe mediul Czapek, formând celule globulare cu dimensiuni de (2,3-5) x (12-16) μm și uneori pseudomiceliu; deci pot folosi azotații ca sursă de azot, dar nu se dezvoltă pe extract de malț agar + 0,5 % acid acetic. Pe mediul Gorodkova, nu se formează spori.

Torulopsis holmii formează pe must de malț-geloză, după 3 zile de cultivare la 25°C, colonii de tip S cu diametrul de 3-5 mm, de culoare albă, perimetru circular, profil slab convex, aplatizate cu margini drepte, cu suprafață lucioasă. După 7 zile, coloniile ajung la diametrul de 4-7,5 mm și își mențin caracterele morfologice. După 30 zile, coloniile gigant sunt colorate crem-cenușiu, au formă circulară, uneori neregulată, margini lobate, aplatizate, lucioase.

Culturile în strii pe must de malț-geloză sunt lucioase, netede, de culoare albă, uniforme, fără relief.

În mustul de struguri, după 3 zile de cultivare, se observă o tulbureală și un sediment fin.

Microscopic, celulele crescute pe must de malț-geloză, timp de 30 zile, sunt mici, elipsoidale, cu dimensiuni de (2-4) x (3-5) μm , izolate sau în lanțuri sau grămăjoare. Pe unele celule se observă muguri dispuși lateral. Pe mediul Gorodkova, nu se formează spori.

Concluzii. 1. Examinarea caracterelor morfologice celulare ale tulpinilor de drojdi izolate a permis diferențierea lor după mărimea și forma celulelor (*Saccharomyces*, *Pichia*, *Candida*, *Rhodotorula*, *Kloeckera* și *Torulopsis*).

2. După capacitatea de formare a sporilor s-au diferențiat tulpinile de drojdi sporogene aparținătoare genurilor *Saccharomyces* și *Pichia*.

3. Examinarea caracterelor culturale (macroscopice) ale tulpinilor de drojdi a evidențiat deosebiri în aspectul coloniilor gigant dezvoltate pe medii adecvate solide, funcție de care pot fi diferențiate aceste tulpi de drojdi.

4. Prezența pigmentilor carotenoizi de culoare roz sau roșu la drojdile aparținătoare genului *Rhodotorula* a permis identificarea drojdiei de contaminare izolate, *Rhodotorula glutinis*.

5. Proprietatea drojdii peliculară (oxidative) de a forma pelicula la suprafață mediilor lichide alcooligene este utilizată drept criteriu de identificare a drojdii de contaminare izolate (*Candida mycoderma* și *Pichia membranaefaciens*).

6. Proprietatea drojdiilor de vin izolate de a lichefia gelatina (proteoliză) este utilizată drept criteriu de diferențiere a celor două tulpini de drojdii de vin predominante în mustul de fermentație: *Saccharomyces cerevisiae* var. *ellipsoideus* și *Saccharomyces bayanus* var. *oviformis*. *S. bayanus* var. *oviformis* nu lichefiază gelatina, iar *S. cerevisiae* var. *ellipsoideus* prezintă acest caracter.

B I B L I O G R A F I E

1. Anghele, I., Tomă, N., Voică, C., Cojocaru, I., *Biologia și tehnologia drojdiilor*, Vol. II, p. 303-324, Ed. Tehn., București, 1989.
2. Kiss, S., *Microbiologie generală*, Vol. II, p. 1-16, Univ. "Babeș-Bolyai", Cluj, 1975.
3. Oprea, L., *Drojdii industriale*, p. 180-196, Univ. "Lucian Blaga", Sibiu, 2002.
4. Riberau - Gayon, J., Peyraud, E., *Traité d'oenologie: Maturation du raisin. Fermentation alcoolique. Vinification*, Vol. I, p. 36-42, Dunod, Paris, 1969.

ADDITIONAL BIBLIOGRAPHY ON ENZYMOLOGY OF DISTURBED SOILS

STEFAN KISS*, MIHAIL DRĂGAN-BULARDA* and DANIELA PAŞCA **

This additional bibliography comprises the papers not referred to in any of the three Parts of the book "*Enzymology of Disturbed Soils*" (XIV + 336 pages), which was elaborated by S. Kiss, D. Paşa and M. Drăgan-Bularda and published by Elsevier Science, Amsterdam, in 1998. In Part I. Enzymology of Oil-Contaminated Soils, Part II. Enzymology of Soils Affected by Industrial Emissions and Part III. Enzymology of Technogenic Soils 74, 149 and 237 papers, respectively, were referred to.

The additional bibliography consists of three separate lists, *i.e.* a separate bibliographical list will be added to each of the three Parts of the book. The papers in each bibliographical list are presented based on the system applied in *Enzymology of Disturbed Soils*. If a paper covers topics belonging to two Parts of the book, it is cited only once, namely in the additional bibliographical list of the first of the two Parts. Similarly, the additional bibliographical list of Part I comprises the papers dealing with topics of all Parts.

The soil enzymological investigations described in the 64, 74 and 61 papers cited in the three additional bibliographical lists were carried out in the following countries: Australia, Austria, Azerbaijan, Byelorussia, Canada, Czech Republic, Denmark, Finland, France, Germany, Hungary, Ireland, Italy, Kuwait, The Netherlands, New Zealand, Poland, Romania, Russian Federation (Russia, Bashkiria, Komi Republic, Tataria), Spain, Switzerland, Turkey, Ukraine, United Kingdom, United States of America, Uzbekistan, Venezuela.

It should be mentioned that of the 64, 74 and 61 papers forming the three additional bibliographical lists, 22, 38 and 19 papers, respectively, were cited in a review article, too [S. Kiss (2001) Advances in soil enzymology (Parts I-III). *Stud. Univ. Babeş-Bolyai, Biol.*, **46** (1), 3-48].

PART I. ENZYMOLOGY OF OIL-CONTAMINATED SOILS

Akhundova, L.Kh. and Maslovetskaya, G.Yu. (1980) Vliyanie nefteproduktov na fermentativnye protsessy v pochve (Effect of petroleum products on the enzymatic processes in soil). In *Nauchnye Osnovy Gigieny Okruzhayushchei Sredy i Infektsionnoi Patologii*, pp.29-33. Baku.

Alekhin, V.G., Emtsev, V.T., Rogozina, E.A. and Fakhrutdinov, A.I. (1998) Biologicheskaya aktivnost' i mikrobiologicheskaya rekul'tivatsiya pochv, zagryaznennykh nefteproduktami (Biological activity and microbiological recultivation of soils polluted with petroleum products). *Biologicheskie Resursy i Prirodopol'zovanie*, No. 2, 95-105.

* Babeş-Bolyai University, Department of Experimental Biology, Laboratory for Environmental Enzymology and Microbiology, 400084 Cluj-Napoca, Romania. E-mail: skiss@bioge.ubbcluj.ro and draganb@bioge.ubbcluj.ro

** Biological Research Institute, 400015 Cluj-Napoca, Romania. E-mail: icb@mail.dntcj.ro

- Allen, C.A. and Wagner, S.C. (1999) Rhizosphere activity of plants inhabiting an oil spill in Roy E. Larsen Sandyland Sanctuary. *Abstracts, 99th General Meeting of the American Society of Microbiology* (Chicago, 1999), p. 609.
- Andreson, R.K., Boiko, T.F., Bagautdinov, F.Ya., Danilenko, L.A., Denezhkin, E.M., Novoselova, E.I., Khaziev, F.Kh. and Andreson, B.A. (1994) Primenenie biologicheskogo metoda dlya ochistki i rekul'tivatsii neftezagryaznennykh pochv (Application of the biological method for decontamination and recultivation of petroleum-polluted soils). *Zashchita ot Korrozii i Okhrana Okruzhayushchey Sredy*, No. 2, 16-18.
- Balba, M.T., Al-Awadhi, N. and Al-Daher, R. (1998) Bioremediation of oil-contaminated soil: microbiological methods for feasibility assessment and field evaluation. *Journal of Microbiological Methods*, **32**, 155-164.
- Bauer, E., Pennerstorfer, C., Kandeler, E. and Braun, R. (1993) Einfluß von Kohlenwasserstoffen auf bodenbiologische Parameter (Influence of hydrocarbons on soil biological parameters). In *Ecoinforma '92. 2. Internationale Tagung und Ausstellung über Umweltinformation und Umweltkommunikation "Bodensanierung, Bodenkontamination, Verhalten und ökotoxikologische Wirkung von Umweltchemikalien im Boden"* (K. Alef, H. Fiedler and O. Hutzinger, Eds.), pp. 351-363. Ecoinforma Press, Bayreuth.
- Brohon, B. and Gourdon, R. (2000) Influence of soil microbial activity level on the determination of contaminated soil toxicity using Lumistox and MetPlate bioassays. *Soil Biology & Biochemistry*, **32**, 853-857.
- Brohon, B., Delolme, C. and Gourdon, R. (2001) Complementarity of bioassays and microbial activity measurements for the evaluation of hydrocarbon-contaminated soils quality. *Soil Biology & Biochemistry*, **33**, 883-891.
- Frankenberger, W.T., Jr. (1997) Factors affecting the fate of urea peroxide added to soil. *Bulletin of Environmental Contamination and Toxicology*, **59**, 50-57.
- Gabbasova, I. M., Abdurakhmanov, R. F. and Batanov, V.N. (1996) Vliyanie zagryazneniya neft'yu i neftepromyslovymi stochnymi vodami (NSV) na vlazhno-lugovye pochvy (Effect of the pollution with petroleum and petroleum-industrial wastewaters (PWW) on humid meadow soils). *Tezisy Dokladov Mezhdunarodnoi Nauchno-Prakticheskoi Konferentsii "Geoekologiya v Uralo-Kaspinskem Regione"* (Ufa, 1996), (Part 2), 176-179.
- Gabbasova, I.M., Abdurakhmanov, R.F., Khabirov, I.K. and Khaziev, F.Kh. (1997) Izmenenie svoistv pochv i sostava gruntovych vod pri zagryaznenii neft'yu i neftepromyslovymi stochnymi vodami v Bashkirii (Changes in soil properties and composition of groundwaters following pollution with petroleum and petroleum-industrial wastewaters in Bashkaria). *Pochvovedenie*, No. 11, 1362-1372. Changes in soil properties and groundwater composition under the impact of pollution by oil and sewage waters from the oil fields in Bashkortostan. *Eurasian Soil Science*, No. 11, 1219-1229.
- Gabbasova, I.M., Khaziev, F.Kh., Suleimanov, R.R., Galimzyanova, N.F. and Boiko, T.F. (1999) Stimulirovaniye mikroflory neftezagryaznennoi seroi lesnoi pochvy s pomoshch'yu organiceskikh dobavok (Stimulation of the microflora in petroleum-polluted grey forest soil by means of organic additions). *Bashkirskii Ekologicheskii Vestnik*, No. 2, 14-18.

ADDITIONAL BIBLIOGRAPHY ON ENZYMOLOGY OF DISTURBED SOILS

- Janke, S. and Kunze, C. (1993) Beeinflussung der bodenmikrobiologischen Aktivität durch Heizöl (Influencing of the soil microbiological activity by fuel oil). In *Ecoinforma '92. 2. Internationale Tagung und Ausstellung über Umweltinformation und Umweltkommunikation "Bodensanierung, Bodenkontamination, Verhalten und ökotoxikologische Wirkung von Umweltchemikalien im Boden"* (K. Alef, H. Fiedler and O. Hutzinger, Eds.), pp. 373-384. Ecoinforma Press, Bayreuth.
- Khaziev, F.Kh. (Ed.) (1997) *Pochvy Bashkirtostana. Tom 2. Vozproizvodstvo Plodorodiya: Zonal'no-Ekologicheskie Aspekty* (Soils of Bashkiria. Vol. 2. Reproduction of Fertility: Zonal-Ecological Aspects), pp. 288-299. Izdatel'stvo Gilem, Ufa.
- Kireeva, N.A. (1999) Biologicheskie metody rekultivatsii narushennykh pochv (Biological methods of the recultivation of disturbed soils). *Bashkirskii Ekologicheskii Vestnik*, No.3, 82-84.
- Kireeva, N.A. and Yamaletdinova, G.F. (1999) Aktivnost' askorbatoksidazy kak diagnosticheskii pokazatel' zagryazneniya pochv neft'yu (Activity of ascorbate oxidase as a diagnostic indicator of the pollution of soils with petroleum). *Itogi Biologicheskikh Issledovanii Bashkirskogo Gosudarstvennogo Universiteta, Ufa, 1998*, No. 5, 71-75.
- Kireeva, N.A. and Yamaletdinova, G.F. (2000) Fenoloksidaznaya aktivnost' neftezagryaznennykh pochv (Phenol oxidase activity of petroleum-polluted soils). *Vestnik Bashkirskogo Universiteta*, No. 1, 48-51.
- Kireeva, N.A. and Yamaletdinova, G.F. (2001) Vliyanie neftyanykh fraktsii na protsessy sernogo obmena pochv (Effect of petroleum fractions on the processes of sulphur metabolism of soils). *Itogi Biologicheskikh Issledovanii Bashkirskogo Gosudarstvennogo Universiteta, Ufa, 2000*, No. 6, 120-122.
- Kireeva, N.A., Novoselova, E.I. and Khaziev, F.Kh. (1998a) Izmenenie svoistv seroi lesnoi pochvy pri zagryaznenii neft'yu i v protsesse rekultivatsii (Changes in properties of grey forest soil following pollution with petroleum and in the process of recultivation). *Bashkirskii Ekologicheskii Vestnik*, No. 3, 3-7.
- Kireeva, N.A., Novoselova, E.I. and Khaziev, F.Kh. (1998b) Aktivnost' karbogidraz v neftezagryaznennykh pochvakh (Activity of carbohydrases in petroleum-polluted soils). *Pochvovedenie*, No. 12, 1444-1448. Carbohydrase activity in oil-contaminated soils. *Eurasian Soil Science*, No. 12, 1314-1318.
- Kireeva, N.A., Vodop'yanov, V.V. and Miftakhova, A.M. (2000a) Vliyanie neftyanoogo zagryazneniya na tsellyulaznuyu aktivnost' pochv (Effect of petroleum pollution on the cellulase activity of soils). *Pochvovedenie*, No. 6, 748-753.
- Kireeva, N.A., Yamaletdinova, G.F., Miftakhova, A.A. and Novoselova, E.I. (2000b) Kompleksnaya sistema pokazatelei biologicheskoi aktivnosti pochv dlya indikatsii toksichnosti neftyanykh zagryaznenii (Complex system of indicators of the biological activity of soils for indication of toxicity of petroleum pollutions). *Bashkirskii Ekologicheskii Vestnik*, No.2, 26-29.
- Kireeva, N.A., Onegova, T.S. and Novoselova, E.I. (2000c) Ispol'zovanie fermentativnoi aktivnosti dlya bioindikatsii neftezagryaznennykh i rekul'tiviruemnykh pochv (Use of the enzymatic activity for bioindication of petroleum-polluted and recultivated soils). *Tezisy Dokladov Vserossiiskoi Nauchno-Prakticheskoi Konferentsii "Aktual'nye Voprosy Monitoringa Ekosistem Antropogenno-Narushennykh Territorii"* (Ulyanovsk, 2000), pp. 51-53.

- Kireeva, N.A., Tarasenko, E.M. and Onegova, T.S. (2001a) Biologicheskaya rekultivatsiya neftezagryaznennykh pochv (Biological recultivation of petroleum-polluted soils). *Itogi Biologicheskikh Issledovanii Bashkirskogo Gosudarstvennogo Universiteta, Ufa, 2000*, No. 6, 115-118.
- Kireeva, N.A., Novoselova, E.I. and Yamaletdinova, G.F. (2001b) Aktivnost' oksidoreduktaz v neftezagryaznennykh i rekultiviruemых pochvakh (Activity of oxidoreductases in petroleum-polluted and recultivated soils). *Agrokhimiya*, No. 4, 53-60.
- Kireeva, N.A., Novoselova, E.I. and Yamaletdinova, G.F. (2001c) Diagnosticheskie kriterii samoochishcheniya pochvy ot nefti (Diagnostic criteria of the self-decontamination of soil from petroleum). *Ekologiya i Promyshlennost' Rossii*, No. 12, 34-35.
- Kireeva, N.A., Vodop'yanov, V.V. and Miftakhova, A.M. (2001d) *Biologicheskaya Aktivnost' Neftezagryaznennykh Pochv* (Biological Activity of Petroleum-Polluted Soils). Izdatel'stvo Gilem, Ufa.
- Kireeva, N.A., Vodop'yanov, V.V., Novoselova, E.I., Onegova, T.S. and Zhdanova, N.V. (2001e) *Mikrobiologicheskaya Rekul'tivatsiya Neftezagryaznennykh Pochv* (Microbiological Recultivation of Petroleum-Polluted Soils). OAO "VNIIOENG", Moscow.
- Kireeva, N.A., Novoselova, E.I. and Yamaletdinova, G.F. (2001f) Otsenka sostoyaniya pochv, zagryaznennykh neft'yu, po pokazatelyami fermentativnoi aktivnosti (Evaluation of the state of soils, polluted with petroleum, according to indicators of enzymatic activity); Enzymatic activity as bioindicator of oil pollution of soils. *Tezisy Dokladov, XI. Mezhdunarodnyi Simpozium po Bioindikatoram "Sovremennye Problemy Bioindikatsii i Biomonitoringa"; Abstracts, XI. International Symposium on Bioindicators "Problems of Today in Bioindication and Biomonitoring"* (Syktyvkar, 2001), pp. 77-78; 270-271.
- Kireeva, N.A., Tarasenko, E.M. and Onegova, T.S. (2001g) Podbor bioindikatsionnykh pokazatelei diagnostiki neftezagryaznennykh i rekultiviruemых pochv; Selection of bioindicator indexes for diagnosis of oil-contaminated and recultivated soils. *Tezisy Dokladov, XI. Mezhdunarodnyi Simpozium po Bioindikatoram "Sovremennye Problemy Bioindikatsii i Biomonitoringa"; Abstracts, XI. International Symposium on Bioindicators "Problems of Today in Bioindication and Biomonitoring"* (Syktyvkar, 2001), pp. 78-79; 271.
- Kireeva, N.A., Yamaletdinova, G.F., Novoselova, E.I. and Khaziev, F.Kh. (2002a) Fermenty sernogo obmena v neftezagryaznennykh pochvakh (Enzymes of the sulphur metabolism in petroleum-polluted soils). *Pochvovedenie*, No. 4, 474-480.
- Kireeva, N.A., Novoselova, E.I. and Onegova, T.S. (2002b) Aktivnost' katalazy i degidrogenazy v pochvakh, zagryaznennykh neft'yu i nefteproduktami (Activity of catalase and dehydrogenase in soils, polluted with petroleum and petroleum products). *Agrokhimiya*, No. 8, 66-74.
- Kiss, S., Drăgan-Bularda, M. and Paşa, D. (2002) Recent environmental enzymological investigations. *Contribuții Botanice* (Cluj-Napoca), **37**, 299-304.
- Kiyamova, S.N. (1999) Biologicheskaya aktivnost' neftezagryaznennykh pochv v protsesse bioremediatsii (Biological activity of petroleum-polluted soils in the process of bioremediation). *Referativnyi Zhurnal, Pochvovedenie i Agrokhimiya*, No. 11, 97 (deposited).

- López-Hernández, D., Hernández, C.L., Briceño, G. and Pulido, R. (1999) Effect of oil spills on microbial and enzymatic soil parameters of savanna soils: long-term and incubation experiments. *Abstracts, International Conference "Enzymes in the Environment: Activity, Ecology and Applications"* (Granada, 1999), p. 128.
- Malachowska-Jutsz, A., Mrozowska, J., Kozielska, M. and Miksch, K. (1997) Aktywność enzymatyczna w glebie skażonej związkami ropopochodnymi w procesie jej detoksykacji (Enzymatic activity in soil contaminated by petroleum derivatives during the process of its detoxification). *Biotechnologia*, No. 1, 79-91
- Margesin, R. (2000) Potential of cold-adapted microorganisms for bioremediation of oil-polluted alpine soils. *International Biodeterioration & Biodegradation*, **46**, 3-10.
- Margesin, R. and Schinner, F. (1997) Bioremediation of diesel-oil-contaminated alpine soils at low temperatures. *Applied Microbiology and Biotechnology*, **47**, 462-468.
- Margesin, R. and Schinner, F. (1999a) Biological decontamination of oil spills in cold environments. *Journal of Chemical Technology and Biotechnology*, **74**, 381-389.
- Margesin, R. and Schinner, R. (1999b) A feasibility study for the *in situ* remediation of a former tank farm. *World Journal of Microbiology & Biotechnology*, **15**, 615-622.
- Margesin, R., Zimmerbauer, A. and Schinner, F. (1999) Soil lipase activity – a useful indicator of oil biodegradation. *Biotechnology Biotechniques*, **13**, 859-863.
- Margesin, R., Walder, G. and Schinner, F. (2000a) The impact of hydrocarbon remediation (diesel oil and polycyclic aromatic hydrocarbons) on enzyme activities and microbial properties of soil. *Acta Biotechnologica*, **20** (3-4), 313-333.
- Margesin, R., Zimmerbauer, A. and Schinner, F. (2000b) Monitoring of bioremediation by soil biological activities. *Chemosphere*, **40**, 339-346.
- Margesin, R., Feller, G., Hämerle, M., Stegner, U. and Schinner, F. (2002) A colorimetric method for the determination of lipase activity in soil. *Biotechnology Letters*, **24**, 27-33.
- Markarova, M.Yu. and Kaneva, Yu.S. (2000) Vzaimosvyaz' mezhdu fermentativnoi i mikrobiologicheskoi aktivnost'yu neftezagryaznenykh pochv Severa (Interrelation between enzymatic and microbiological activity of petroleum-polluted soils of the North). *Tezisy Dokladov, XIV.Komi Respublikanskaya Molodezhnaya Nauchnaya Konferentsiya. Tom 2. Aktual'nye Problemy Biologii i Ekologii* (Syktyvkar, 2000), pp. 132-133.
- Megharaj, M., Singleton, I., McClure, N.C. and Naidu, R. (2000) Influence of petroleum hydrocarbon contamination on microalgae and microbial activities in a long-term contaminated soil. *Archives of Environmental Contamination and Toxicology*, **38**, 439-445.
- Miethe, D., Riis, V. and Babel, W. (1994) The relationship between the microbial activity of the autochthonous microorganisms of pristine and contaminated soils and their potential for the degradation of mineral oil hydrocarbons. *Acta Biotechnologica*, **14** (2), 131-140.
- Panichkina, I.V. (1997) Sposoby rekul'tivatsii neftezagryaznenykh pochv (Methods for the recultivation of petroleum-polluted soils). *Tezisy Dokladov Nauchnoi Konferentsii Molodykh Uchenykh Goroda Pushchino* (Pushchino, 1997), pp. 182-183.

- Peña, W., Leirós, M.C., Trasar-Cepeda, C. and Gil-Sotres, F. (1999) Soil pollution by diesel: modification of soil microbiological and biochemical properties. *Abstracts, International Conference “Enzymes in the Environment: Activity, Ecology and Applications”* (Granada, 1999), p. 131.
- Ponomareva, L.V., Krunchak, V.G., Torgovanova, V.A., Tsvetkova, N.P. and Osipov, A.I. (1998) Bioremediatsiya neftezagryaznennoi pochvy s ispol'zovaniem biopreparata “Bioset” i peroksiда kal'tsiya (Bioremediation of oil-polluted soil with the use of the biopreparation “Bioset” and calcium peroxide). *Biotehnologiya*, No. 1, 79-84.
- Popa, A. (1999) Activitatea dehidrogenazică în sol ca test ecotoxicologic pentru poluanți anorganici și organici (Dehydrogenase activity in soil as ecotoxicological test for inorganic and organic pollutants). *Studia Universitatis Babeş-Bolyai, Biologia*, **44** (1-2), 169-178.
- Popa, A. (2000) Inducția enzimatică în sol ca test ecotoxicologic pentru poluanți anorganici și organici (Enzymatic induction in soil as ecotoxicological test for inorganic and organic pollutants). *Studia Universitatis Babeş-Bolyai, Biologia*, **45** (1), 129-138.
- Samosova, S.M., Fil'chenkova, V.I., Musina, G.Kh., Kiprova, R.R. and Gubaidullina, T.S. (1979) Izyskanie putei stimulyatsii biodegradatsii nefti v pochve (Search of the ways for stimulation of the biodegradation of petroleum in soil). *Tezisy Dokladov Konferentsii “Mikrobiologicheskie Metody Bor'by s Zagryazneniem Okruzhayushchey Sredy”* (Pushchino, 1979), pp. 8-10.
- Sattarov, U.G., Samosova, S.M., Khramov, I.T., Artem'eva, T.I., Minibaev, V.G., Tribrat, T.G. and Gilyazov, M.Yu. (1982) Nekotorye rezul'taty kompleksnykh issledovanii po rekul'tivatsii zemel' v ob'edinenii “Tatneft” (Some results of the complex researches for recultivation of earths in the association “Tatneft”). *Neftyanaya Promyshlennost'. Seriya Korroziya i Zashchita Neftegazovoi Promyshlennosti* (Moscow), No. 3, 29-31.
- Schinner, F., Zimmerbauer, A. and Margesin, R. (1999) Soil enzymes – useful tools to monitor hydrocarbon biodegradation. *Abstracts, International Conference “Enzymes in the Environment: Activity, Ecology and Applications”* (Granada, 1999), p. 132.
- Tiberg, E., Adam, R., Fischer, W., Knoblauch, H., Genz, U. and Kaun, E. (1993) Laboruntersuchungen zur mikrobiellen Abbaubarkeit verschiedener mineralöl-kontaminiert Böden (Laboratory investigations on the microbial degradability of different mineral oil-contaminated soils). *Fachgespräche Umweltschutz, “Bewertung und Sanierung mineralöl-kontaminiert Böden”* (Heidelberg, Leipzig, 1992) (G. Kreysa, J. Wiesner and J. Klein, Eds.), pp. 442-447. DECHEMA, Frankfurt am Main.
- Trasar-Cepeda, C., Leirós, M.C., Seoane, S. and Gil-Sotres, F. (1999; 2000) Limitations of soil enzymes as indicators of soil pollution. *Abstracts, International Conference “Enzymes in the Environment: Activity, Ecology and Applications”* (Granada, 1999), p. 120; *Soil Biology & Biochemistry*, **32**, 1867-1875.
- van der Waarde, J.J., Dijkhuis, E.J., Henssen, M.J.C. and Keuning, S. (1995a) Enzyme assays as indicators for biodegradation. In *Contaminated Soil '95* (W.J. van den Brink, R. Bosman and F. Arendt, Eds.), pp. 1377-1378. Kluwer, Dordrecht.

ADDITIONAL BIBLIOGRAPHY ON ENZYMOLOGY OF DISTURBED SOILS

- van der Waarde, J.J., Dijkhuis, E.J., Heijs, S.K., Henssen, M.J.C. and Keuning, S. (1995b) Monitoring bioventing of soil contaminated with mineral oil. In *In Situ Aeration: Air Sparging, Bioventing, and Related Remediation Processes* (R.E. Hinchee, R.N. Miller and P.C. Johnson, Eds.), pp. 409-417. Battelle Press, Columbus, Ohio.
- Wilke, B.-M. (1997) Effects of non-pesticide organic pollutants on soil microbial activity. *Advances in GeoEcology* (Reiskirchen), **30**, 117-132.
- Wilke, B.-M. and Koch, C. (1998) Combination effects of selected PAHs, PCPs, and heavy metals on bacteria and dehydrogenase activity of sewage farm soils. *16^e Congrès Mondial de Science du Sol* (Montpellier, 1998), *Résumés*, **2**, 685.
- Wyszkowska, J. and Kucharski, J. (2000) Biochemical properties of soil contaminated by petrol. *Polish Journal of Environmental Studies*, **9** (6), 479-485.
- Yamaletdinova, G.F. (2000) Izuchenie toksicheskogo deistviya nefti na biokhimicheskie protsessy v pochve (Study of the toxic effect of petroleum on the biochemical processes in soil). *Sbornik Tezisov, IV. Mezhdunarodnaya Ekologicheskaya Konferentsiya Studentov i Molodykh Uchenykh "Rol' Nauki i Obrazovaniya dlya Ustoichivogo Razvitiya na Poroge Tret'ego Tysyacheletiya"* (Moscow, 2000), **1**, 112-113.
- Yamaletdinova, G.F. (2001) Izuchenie vozmozhnosti ispol'zovaniya aktivnosti oksidoreдuktaz dlya indikatsii neftezagryazneniykh pochv (Study of the possibility of using activity of oxidoreductases for indication of petroleum-contaminated soils). *Sbornik Tezisov, V. Pushchinskaya Konferentsiya Molodykh Uchenykh "Biologiya – Nauka 21-go Veka"* (Pushchino, 2001), p. 320.

PART II. ENZYMOLOGY OF SOILS AFFECTED BY INDUSTRIAL EMISSIONS

ADDENDA. Nuclear fuel waste disposal. Urban soils. Roadside soils

- Antonenko, A.M. (1997) Allyuvial'nye tekhnogenno-preobrazovannye pochvy oz. Baikal: mikroflora i biokhimicheskie protsessy (The anthropogenically transformed alluvial soils of the Baikal lake: microflora and biochemical processes). *Tezisy Dokladov Mezhdunarodnoi Konferentsii "Problemy Antropogenennogo Pochvoobrazovaniya"* (Moscow, 1997), **1**, 260-262.
- Arcak, S., Haktanir, K. and Karaka, A. (1996) The effect of vehicle-generated heavy metals in the highway vicinity on urease enzyme activity. *Turkish Journal of Agriculture and Forestry*, **20** (Suppl.), 101-107 (in Turkish). *Chemical Abstracts*, 1996, **125**, 320049 f.
- Bewley, R.J.F. and Parkinson, D. (1984) Effects of sulphur dioxide pollution on forest soil microorganisms. *Canadian Journal of Microbiology*, **30**, 179-185.
- Bewley, R.J.F. and Parkinson, D. (1986) Sensitivity of certain soil microbial processes to acid deposition. *Pedobiologia*, **29**, 73-84.
- Bezuglova, O.S., Gorbov, S.N., Evseeva, N.V. and Urkova, E.G. (1999) Changing of the biological activity of urban soils in the case of sealing-up. *Abstracts, International Conference "Enzymes in the Environment: Activity, Ecology and Applications"* (Granada, 1999), p. 122.

- Bezuglova, O., Gorbov, S., Morozov, I. and Privalenko, V. (2000) To a question about the urban soil mapping (a.e. of Rostov-na-Donu). *Proceedings of the First International Conference on Soils of Urban, Industrial, Traffic and Mining Areas* (Essen, 2000) (W. Burghardt and C. Dornau, Eds.), **1**, 67-70.
- Bielińska, E.J., Domżał, H., Kopińska, A. and Plecha, R. (1998) Aktywność enzymatyczna gleby zdegradowanych siedlisk leśnych w obszarze oddziaływanego zakładów azotowych "Pulawy" S.A. (Enzymatic activity of soil in degraded forest stands in the vicinity of the "Pulawy" nitrogen fertiliser factory). *Chemia i Inżynieria Ekologiczna*, **5** (11), 953-961.
- Broll, G. and Keplin, B. (1995) Bodenökologische Untersuchungen auf städtischen Grünflächen (Soil ecological investigations on urban lawns). *Verhandlungen der Gesellschaft für Ökologie*, **24**, 385-389.
- Cassidy, D.P. and Irvine, R.L. (1999) Use of calcium peroxide to provide oxygen for contaminant biodegradation in a saturated soil. *Journal of Hazardous Materials*, **69**, 25-39.
- Dahm, H., Li, C.-Y. and Januszek, K. (1997) Development of microorganisms and oxidation of some organic compounds in soil polluted with heavy metals. *Polish Journal of Soil Science*, **30** (2), 55-63.
- Dick, R.P. (1994) Soil enzyme activities as indicators of soil quality. In *Defining Soil Quality for a Sustainable Environment* (J.W. Doran, D.C. Coleman, D.F. Bezdicek and B.A. Stewart, Eds.), pp. 107-124. Soil Science Society of America, Madison, Wisconsin.
- Dick, R.P. (1997) Soil enzyme activities as integrated indicators of soil health. In *Biological Indicators of Soil Health* (C.E. Pankhurst, B.M. Doube and V.V.S.R. Gupta, Eds.), pp. 121-156. CAB International, Wallingford, U.K.
- Dick, W.A. and Tabatabai, M.A. (1992) Significance and potential uses of soil enzymes. In *Soil Microbial Ecology. Applications in Agriculture and Environmental Management* (F.B. Jr. Metting, Ed.), pp. 95-127. Dekker, New York.
- Dick, R.P., Breakwell, D.P. and Turco, R.F. (1996) Soil enzyme activities and biodiversity measurements as integrative microbiological indicators. In *Methods for Assessing Soil Quality* (J.W. Doran and A.J. Jones, Eds.), pp. 247-271. Soil Science Society of America, Madison, Wisconsin.
- Dungan, R.S. and Frankenberger, W.T., Jr. (2002) Enzyme-mediated transformations of heavy metals/metalloids. Applications in bioremediation. In *Enzymes in the Environment. Activity, Ecology, and Applications* (R.G. Burns and R.P. Dick, Eds.), pp. 539-565. Dekker, New York.
- Ellis, R.J., Neish, B., Trett, M.W., Best, J.G., Weightman, A.J., Morgan, P. and Fry, J.C. (2001) Comparison of microbial and meiofaunal community analyses for determining impact of heavy metal contamination. *Journal of Microbiological Methods*, **45**, 171-185.
- Evdokimova, G.A. and Mozgova, N.P. (2000) Vliyanie vybrosov predpriyatiya tsvetnoi metallurgii na pochvy v usloviyakh model'nogo opyta. *Pochvovedenie*, No.5, 630-638; The impact of emissions from the nonferrous metallurgical plant on soil in a model experiment. *Eurasian Soil Science*, No. 5, 552-559.

ADDITIONAL BIBLIOGRAPHY ON ENZYMOLOGY OF DISTURBED SOILS

- Fedorov, A.S. (1997) Tekhnogennoe izmenenie svoistv pochv pod vozdeistviem otkhodov predpriyatiis tsvetnoi metallurgii (Technogenic changes in properties of soils under the action of wastes of the nonferrous metallurgical plants). *Tezisy Dokladov Mezhdunarodnoi Konferentsii "Problemy Antropogennogo Pochvoobrazovaniya"* (Moscow, 1997), **3**, 183-185.
- Freedman, B. and Hutchinson, T.C. (1980) Smelter pollution near Sudbury, Ontario, Canada, and effects on forest litter decomposition. *Global Environmental Change* (New York), **4**, 395-434.
- Galiulin, R.V., Bashkin, V.N., Galiulina, R.R., Kucharski, R., Malkowski, E. and Marchwinska, E. (1998) Vliyanie effektorov fitoekstraktsii na fermentativnuyu aktivnost' pochvy, zagryaznennoi tyazhelyimi metallami (Effect of the effectors of phytoextraction on the enzymatic activity of soil, polluted with heavy metals). *Agrokhimiya*, No.7, 77-86.
- Gianfreda, L. and Bollag, J.-M. (2002) Isolated enzymes for the transformation and detoxification of organic pollutants. In *Enzymes in the Environment. Activity, Ecology, and Applications* (R.G.. Burns and R.P. Dick, Eds.), pp. 495-538. Dekker, New York.
- Grishko, V.N. (2000) Biologicheskaya aktivnost' pochv, ispytyvayushchikh deistvie ftosoderzhashchikh promyshlennyk vybrosov (Biological activity of soils, exposed to the effect of fluorine-containing industrial emissions). *Dop. Nats. AN Ukrainsi*, No. 6, 192-198. *Referativnyi Zhurnal, Pochvovedenie i Agrokhimiya*, 2001, No. 8, 115.
- Ivleva, S.N. and Shimko, N.A. (1999) Proteoliticheskaya i fosfataznaya aktivnost' dernovo-podzolistoi suglinistoi pochvy v zone deistviya Soligorskogo kaliinogo kombinata (Proteolytic and phosphatase activity of the loamy soddy-podzolic soil in the zone affected by the Soligorsk potash plant). *Pochvovedenie*, No. 7, 860-865.
- Januszek, K. (1992) Aktywnosc enzymatyczna butwin wybranych gleb swierczyn Sudetow Zachodnich i Tatr (Enzymatic activity of raw humus in selected soils of spruce stands in the Western Sudeten and Tatra). *Zeszyty Naukowe Akademii Rolniczej w Krakowie, Leśnictwo*, No. 22, 81-101.
- Januszek, K. (1994) Aktywnosc fosfataz w glebach leśnych Sudetów Zachodnich oznaka skażenia środowiska? (Activity of phosphatases in forest soils of the Western Sudeten as indicator of the environmental pollution?). *Prace Instytutu Badawczego Leśnictwa, Seria B* (Warsaw), No. 21, 163-172.
- Januszek, K. (1999) Aktywnosc enzymatyczna wybranych gleb leśnych Polski poludniowej w świetle badań polowych i laboratoryjnych (Enzymatic activity in selected forest soils of Southern Poland in the light of field and laboratory investigations). *Zeszyty Naukowe Akademii Rolniczej w Krakowie*, No. 250, 1-132.
- Kaigorodova, S. Yu. (1997) Osobennosti transformatsii yuzhno-taezhnykh pochv v zone vozdeistviya Sredne-Ural'skogo medeplavil'nogo kombinata (Peculiarities of the transformation of southern taiga soils in the zone affected by the copper smelter plant in Sredne-Uralsk). *Tezisy Dokladov Mezhdunarodnoi Konferentsii "Problemy Antropogennogo Pochvoobrazovaniya"* (Moscow, 1997), **3**, 144-147.

- Kaigorodova, S.Yu. and Vorobeichik, E.L. (1996a) K voprosy o diagnostike khimicheski zagryaznennykh pochv (To the question of diagnosing chemically polluted soils). *Tezisy Dokladov 2. S'ezda Obshchestva Pochvovedov* (Sankt-Petersburg, 1996), **2**, 60-61.
- Kaigorodova, S.Yu. and Vorobeichik, E.L. (1996b) Transformatsiya nekotorykh svoistv serykh lesnykh pochv pod deistviem vybrosov medeplavil'nogo kombinata (Transformation of some properties of grey forest soils under the effect of emissions from copper smelter plant). *Ekologiya*, No. 3, 187-193.
- Kelly, J.J. and Tate, R.L. III (1998) Effects of heavy metal contamination and remediation on soil microbial communities in the vicinity of a zinc smelter. *Journal of Environmental Quality*, **27**, 609-617.
- Kelly, J.J., Häggblom, M. and Tate, R.L. III (1999) Changes in soil microbial communities over time resulting from one time application of zinc: a laboratory microcosm study. *Soil Biology & Biochemistry*, **31**, 1455-1465.
- Kiss, S., Drăgan-Bularda, M. and Pașca, D. (2002) Enzimologia mediului înconjurător la începutul secolului al XXI-lea (Environmental enzymology at the beginning of the XXIst century). *Program și Rezumate, Simpozionul "Mediu - Cercetare, Protecție și Gestiune"* (Universitatea "Babeș-Bolyai", 2002), p. 59.
- Kolosváry, I. (1998) Data concerning the possibility of using the abundance of the *Azotobacter* cells as a bioindicator of soil pollution. *Studia Universitatis Babeș-Bolyai, Biologia*, **43** (1-2), 137-141.
- Kovalenko, L.A., Koroleva, L.I., Kokovkina, T.F., Simonova, L.I. and Shebalova, N.M. (1997) Bioindikatsiya sostoyaniya lesnykh pochv metodom fermentativnogo analiza v zone promyshlennogo zagryazneniya ftor- i sul'fatsoderzhashchimi pollyutantami (Bioindication of the state of forest soils by the method of enzyme analysis in the zone of industrial contamination with fluorine- and sulphate-containing pollutants). *Materialy Nauchno-Prakticheskoi Konferentsii "Lesa Bashkortostana: Sovremennoe Sostoyanie i Perspektivy"* (Ufa, 1997), pp. 126-128.
- Kuperman, R.G. (1996) A hierarchical approach to ecological assessment of contaminated soils at Aberdeen Proving Ground, USA. In *Bioindicator Systems for Soil Pollution* (N.M. van Straalen and D.A. Krivolutsky, Eds.), pp. 197-212. Kluwer, Dordrecht.
- Kuperman, R.G. and Edwards, C.A. (1997) Effects of acidic deposition on soil invertebrates and microorganisms. *Reviews of Environmental Contamination and Toxicology*, **148**, 35-138.
- Langer, U. (1999) Langzeiteinfluß von Emissionen aus Phosphat-Düngemittelproduktion auf bodenmikrobielle Aktivitäten in Queckenrasen auf Muschelkalk (Long-term influence of emissions from phosphate fertiliser production on soil microbial activities in couch-grass stands on shelly limestone). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, **91**, 662-665.
- Langer, U. and Günther, T. (2001) Effects of alkaline dust deposits from phosphate fertilizer production on microbial biomass and enzyme activities in grassland soils. *Environmental Pollution*, **112**, 321-327.

- Langer, U., Günther, T. and Fritsche, W. (1999) Long-term influence of alkaline dust deposits of phosphate fertilizer production on soil microbial activities. *Abstracts, International Conference “Enzymes in the Environment: Activity, Ecology and Applications”* (Granada, 1999), p. 128.
- Lattaud, C., Mora, P., Dilumbu, M., Locati, S. and Rouland, C. (1998) Impact of heavy metals on phosphatase activities in the industrial fallow soil (Nord-Pas de Calais, France). *16^e Congrès Mondial de Science du Sol* (Montpellier, 1998), *Symposium No. 37, Poster, 11 pages + Résumés*, **2**, 682.
- Machulla, G. (2000) Microbial biomass and activity in soils developed on man-made substrates. *Proceedings of the First International Conference on Soils of Urban, Industrial, Traffic and Mining Areas* (Essen, 2000) (W. Burghardt and C. Dornau, Eds.), **3**, 587-592.
- Markkola, A.M. and Ohtonen, R. (1988) Mycorrhizal fungi and biological activity of humus layer in polluted pine forests in the surroundings of Oulu. *Karstenia*, **28**, 45-47.
- Maryskevych, O. and Shpakivska, I. (2000) Soil enzymes and soil microbial activity as indicators of soil quality in the urban soils. *Proceedings of the First International Conference on Soils of Urban, Industrial, Traffic and Mining Areas* (Essen, 2000) (W. Burghardt and C. Dornau, Eds.), **3**, 869-873.
- McGrath, R. and Singleton, I. (2000) Pentachlorophenol transformation in soil: a toxicological assessment. *Soil Biology & Biochemistry*, **32**, 1311-1314.
- Megharaj, M., Singleton, I. and McClure, N.C. (1998) Effect of pentachlorophenol pollution towards microalgae and microbial activities in soil from a former timber processing facility. *Bulletin of Environmental Contamination and Toxicology*, **61**, 108-115.
- Mukatanov, A.Kh. and Shigapov, Z.Kh. (1997) Pochvennaya otsenka sosnyakovych ekosistem na Karabashskom monitoringovom statsionare (Soil evaluation of the pine forest ecosystems on the monitoring station in Karabash). *Problemy Ekologii Yuzhnogo Urala* (Chelyabinsk), No. 3, 44-46.
- Müller, A.K., Westergaard, K., Christensen, S. and Sørensen, S.J. (2001) The effect of long-term mercury pollution on the soil microbial community. *FEMS Microbiology Ecology*, **36**, 11-19.
- Nannipieri, P. (1994) The potential use of soil enzymes as indicators of productivity, sustainability and pollution. In *Soil Biota. Management in Sustainable Farming Systems* (C.E. Pankhurst, B.M. Doube, V.V.S.R. Gupta and P.R. Grace, Eds.), pp. 238-244. CSIRO, Melbourne.
- Naprasnikova, E.V. (1998) Osnovnye podkhody, metody i rezul'taty izucheniya biogenicheskikh svoistv pochv Sharypovskogo promuzla (zona KATEKa) (Basic approaches, methods and results of the study of the biogenic properties of soils in the Sharypovo industrial junction (zone of KAFEC)). In *Ekologicheskie Problemy Urbanizirovannykh Territorii*, pp. 105-113. Institut Geografii, Irkutsk.
- Naprasnikova, E.V. (2001) Biodiagnostika pochv antropogenicheskikh geosistem (Biodiagnosis of soils of anthropogenic geosystems). *Geografiya i Prirodnye Resursy*, No. 1, 55-59.

- Naprasnikova, E.V. (2002) Pochvenno-ekologicheskie modifikatsii antropogennykh geosistem (Soil ecological modifications of anthropogenic geosystems). In *Nauchnye Chteniya Pamyati Akademika Viktora Borisovicha Sochavy* (V.A. Snytko and Yu.M. Semenov, Eds.), pp. 87-91. Institut Geografii, Irkutsk.
- Naprasnikova, E.V. and Makarova, A.P. (1999) Sanitarno-mikrobiologicheskaya i biokhimicheskaya kharakteristika pochv v usloviyakh urbanizatsii (Sanitary-microbiological and biochemical characteristics of soils under conditions of urbanisation). *Gigiena i Sanitariya*, No. 3, 15-17.
- Naprasnikova, E.V. and Snytko, V.A. (1999) Estestvennaya i antropogennaya izmenchivost' svoistv pochv Pribaikal'ya: biokhimicheskii aspekt (Natural and anthropogenic variability of soil properties in the area located along the Baikal shore: biochemical aspect). In *Sovremennoe Prirodopol'zovanie i Tekhnogennye Protsessy* (V.A. Snytko and T. Szczypek, Eds.), pp. 104-107. Institute of Geography, Irkutsk and University of Silesia, Sosnowiec.
- Naprasnikova, E.V. and Snytko, V.A. (2002) Biochemical activity of some urban soils in Eastern Siberia. *Studia Universitatis Babeş-Bolyai, Biologia*, **47** (2), 55-58.
- Ohtonen, R. (1994) Accumulation of organic matter along a pollution gradient: application of Odum's theory of ecosystem energetics, *Microbial Ecology*, **27**, 43-55.
- Olszowska, G. (1997) Aktywność enzymatyczna gleb leśnych w rejonie oddziaływanego imisji huty cynku i ołowiu "Miasteczko Śląskie" (Enzymatic activity of forest soils in the region of the influence of emissions from the zinc and lead smelter "Miasteczko Śląskie"). *Prace Instytutu Badawczego Leśnictwa, Seria A* (Warsaw), No. 834, 107-130.
- Olszowska, G. (1998) Wpływ pyłów kadmowo-cynkowych na aktywność wybranych enzymów glebowych (Effect of cadmium-zinc dusts on the activity of selected soil enzymes). *Prace Instytutu Badawczego Leśnictwa, Seria A* (Warsaw), No. 847, 111-125.
- Olszowska, G. (1999) Wpływ nawożenia mineralnego na aktywność biochemiczną gleb leśnych skażonych pyłami kadmowo-cynkowymi (Effect of mineral fertilisation on the biochemical activity of forest soils contaminated with cadmium-zinc dusts). *Prace Instytutu Badawczego Leśnictwa, Seria A* (Warsaw), No. 881, 51-60.
- Polomski, J., Oterdoom, H. and Egli, S. (1983) Wirkung von anthropogenem Fluor auf die mikrobielle Aktivität im Boden (Effect of anthropogenic fluorine on the microbial activity in the soil). *Bulletin der Bodenkundlichen Gesellschaft, Schweiz*, **7**, 20-25.
- Rossel, D., Tarradellas, J., Bitton, G. and Morel, J.-L. (1997) Use of enzymes in soil ecotoxicology: a case for dehydrogenase and hydrolytic enzymes. In *Soil Ecotoxicology* (J. Tarradellas, G. Bitton and D. Rossel, Eds.), pp. 179-206. CRC Lewis, Boca Raton, Florida.
- Shebalova, N.M. and Babushkina, L.G. (1997) Mikrobiologicheskaya i fermentativnaya aktivnost' lesnoi podstilki i verkhnykh gorizontov pochvy v usloviyakh tekhnogennogo zagryazneniya (Microbiological and enzymatic activity of forest litter and upper horizons of soil under conditions of technogenic pollution). *Trudy Komi Nauchnogo Tsentra*, No. 155, 125-136.

ADDITIONAL BIBLIOGRAPHY ON ENZYMOLOGY OF DISTURBED SOILS

- Shebalova, N.M., Kovalenko, L.A., Koroleva, L.I. and Babushkina, L.G. (1994) Aktivnost' fermentov kak pokazatel' stepeni zagryazneniya lesnoi podstilki i pochvy (Enzyme activity as indicator of the degree of pollution of forest litter and soil). *Tezisy Dokladov Konferentsii Molodykh Uchenykh Fakul'teta Pochvovedeniya MGU "Sovremennye Problemy Pochvovedeniya i Ekologii"* (Moscow, 1994), p. 66.
- Shelukho, V.P. (1999) Izmenenie lesnykh pochv i aktivnosti tsellyulozorazrushayushchego kompleksa pri deistvii pylei tsementnogo proizvodstva (Change of the forest soils and of the activity of cellulose-decomposing complex due to the effect of dusts from the cement production). *Materialy Mezhdunarodnoi Nauchnoi Konferentsii Molodykh Uchenykh "Les, Nauka, Molodezh"* (Gomel, 1999), **2**, 106-108. *Referativnyi Zhurnal, Pochvovedenie i Agrokhimiya*, 2001, No.2, 166.
- Sheppard, M.I., Hawkins, J.L. and Smith, P.A. (1996) Linearity of iodine sorption and sorption capacities for seven soils. *Journal of Environmental Quality*, **25**, 1261-1267.
- Sidorenko, N.N., Lysak, L.V., Kozhevnik, P.A. and Zvyagintsev, D.G. (1998) Osobennosti mikrobiynykh kompleksov gorodskikh pochv (Peculiarities of the microbial complexes of urban soils). *Vestnik Moskovskogo Universiteta, Seriya 17. Pochvovedenie*, No. 2, 45-49.
- Sorokin, N.D. (1993) Mikrobiologicheskii monitoring lesnykh ekosistem Sibiri pri razlichnykh antropogennykh vozdeistviyah (Microbiological monitoring of the forest ecosystems of Siberia exposed to different anthropogenic effects). *Uspekhi Sovremennoi Biologii*, **113** (2), 131-140.
- Sorokin, N.D. and Gukasyan, A.B. (1990) Mikrobiologicheskaya diagnostika lesnykh pochv i filosfery v zone tekhnogennykh vybrosov (Microbiological diagnosis of forest soils and phyllosphere in the zone of technogenic emissions). In *Sovremennoe Sostoyanie Biogeotsenozov zony KATEKa*, pp. 152-161. Gidrometeoizdatel'stvo, Sankt-Petersburg.
- Tesařová, M., Šroubková, E., Záhora, J., Dušek, L., Vakula, J., Kardiovský, R. and Filip, Z. (1998) An international approach to assess soil quality by biological methods: experience from anthropogenically affected grassland soils in the Czech Republic. *16^e Congrès Mondial de Science du Sol* (Montpellier, 1998), Symposium No. 37, Poster, 6 pages + Résumés, **2**, 684.
- Tscherko, D. and Kandeler, E. (1997a) Ecotoxicological effects of fluorine deposits on microbial biomass and enzyme activity in grassland. *European Journal of Soil Science*, **48**, 329-335.
- Tscherko, D. and Kandeler, E. (1997b) Biomonitoring of soils – denitrification enzyme activity and soil microbial processes as indicators for environmental stress. *Proceedings of the 7th International Workshop on Nitrous Oxide Emissions* (Cologne, Germany, 1997) (K.H. Becker and P. Wiesen, Eds.), *Bergische Universität, Gesamthochschule Wuppertal, Physikalische Chemie, Bericht*, No. 41, 373-381.
- Tscherko, D. and Kandeler, E. (1999) Classification and monitoring of soil microbial biomass, N-mineralization and enzyme activities to indicate environmental changes. *Die Bodenkultur*, **50** (4), 215-226.

- Udachin, V.N. (1995) O vozmozhnosti otsenki stepeni khimicheskogo zagryazneniya pochv po aktivnosti pochvennykh fermentov (On the possibility for evaluation of the degree of chemical pollution of soils by activity of soil enzymes). *Problemy Ekologii Yuzhnogo Urala* (Chelyabinsk), No. 4, 16-17.
- van Beelen, P. and Doelman, P. (1997) Significance and application of microbial toxicity tests in assessing ecotoxicological risks of contaminants in soil and sediment. *Chemosphere*, **34**, 455-499.
- Zwoliński, J., Olszowska, G. and Zwolińska, B. (1987) Soil biological activity as an indicator of industrial pressure on the forest environment (microbiological and biochemical activity). *Acta Agraria et Silvestria, Series Silvestris*, **26**, 25-44.

PART III. ENZYMOLOGY OF TECHNOGENIC SOILS

- Belova, E.V. (1999) Fermentativnaya aktivnost' vyrabotannykh torfyanykh pochv (Enzymatic activity of exploited peat soils). *Trudy Regional'noi Nauchno-Prakticheskoi Konferentsii Studentov, Aspirantov i Molodykh Uchenykh* (Tomsk, 1998), **5**, 55-56. *Referativnyi Zhurnal, Pochvovedenie i Agrokhimiya*, 2001, No. 7, 199.
- Belova, E.V. (2000) Katalaznaya aktivnost' rekul'tivirovannykh torfyanykh pochv (Catalase activity of recultivated peat soils). *Problemy Regional'noi Ekologii* (Tomsk), No. 6, 60-61. *Referativnyi Zhurnal, Pochvovedenie i Agrokhimiya*, 2001, No. 6, 134.
- Blaga, G., Vulcan, R., Clapa, D. and Lechințan, T. (1997) Recherches sur l'activité enzymatique dans les haldes de stérile (protosol anthropique) résultant de l'exploitation superficielle minière en Transylvanie (Researches on the enzymatic activity of spoils (anthropic protosol) resulting from surface mining in Transylvania). *Lucrări Științifice, Universitatea Agronomică și de Medicină Veterinară, Iași, Seria Agronomie*, **40**, 112-115.
- Blaga, G., Vesa, S., Spînu, M. and Paulette, L. (1998) Parametrii microbiologici și enzimologici ai protosolului antropic de la Căpuș, jud. Cluj, fertilizat cu nămol orășenesc (Microbiological and enzymological parameters of the anthropic protosol in Căpuș, Cluj county, fertilised with municipal sludge). *Buletinul Universității de Științe Agricole și Medicină Veterinară, Cluj-Napoca, Seria Agricultură și Horticultură*, **52**, 31-35.
- Bonmatí, M., Gascón, X., Pujolà, M. and Serra, L. (1997) Effects of sewage sludge addition on organic matter composition and biochemical parameters of a soil in a reclamation field experiment. In *The Role of Humic Substances in the Ecosystems and in Environmental Protection* (J. Drozd, S.S. Gonet, N. Senesi and J. Weber, Eds.), pp. 835-840. IHSS – Polish Society of Humic Substances, Wroclaw.
- Bonmatí, M., Jiménez, P., Garrido, A., Iñíguez, D., Lluis, R., Oyague, I. and Varela, M. (1999) Organic matter and enzyme activities as indicators for sewage sludge effect on limestone quarry reclamation. *Abstracts, International Conference "Enzymes in the Environment: Activity, Ecology and Applications"* (Granada, 1999), p. 64.

ADDITIONAL BIBLIOGRAPHY ON ENZYMOLOGY OF DISTURBED SOILS

- Drab, M. and Piontek, M. (1999) Aktywność enzymatyczna gruntów poeksplotacyjnych kruszywa budowlanego w Dobroszowie Wielkim w województwie Zielonogórskim (Enzymatic activity of ground after exploitation of gravel as building material near the village of Dobroszów Wielki in Zielona Góra). *Archiwum Ochrony Środowiska*, **25** (4), 81-88.
- Dumbeck, G., Felix-Henningsen, P., Harrach, T., Horn, R., Rademacher, D., Schneider, R., Schröder, D., Tenholtern, R. and Zepp, H. (1996) Lignite mining in the Lower Rhine Basin and landscape recultivation. *ISCO 9th Conference* (Bonn, 1996), *Excursion A*, 18 pages; *Skript zur Rekultivierung, Herbst 1999*, pp. 221-229. Abteilung Bodenkunde, Universität Trier.
- Emmerling, C. and Liebner, C. (1997) Förderung bodenmikrobiologischer Eigenschaften von Kippenböden durch Einsatz organischer Reststoffe (Stimulation of soil microbiological properties of reclaimed mine soils by addition of organic waste materials). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, **83**, 145-148.
- Emmerling, C., Embacher, A., Haubold-Rosar, M. and Schröder, D. (1996) Initierung und Förderung der mikrobiellen Biomasse und mikrobieller Aktivitäten in jungen Kippsubstraten durch organische Reststoffe (Initiation and stimulation of the microbial biomass and microbial activities in young reclaimed mine soils by organic waste materials). *VDLUFA-Schriftenreihe*, **44**, 579-582.
- Emmerling, C., Wermber, N., Schimmelpfennig, H. and Schröder, D. (1999) Bodenmikrobiologische Eigenschaften rekultivierter Böden und die Besiedlung durch Regenwürmer in den Leipziger und Lausitzer Braunkohlenrevieren in Abhängigkeit von Substrat, Nutzung und Alter (Soil microbiological properties of recultivated soils and the colonisation by earthworms in the Leipzig and Lausitz Brown Coal Districts in dependence on substrate, use and age). *Geowissenschaften & Umwelt*, pp. 67-86; *Skript zur Rekultivierung, Herbst 1999*, pp. 277-290. Abteilung Bodenkunde, Universität Trier.
- Emmerling, C., Liebner, C., Haubold-Rosar, M., Katzur, J. and Schröder, D. (2000) Impact of application of organic waste materials on microbial and enzyme activities of mine soils in the Lusatian Coal Mining Region. *Plant and Soil*, **220**, 129-138.
- García-Fernández, F., Chao, B., Trasar-Cepeda, C., Leirós, C. and Gil-Sotres, F. (1999) Effects of copper on microbiological and biochemical soils properties. *Abstracts, International Conference "Enzymes in the Environment: Activity, Ecology and Applications"* (Granada, 1999), p. 125.
- González, S., Nicolás, M.A., Navarro, M. and López, J.I. (1999) The relation between phosphatase activity and phosphorus available as recovery measure in the land waste of manganese mine. *Abstracts, International Conference "Enzymes in the Environment: Activity, Ecology and Applications"* (Granada, 1999), p. 45.
- Gornoavă, M. and Drăgan-Bularda, M. (2002) Evaluarea bioremedierii solului tehnogen din zona Rovinari-Gorj prin teste microbiologice și enzimologice (Evaluation of the bioremediation of the technogenic soil in the Rovinari-Gorj zone by microbiological and enzymological tests). *Rezumatele Lucrărilor, Sesiunea Științifică "Actualități în Biologia Vegetală"* (Universitatea "Babeș-Bolyai", 2002), pp. 36-37; *Program și Rezumate, Simpozionul "Mediu – Cercetare, Protecție și Gestiune"* (Universitatea "Babeș-Bolyai", 2002), pp. 50-51.

- Hübl, M. (2000) Mikrobiologische Eigenschaften von Kippenböden und natürlichen Böden des Westelbischen Braunkohlenreviers (Microbiological properties of the reclaimed mine soils and natural soils of the Westelbe Brown Coal District). *Trierer Bodenkundliche Schriften*, **1**, 93-98.
- Hübl, M., Wermbter, N., Schneider, R. and Schröder, D. (1999) Chemische und mikrobiologische Eigenschaften rekultivierter, ackerbaulich genutzter Böden im Raum Halle-Leipzig (Chemical and microbiological properties of the recultivated, agriculturally used soils in the Halle-Leipzig area). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, pp. 1237-1240; *Skript zur Rekultivierung, Herbst 1999*, pp. 296-297. Abteilung Bodenkunde, Universität Trier.
- Karyuchina, O.A. and Dyrin, V.A. (1999) Aktivnost' mikroflory i fermentov v ostatochnom torfe nizmennoi bolotnoi ekosistemy (BES) "Tagan" posle 13 let rekul'tivatsii (Activity of microflora and enzymes in residual peat of the "Tagan" low moor ecosystem (MES) after 13 years of recultivation). *Trudy Regional'noi Nauchno-Prakticheskoi Konferentsii Studentov, Aspirantov i Molodykh Uchenykh* (Tomsk, 1998), **5**, 54-55. *Referativnyi Zhurnal, Pochvovedenie i Agrokhimiya*, 2001, No. 7, 191.
- Kubát, J., Cerhanová, D., Mikanová, O., Hanzlíková, A. and Filip, Z. (1996) Vliv dlouhodobého hnojení a antropogenní zátěže na aktivitu půdní mikroflóry (Effects of long-term fertilisation and anthropogenic load on activity of the soil microflora). *Rostlinná Výroba*, **42**, 399-404.
- Kubát, J., Mikanová, O., Nováková, J. and Filip, Z. (1998) Assessment of soil quality by biological methods: experience from arable soils in the Czech Republic. *16^e Congrès Mondial de Science du Sol* (Montpellier, 1998), Symposium No. 37, Poster, 5 pages + Résumés, **2**, 681.
- Leirós, M.C., Trasar-Cepeda, M.C., Galán, P., Gil-Sotres, F. and Saá, A. (1994) Evolution of biochemical properties in lignite mine spoils treated with cattle slurry or inorganic fertilizer. *VII. Congresso Internazionale "L'Approccio Integrato della Moderna Biologia: Uomo, Territorio, Ambiente"* (Vieste, Italy, 1994), 1 page.
- Leirós, M.C., Trasar-Cepeda, C., García-Fernández, F. and Gil-Sotres, F. (1999) Defining the validity of a biochemical index of soil quality. *Biology and Fertility of Soils*, **30**, 140-146.
- Marschner, P., Kandeler, E. and Marschner, B. (2003) Structure and function of the soil microbial community in a long-term fertilizer experiment. *Soil Biology & Biochemistry*, **35**, 453-461.
- Maryskevych, O., Shpakivska, I. and Temchuk, V. (2000) Chemical and biological properties of technogenic dumps on surface areas in Western Ukraine. *Proceedings of the First International Conference on Soils of Urban, Industrial, Traffic and Mining Areas* (Essen, 2000) (W. Burghardt and C. Dornau, Eds.), **3**, 999.
- Máthé-Gáspár, G., Máthé, P. and Anton, A. (2003) Factors affecting the phosphomonoesterase activity in lignite mine spoils. *Proceedings COST (European Cooperation in the Field of Scientific and Technological Research) Action 831 "Biotechnology of Soil: Monitoring, Conservation and Remediation" Workshop "Managing Soil Quality"* (Budapest, 2002) (T. Szili-Kovács, Ed.), pp., 92-94. Research Institute for Soil Science and Agricultural Chemistry, Budapest.

ADDITIONAL BIBLIOGRAPHY ON ENZYMOLOGY OF DISTURBED SOILS

- Mikanová, O., Nováková, J., Hanzlíková, A., Kubát, J. and Filip, Z. (1996) Vliv dlouhodobého hnojení a antropogenní zátěže na biomasu a zastoupení některých skupin mikroorganismů (Effect of long-term fertilisation and anthropogenic load on biomass and the incidence of some groups of microorganisms). *Rostlinná Výroba*, **42**, 391-397.
- Müller, M., Harksen, E., Juhr, E., Richter, S., Grund, J., Lebentrau, B. and Häfner, B. (1993) Mikrobiologisches Untersuchungs- und Bewertungsprogramm – MUB – zur Beurteilung von Böden (Microbiological research and evaluation – MRE - programme for assessment of soils). In *Ecoinforma '92. 2. Internationale Tagung und Ausstellung über Umweltinformation und Umweltkommunikation "Bodensanierung, Bodenkontamination, Verhalten und ökotoxikologische Wirkung von Umweltchemikalien im Boden"* (K. Alef, H. Fiedler and O. Hutzinger, Eds.), pp. 139-160. Ecoinforma Press, Bayreuth.
- Muntean, V., Pașca, D., Crișan, R. and Kiss, S. (2001) Potențialul enzimatic în sterilul de la o mină de plumb și zinc supus recultivării biologice (Enzymatic potential in lead and zinc mine spoils submitted to biological recultivation). *Lucrările Conferinței Naționale pentru Știința Solului* (Suceava, 2000), *Publicațiile Societății Naționale Române pentru Știința Solului* (Bucharest), **30B**, 65-73.
- Naprasnikova, E.V. (1997) Osobennosti biokhimicheskikh protsessov formirovaniya molodykh pochv narushennykh zemel' (Peculiarities of the biochemical processes during formation of young soils on disturbed earths). *Tezisy Dokladov Mezhdunarodnoi Konferentsii "Problemy Antropogennogo Pochvoobrazovaniya"* (Moscow, 1997), **3**, 95-97.
- Pașca, D., Crișan, R., Muntean, V., Popovici, I., Kiss, S. and Drăgan-Bularda, M. (1998) Enzymological and microbiological study of the evolution of a technogenic soil submitted to biological recultivation at the lead and zinc mine in Rodna (Romania). *16^e Congrès Mondial de Science du Sol* (Montpellier, 1998), *Résumés*, **1**, 396; *Soil & Tillage Research*, **47**, 163-168.
- Pașca, D., Crișan, R., Muntean, V. and Kiss, S. (2000) Enzyme activities in biologically recultivated lead and zinc mine spoils as indicators of the recultivation efficiency. *Proceedings of the First International Conference on Soils of Urban, Industrial, Traffic and Mining Areas* (Essen, 2000) (W. Burghardt and C. Dornauf, Eds.), **3**, 1039-1044.
- Pașca, D., Kiss, S., Muntean, V. and Drăgan-Bularda, M. (2002) Crearea în condiții de laborator a unui potențial enzimatic într-un sol tehnogen (Creation of an enzymatic potential in a technogenic soil under laboratory conditions). *Program și Rezumate, Simpozionul "Mediu–Cercetare, Protecție și Gestire"* (Universitatea "Babeș-Bolyai", 2002), p. 70.
- Rumpel, C. and Kögel-Knabner, I. (2002) The role of lignite in the carbon cycle of lignite-containing mine soils: evidence from carbon mineralisation and humic acid extraction. *Organic Geochemistry*, **33**, 393-399.
- Schneider, R. (1993) Mikro- und makroskopische Untersuchungen zur Gefügeentwicklung in rekultivierten Lößböden bei unterschiedlicher Nutzung im Rheinischen Braunkohlenrevier (Micro- and macroscopical researches on the development of structure in recultivated loess soils under different use in the Rhine Brown Coal District). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, **72**, 231-234.

- Schneider, R. (1997) Bodenökologisch bedeutsame Verluste durch die Braunkohlegewinnung im Rheinischen Revier (Soil ecologically important losses through brown coal exploitation in the Rhine District). In *Braunkohle und Rekultivierung. Dokumentation der Tagung der Natur- und Umweltschutzakademie des Landes Nordrhein-Westfalen* (Jülich, 1997), pp. 20-30. *Skript zur Rekultivierung, Herbst 1999*, pp. 251-256. Abteilung Bodenkunde, Universität Trier.
- Schneider, H. and Schröder, D. (1997a) Der Einfluß ausgewählter Bodentiergruppen auf die Entwicklung von rekultivierten Böden des Braunkohleabbaus im Kölner Raum, untersucht anhand eines Laborversuchs (The influence of selected soil animal groups on the development of recultivated soils of the brown coal mining in the Köln area, as studied in a laboratory experiment). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, **83**, 195-198.
- Schneider, R. and Schröder, D. (1997b) Entwicklung und Eigenschaften von Neulandböden im Rheinischen Braunkohlenrevier und in der Wetterau (Development and properties of new reclaimed soils in the Rhine Brown Coal District and in the Wetterau). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, **84**, 57-60.
- Schneider, H., Schneider, R. and Schröder, D. (1990) Zusammenhänge zwischen bodenmikrobiologischen und -zoologischen Parametern sowie Gefügeeigenschaften in Neulandböden aus Löss und Hafenschlick (Relationships between soil microbiological and zoological parameters as well as properties of the structure of new reclaimed soils from loess and harbour mud). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, **62**, 87-90.
- Schneider, H., Schneider, R. and Schröder, D. (1991) Die mikrobiologischen Eigenschaften von Neulandböden, untersucht an gesiebtem, aggregiertem und ungestörtem Boden sowie an geschälten Aggregaten (The microbiological properties of new reclaimed soils, as studied on sieved, aggregated and undisturbed soil as well as on divided aggregates). *VDLUFA-Schriftenreihe*, **33**, 281-286.
- Schneider, R., Schneider, H. and Schröder, D. (1997) Der Einfluß acker- und waldbaulicher Nutzung auf die Eigenschaften von rekultivierten Böden (The influence of agri-and sylvicultural land use on the properties of recultivated soils). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, **83**, 199-202.
- Schneider, R., Schröder, D. and Weyers, M (1998) Entwicklung, Sanierung und Schutz des Bodengefüges von Neulandböden (Development, sanation and protection of the soil structure of new reclaimed soils). In *Braunkohlentagebau und Rekultivierung. Landschaftsökologie – Folgenutzung - Naturschutz* (W. Pflug, ed.), pp. 199-213. Springer, Berlin.
- Schneider, H., Bräunig, A. and Wünsche, M. (2000a) Eigenschaften und Entwicklung von forst- und landwirtschaftlich rekultivierten Kipp-Schluffen – Fallstudie an einer Braunkohlen -Bergbaufolgelandschaft südlich Leipzig (Properties and development of sylvi- and agriculturally recultivated mine spoil silts – Case study on a landscape following brown coal mining south of Leipzig). *Trierer Bodenkundliche Schriften*, **1**, 79-84.

ADDITIONAL BIBLIOGRAPHY ON ENZYMOLOGY OF DISTURBED SOILS

- Schneider, R., Schneider, H.C., Schröder, D. and Martinovic, L. (2000b) Eigenschaften und Entwicklung von Neulandböden aus Löß bei unterschiedlicher Nutzung im Rheinischen Braunkohlenrevier und in der Wetterau (Properties and development of new reclaimed soils from loess under different use in the Rhine Brown Coal District and in the Wetterau). In *Rekultivierung in Bergbaufolgelandschaften* (G. Broll, W. Dunger, B. Keplin and W. Topp, Eds.), pp. 239-259. Springer, Berlin.
- Schröder, D. (1988) Verwendung von Siedlungsabfällen auf rekultivierten Böden (Application of communal waste materials on recultivated soils). *Landwirtschaftliche Zeitschrift Rheinland*, No. 43, 2501-2502.
- Schröder, D. (1991) Die Wiederherstellung von Böden auf Tagebau- und Überschlickungsflächen (Reconstitution of soils on strip mine spoil and harbour mud areas). In *Wasser und Boden im Lichte Neuer Geowissenschaftlicher Forschung. Vorträge im Rahmen des Geographischen Umweltkolloquiums* (M.J. Müller, Ed.), *Flensburger Regionale Studien*, 5, 9-38.
- Schröder, D. (1995) Eigenschaften landwirtschaftlich rekultivierter Böden im Rheinischen Braunkohlenrevier und Anforderungen an die Rekultivierung (Properties of the agriculturally recultivated soils in the Rhine Brown Coal District and requirements for the recultivation). *Workshop Rekultivierung 1991/92. Richtlinien für die Rekultivierung der Neulandböden in den Braunkohleplanungsgebieten Ostdeutschlands* (Senftenberg, 1991), pp. 21-26. Lausitzer Braunkohle AG, Senftenberg. *Skript zur Rekultivierung, Herbst 1999*, pp. 215-220. Abteilung Bodenkunde, Universität Trier.
- Schröder, D. (2000) Problematik der Bodenschätzung auf Kippen (Problems of soil assessment on strip mine spoils). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, 93, 282-285.
- Schröder, D. und Ganitta, U. (1989) Kompost und Bodenfruchtbarkeit (Compost and soil fertility). In *Kompost und Landwirtschaft*, pp. 20-36. Landwirtschaftliche Untersuchungs- und Forschungsanstalt, Bonn.
- Schröder, D., Schneider, R. and Haubold-Rosar, M. (1994) Rekultivierung und rekultivierte Böden (Recultivation and recultivated soils). In *Planung, Gestaltung und Schutz der Umwelt* (H. Große, G. Lehmann and M. Mittag, Eds.), 5 (Ergänzungen), 228-265. IRB Verlag, Stuttgart. *Skript zur Rekultivierung, Herbst 1999*, pp. 170-189. Abteilung Bodenkunde, Universität Trier.
- Scullion, J. and Malik, A. (2000) Earthworm activity affecting organic matter, aggregation and microbial activity in soils restored after opencast mining for coal. *Soil Biology & Biochemistry*, 32, 119-126.
- Shugalei, L.S. (1997) Pervichnoe pochvoobrazovanie na otvalakh vskryshnykh porod pod kul'turoi sosny (Primary soil formation on strip mine spoils under pine culture). *Pochvovedenie*, No. 2, 247-253.
- Shugalei, L.S., Yashikhin, G.I. and Nefodina, N.L. (1984) Formirovanie lesnykh biogeotsenozov na rekul'tivirovannykh zemlyakh KATEKa (Formation of forest biogeocoenoses on the recultivated earths of KAFEC). *Geografiya i Prirodnye Resursy*, No. 1, 30-36.
- Shugalei, L.S., Vedrova, E.F., Kuz'mina, G.P. and Yashikhin, G.I. (1996) Vliyanie udobrenii na formirovanie iskusstvennykh lesnykh biogeotsenozov na otvalakh vskryshnykh porod ugol'nykh razrezov (Effect of fertilisers on the formation of artificial forest biogeocoenoses on the spoils of opencast coal mines). *Agrokhimiya*, No. 5, 65-72.

- Šíša, R., Sixta, J., Růžek, L. and Štokánová, G. (2000) Biologická aktivita antropogenních půd (Biological activity of anthropogenic soils). *Rostlinná Výroba*, **46**, 55-61.
- Speir, T.W. and Ross, D.J. (2002) Hydrolytic enzyme activities to assess soil degradation and recovery. In *Enzymes in the Environment. Activity, Ecology, and Applications* (R.G. Burns and R.P. Dick, Eds.), pp. 407-431. Dekker, New York.
- Voříšek, K., Růžek, L. and Šíša, R. (1997) Microbiological activities in reclaimed and undisturbed soils. *Abstracts, Symposium and Status Seminar "Organic Matter Application and Element Turnover in Disturbed Terrestrial Ecosystems"* (Cottbus, 1997), pp. 13-15.
- Wermbter, N., Emmerling, C. and Schröder, D. (1997) Charakterisierung bodenbiologischer Eigenschaften von Kippenböden im Lausitzer und Leipziger Braunkohlenrevier im Hinblick auf eine standortgerechte Folgenutzung (Characterisation of the soil biological properties of the reclaimed mine soils in the Lausitz and Leipzig Brown Coal District regarding site-oriented land use after mining). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, **85**, 1607-1610.
- Wermbter, N., Sehy, U., Weber, G., Emmerling, C. and Schröder, D. (1998) Bodenbiologische Eigenschaften rekultivierter Böden im Leipziger und Lausitzer Braunkohlenrevier in Abhängigkeit von Nutzung und Substrat (Soil biological properties of the recultivated soils in the Leipzig and Lausitz Brown Coal District in dependence on use and substrate). *VDLUFA-Schriftenreihe*, **49**, 653-656.
- Wermbter, N., Sehy, U. and Weber, G. (1999a) Bodenbiologische Eigenschaften von Kippböden in Abhängigkeit von Nutzung, Bodenbearbeitung und Substrat (Soil biological properties of reclaimed mine soils in dependence on use, tillage and substrate). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, **89**, 301-304.
- Wermbter, N., Sehy, U., Emmerling, C. and Schröder, D. (1999b) Bedeutung der konservierenden Bodenbearbeitung für die bodenbiologischen Eigenschaften von rekultivierten Kippböden (Importance of the conservant, reduced tillage on the soil biological properties of recultivated reclaimed mine soils). *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft*, pp. 723-726; *Skript zur Rekultivierung, Herbst 1999*, pp. 294-295. Abteilung Bodenkunde, Universität Trier.
- Wünsche, M. and Thum, J. (1990) Bodensubstrate und Bodenentwicklung der landwirtschaftlich genutzten Flurkippe Espenhain (Sachsen) (Soil substrates and soil development of the agriculturally used spoil heaps at Espenhain (Saxony)). *Archiv für Naturschutz und Landschaftsforschung* (Berlin), **30** (4), 217-229.

IN MEMORIAM

Prof. Univ. dr. *TRAIAN CEUCA*
(18 ianuarie 1921 – 12 octombrie 2003)



Profesorul și diplopodologul Traian Ceuca s-a născut la 18 ianuarie 1921, în localitatea Salva, județul Bistrița-Năsăud, România. A urmat una dintre cele mai prestigioase școli din țară, Liceul „George Coșbuc” din Năsăud. În perioada 1940-1942 a urmat cursurile Facultății de Științe la Timișoara, unde s-a refugiat Universitatea clujeană în urma Dictatului de la Viena. După o intrerupere datorată războiului (sfârșitul acestuia găsindu-l pe frontul din Cehoslovacia), și-a definitivat studiile la Cluj. Profilul științific, moral și civic al profesorului Ceuca a fost puternic marcat de personalitatea unor prestigioși dascăli, precum Vasile

Gh. Radu, Emil Pop și Victor Pop. Din 1946, după revenirea Universității la Cluj, și-a continuat studiile întrerupte cu 4 ani, fiind angajat în același timp ca desenator la Catedra de Zoologie, condusă de prof. V. Gh. Radu. Calitățile de fin observator și remarcabil desenator l-au făcut, ca imediat după terminarea studiilor universitare, să fie reținut la catedră ca preparator, iar în decursul celor 40 de ani de carieră, a urcat toate gradele didactice până în 1974, când a devenit profesor universitar.

Studiul diplopodelor a fost început încă din anul 1949, sub îndrumarea directă a prof. dr. V.Gh. Radu, studiu care s-a transformat într-o veritabilă pasiune de o viață și a cuprins mai multe etape. Studiul faunei de diplopode din Europa – ca cercetare de bază, a inclus în primul rând inventarierea colecției „Biospeologica”, aflată la Institutul de Speologie „Emil Racoviță” din Cluj-Napoca. Începând din 1950, prin deplasările făcute pe teren, a colectat peste 10.000 de indivizi, din aproape 500 de localități din cele mai diferite regiuni și biotopuri din țară. Prima lucrare științifică a apărut în 1951, referitoare la câteva specii de *Trachisphaera (Gervaisia)*. Aprofundând studiul acestui grup faunistic, a publicat lucrări și despre reprezentanții altor grupe de diplopode din fauna țării noastre, atât forme epigee, cât și cavernicole, completând în același timp aceste lucrări, cu exceptionale desene. De asemenea, a avut colaborări cu specialiști străini, precum și studii și comunicări făcute pe material colectat din Italia, Spania, Germania, Franța, Austria, Polonia, fostă Iugoslavie, Rusia, S.U.A., Anglia, Brazilia, Tunisia, Turcia, Grecia, Bulgaria, Somalia, nordul Africii, Ungaria. Lucrarea de doctorat, cu titlul „Studiul sistematic și ecologic al diplopodelor proterandrice din fauna României” a fost susținută în 1968.

IN MEMORIAM

Profesorul Traian Ceuca a descris următorii taxoni noi pentru știință: o familie: *Hungarosomidae*, 3 genuri: *Romanosoma*, *Paraporatia*, *Napocodesmus*, 4 subgenuri: *Pseudomastuchus*, *Heteranthroleucusoma*, *Moldavobielzia*, *Spelaeoiulus*, 22 specii: *Polydesmus triacantos*, *P. brachydesmoides*, *Entomobielzia getica*, *E. varvarai*, *Napocodesmus endogeus*, *Romanosoma cavernicola*, *R. oltenica*, *R. bîrtei*, *R. odici*, *Hungarosoma inexpectata*, *Paraporatia racovitzai*, *Karpatophyllum dacicus*, *K. banaticum*, *K. carpaticum*, *Mastigophorophyllum bănărescui*, *M. aberatum*, *M. carpaticus*, *Anthroleucusoma spelaea*, *Stenophyllum semenicensis*, *Typhloius serbani*, *T. unilineatus* și *Glomeris pachitelopoda*. De asemenea, au mai fost citate 7 forme noi pentru fauna diferitelor țări: Franța, Tunisia, Bulgaria, Ungaria, Turcia. Întreaga colecție de diplopode a profesorului Ceuca are valoare de patrimoniu și a fost depusă la Muzeul Zoologic al Universității „Babeș-Bolyai” din Cluj-Napoca.

În afară de pasiunea pentru studiul diplopodelor, în calitate de zoolog, s-a preocupat și de studiul vertebratelor. Lucrările publicate în acest domeniu conțin date despre: modul de recepționare al sunetelor de către șerpi, referințe la cele trei exemplare de *Archaeopterix*, studii de anatomie comparată ale unor resturi de oase incinerate dintr-un mormânt dacic, precum și concluzii legate de numărul animalelor și al oamenilor pe care îl conținea acesta, originea și evoluția vertebratelor, originea primaterelor (cu referiri la *Tupaidae*), originea mamiferelor etc.

Cu toate vicisitudinile și îngădirile din perioada în care și-a desfășurat activitatea, profesorul Ceuca s-a impus ca un specialist de vîrf pe plan european; astfel, a participat cu lucrări la cele mai prestigioase congrese internaționale de miriapolodologie: Paris, 1968, cu lucrarea „Peut-on parler de périodomorphose en absence de certains caractères secondaires”, Bull. Mus. d’Hist. Nat. T. 41. Paris, 1969; Innsbruck, 1990, cu lucrarea „Quelques aspects sur la taxonomie, l’écologie, et zoogéografie des Diplopodes de la région Balcanique” și Paris, 1993, cu lucrarea „*Mastigophorophyllum* (Verh. 1897) et *Karpatophyllum* (Jawl. 1928), genres carpathiques (*Diplopoda Chordeumida*)”.

A publicat 73 de lucrări, dintre care 19 apărute în străinătate. Dar din păcate, lucrarea sa de referință, „Les Diplopodes de la Faune de Roumanie”, nu a putut vedea lumina tiparului în timpul vieții autorului. Actualmente se fac eforturi ca această lucrare să apară la „Danicel” –Édition de l’ Association Franco- Roumaine, France.

Profesorul Traian Ceuca a fost și un erudit dascăl, care a format numeroase generații de naturaliști biologi. În cei 40 de ani de carieră universitară a predat un număr mare de cursuri universitare: zoologia vertebratelor, anatomia comparată a vertebratelor, anatomia omului, zoogeografie, zoologia nevertebratelor, tehnica preparării și conservării materialului biologic.

A fost membru al Comitetului Internațional de Miriapolodologie (C.I.M.), cu sediul la Paris și membru al Societății Germane de Zoologie (Deutsche Zoologische Gesellschaft), cu sediul la München.

Valoarea profesorului Traian Ceuca poate fi rezumată în spusele unui fost student de-al său, actual cadru didactic al facultății noastre: „Nu-l plângeti că a murit, bucurați-vă că a trăit”.

DELIA CEUCA și IOAN COROIU

RECENZII – BOOK REVIEWS

Universität Trier. Abteilung Bodenkunde (Trier University. Soil Science Department), **Skript zur Rekultivierung** (*Scripta on Recultivation*), Überarbeitete Version, Herbst, 1999 (Revised Version, autumn 1999), VII + 301 pages.

The book of „*Scripta on Recultivation*“ is dedicated to description of the investigations performed, since the 80' by the Trier University through its Soil Science Department headed by Professor Dietmar Schröder, along the following four lines: 1. spoils from numerous strip brown coal mines, most of which are located within the Rhine and Wetterau Regions, in the Leipzig and Lusatia Basins and recultivation of these spoils into arable or grass- or forestal lands; 2. pumice stone mine spoils in the Neuwid Basin and their recultivation into agricultural soils; 3. harbour muds from the Emden harbour and from the estuary of the Ems River and transformation of these muds into agricultural soils and 4. soils heavily contaminated with carcinogenic polycyclic aromatic hydrocarbons (PAHs) on the territory of a former coking and gas works in Saarbrücken and of a former iron works in Saarbrücken-Burbach and rehabilitation of these contaminated soils.

Both laboratory and field investigations were carried out.

The spoils, soils and harbour muds were submitted to mechanical, physical, micromorphological, chemical, biochemical (including enzymological), microbiological and zoological analyses. Redoxase and hydrolase enzyme activities were measured. Dehydrogenase, invertase and alkaline phosphatase activities were preferably determined. Substrate-induced respiration, microbial biomass and cellulose decomposition were the most frequently examined microbial parameters. Among the animals special attention was always paid to the earthworms. The results were presented in details based on their statistical evaluation. The recultivation technologies applied were also de-

scribed in details. The economic value of the recultivated soils was also dealt with.

The book of „*Scripta on Recultivation*“ consists of 66 papers (62 in German and 4 in English), each in form of a reprint or a photostat copy. Most of the papers (61) were published in the 1984-1999 period and 5 papers were under press in the autumn of 1999. At present, these five papers should also be taken as published. The 66 papers, comprising 473 printed pages with 171 figures and 131 tables in the text, were elaborated by 38 investigators. As individuals 22 of them (not listed below) contributed to a single paper, whereas 16 of them (listed below) are contributors to 2 or to more papers: W. Bartel, L. Hankes, J. Katzur, M. Kraatz, H. Kutsch and G. Weber (2 papers); M. Haubold and U. Sehy (3 papers); M. Haubold-Rosar and B. Schumacher (4 papers); N. Wermber (6 papers); C. Emmerling, H. Schneider and M. Weyers (7 papers); R. Schneider (27 papers) and D. Schröder (54 papers).

The papers, according to their contents, can be grouped into five categories: original papers in journals (47 papers); original papers in proceedings of scientific meetings (5); review papers in edited works (9); dissertation abstracts (4) and excursion guide (1). Most of the original papers, namely 31 plus 7 papers appeared in two journals: *Mitteilungen der Deutschen Bodenkundlichen Gesellschaft* and *Landwirtschaftliche Zeitschrift (Rheinland)*, respectively.

The book of „*Scripta on Recultivation*“ can be considered a handbook as all the investigations described in it can serve as models for the experts and decision makers interested in initiation or development of investigations aiming at the agri- or sylvicultural recultivation of mine spoils, transformation of harbour muds into agricultural soils and rehabilitation of soils contaminated with PAHs.

MIHAIL DRĂGAN-BULARDA
and STEFAN KISS

Tibor Szilli-Kovács (Editor), **Proceedings of the Workshop on Managing Soil Quality – Using Microbial Resources** (12-13 September 2002, Budapest, Hungary), published in form of compact disk (CD) by the Research Institute for Soil Science and Agricultural Chemistry of the Hungarian Academy of Sciences, Budapest, in 2003, on IV + 94 pages including 51 figures and 17 tables in the text.

The workshop took place as a part of the COST (European Cooperation in the Field of Scientific and Technological Research) Action 831. Biotechnology of Soil: Monitoring, Conservation and Remediation.

The Proceedings comprise two reviews and 21 original papers. Both reviews deal with the activity (investigations and meetings) within the COST Action 831, in the period between October 1997 and September 2002. The first, one-page review was elaborated by the Editor and Dr. Oliver Dilly (Germany), vicechairman of the COST Action 831, whereas the second, 4-page review was written by Professor Anna Benedetti (Italy), chairwoman of the COST Action 831.

The names and countries of the 79 authors and coauthors and titles of the 21 original papers are listed below.

1. O. Dilly (Germany): *Microbial respiratory quotient as an indicator of nutritional conditions in agricultural and forest soils.*

2. D. Pașca, S. Kiss, M. Drăgan-Bularda, R. Crișan, V. Muntean (Romania): *Enzymological evaluation of therapeutic mud analogues obtained during incubation of peat mixed with a clay mineral or zeolitic tuff and inoculated with microorganisms from salt lakes.*

3. M. Pesaro, H. Bürgmann, F. Widmer, J. Zeyer (Switzerland): *Soil DNA content – a novel quantitative bulk parameter for description of microbiological soil characteristics.*

4. J.A. Pascal, C. García, T. Hernández, J.L. Moreno, M. Ros (Spain): *The future of composts as biopesticides.*

5. M.T. Pereyra de la Iglesia, J. Domenech, B. Ramos, F.J. Gutierrez Mañero (Spain): *Screening for diesel-degrading rhizobacteria associated to *Piptatherum miliaceum* rhizosphere in a polluted soil.*

6. B. Stres, J.M. Tiedje (USA), I. Mahne, G. Avguštin (Slovenia): *Diversity of nosZ gene in two agricultural soils.*

7. K. Yrjälä, P.E. Galand, H. Fritze (Finland): *Diversity of methanogens in the light of ribosomal 16S- and functional MCR-gene occurrence in Finnish oligotrophic fen.*

8. I. Ignatiadis, C. Michel, F. Battaglia-Brunet, M. Bruschi, P. Bianco, E. Lojou, C. Tran Minh (France): *Development of an enzymatic amperometric biosensor using cytochromes C₃ for the fast quantification of chromate bioavailability in the environment.*

9. A. Anton, P. Máthé, G. Füleky (Hungary): *The effect of phosphorus fertilizer on the phosphomonoesterase activity of *Capnium nannuum* L. rhizosphere.*

10. C. Trasar-Cepeda, M.C. Leirós, F. Gil-Sotres (Spain): *Sensitivity of soil biochemical properties to contamination with heavy metals.*

11. K. Svensson (Sweden): *The substrate-induced microbial biomass as a soil quality and fertility indicator in arable land.*

12. N. Riddech, S. Klammer, H. Insam (Austria), C. Mondini (Italy): *Comparison of the ability of GN and Eco microplates to discriminate microbial communities of different composting ages.*

13. J.L. Niqui-Arroyo, M. Bueno-Montes, J.J. Ortega-Calvo (Spain): *Biological and physico-chemical mechanisms involved in the bioavailability of polycyclic aromatic hydrocarbons in polluted soils.*

14. T. Szili-Kovács, K. Török, M. Halassy, R. Szabó (Hungary), D. Elhottová (Czechia): *Monitoring of restoration of sand grassland communities in abandoned fields under the manipulation of soil N availability.*

15. T. Takács, I. Vörös (Hungary): *Occurrence of indigenous AM fungi in calcareous sandy and chernozem soils and their effect on nutrient uptake of maize.*

16. A. Viterbo, O. Ramot, I. Chet (Israel): *Synergistic antifungal activity between two chitinases from the biocontrol agent T. a s - p e r e l l u m T- 203.*
17. A. Prinčič, I. Mahne (Slovenia): *Responses of nitrifying bacteria to selective pressure of the changing environments – What we know and what we need to know.*
18. S.O. Petersen, K. Henriksen, G.K. Mortensen, P.H. Krogh, K.K. Brandt, J. Sørensen, T. Madsen, J. Petersen, C. Grøn (Denmark): *Recycling of sewage sludge and household compost to arable land: Fate of organic contaminants, and impact on soil fertility.*
19. J.K. Friedel, C. Kobel, M. Stemmer, W.J. Fitz, B. Freyer, W.W. Wenzel (Austria): *Advances in estimating soil microbial biomass and enzymatic activities in the rhizosphere.*
20. A. Halbritter, T. Mogyoróssy (Hungary): *Rhizosphere effect in a peat soil demonstrated via phospholipid fatty acid (PLFA) indication.*
21. G. Máthé-Gáspár, P. Máthé, A. Anton (Hungary): *Factors affecting the phosphomonoesterase activity of lignite mine spoils.*

One can suppose based only on the titles of these 21 papers and convincingly establish based on their full texts that the *Proceedings of the Workshop on Managing Soil Quality – Using Microbial Resources* are a very useful source of information for all students and experts whose professional activity is related to soils, especially for soil microbiologists and biochemists, including enzymologists and also for decision makers in problems of soil degradation and pollution and restoration of degraded and polluted soils.

DANIELA PAŞCA,
STEFAN KISS and
MIHAIL DRĂGAN-BULARDA

I s t v á n K i s s , Az erodált talajok enzimológiája (*Enzymology of Eroded Soils*), Scientia Kiadó, Kolozsvár – Scientia Publisher, Cluj-Napoca, 2003, 160 pages with 34 tables in the text.

The book is a review of the enzymological investigations on eroded soils and also on slope and riparian soils very exposed to erosion. These investigations, originally described in 231 papers, had been carried out in 33 countries and, consequently, the book comprises 33 chapters. The order of countries was established based on the year in which the first paper on enzyme activities in eroded, slope or riparian soils has appeared. According to this year, the 33 countries (and chapters) present the following order: 1. Germany (1954); 2. Armenia (1965); 3. Romania (1966); 4. Russian Federation – Federal and Autonomous Republics: Russia (1970); Bashkiria (1979); Buryatia (1980); Udmurtia (1986); Evenkia (1999); 5. Byelorussia (1970); 6. Ukraine (1972); 7.

Kirghizia (1974); 8. Azerbaijan (1975); 9. Uzbekistan (1975); 10. France (1976); 11. Belgium (1977); 12. Austria (1978); 13. Czech Republic (1978); 14. Georgia (Gruzia) (1980); 15. Bulgaria (1980); 16. Japan (1980); 17. Spain (1982); 18. India (1983); 19. Egypt (1984); 20. Moldavian Republic (1985); 21. Lithuania (1985); 22. Poland (1986); 23. Kazakhstan (1986); 24. New Zealand (1986); 25. Slovakia (1989); 26. U.S.A. (1991); 27. Costa Rica (1991); 28. Israel (1992); 29. Denmark (1993); 30. Canada (1994); 31. Peru (1994); 32. China (1995); 33. United Kingdom (1996).

The chapters comprise at least one of the subchapters entitled as follows: Soils eroded to different extent; Prevention of soil erosion; Improvement (rehabilitation) of eroded soils; Soils of pastures overgrazed and eroded to different extent; Improvement of overgrazed and eroded pastures; Slope soils; Soils of plateaus and depressions of regions having ondulating relief; Riparian soils.

The results of the investigations reviewed in the present book made it possible to draw the general conclusion that enzyme activities are sensitive indicators of the extent of soil erosion, of the efficiency of the methods applied for prevention of soil erosion and for the improvement (rehabilitation) of the eroded soils.

The present book is a valuable source of information not only for soil biochemists,

microbiologists and technologists, but also for decision makers in problems of soil erosion.

As this book on enzymology of eroded soils is such a comprehensive and updated review which constitutes a novelty in the universal soil enzymological literature, we warmly recommend its translation into and publication in a world-wide spoken language.

KATALIN BARTÓK

D o n B r a d s h a w, Vertebrate Ecophysiology. An Introduction to Its Principles and Applications, Cambridge Univ. Press, 2003, 287 pages.

The work is a manual, an introduction to the study of vertebrate ecophysiology. Its author is a Professor at the University of Western Australia, with an outstanding activity of over 30 years in the investigation of ecophysiology. This allows him to have an extremely complex functional image about animals, which is impressively reflected in the pages of this book. The vertebrates presented in the book live in habitats with extreme environmental conditions, anthropomorphically inhospitable: deserts or cold climates (arctic, antarctic). The information is exposed as case studies which present the working techniques and methods, the results of the research and their interpretation.

Many of the case studies refer to endangered animals in situations under which their internal environment is disturbed, and therefore the author structures his book in accordance with this fact. Thus, the book contains 8 chapters: 1. *Homeostasis: a fundamental organising paradigm in ecophysiology*; 2. *Stress: the concept and the reality*; 3. *Basic methods used in ecophysiological studies*; 4. *Turnover methodology: theory and practice*; 5. *Case studies of stress: incidence and intensity*; 6. *Survival in deserts*; 7. *Torpor and hibernation in cold climates*; 8. *Marine birds and mammals*.

In the first Chapter, the factors that influence the state of internal environment are presented along with the regulating role of the endocrine and nervous system. At the very beginning of this Chapter, a special consideration for their contribution to the grounding of the concept of *homeostasis* is given to "the genius of the American physiologist" Walter Cannon, who invented the term *homeostasis* in 1929 and to the "great French physiologist" Claude Bernard, who in 1878 introduced the notion of *internal environment*. He was the first to suggest that animals regulate constantly their internal environment. As a sign of high esteem for the French physiologist, the French expression *milieu intérieur* is used throughout the book instead of the English *internal environment*. It is worth noting the author's remark that the internal environment actually varies within rather wide limits, especially in the passage phases from the juvenile to the adult state or in the reproductive period in the case of the females (after the embryo's implantation). Some pathological states, as for instance fever, presuppose a different homeostatic régime, which "resets" the upper and lower set-points.

The problem of animal stress in the natural habitat is treated mainly in Chapters 2 and 5, but the idea of stress crosses as a red line most pages of the book. Having pointed to the basic contribution of Hans Selye to the study of stress (in a nutshell: he identified three signs of the state of stress: adrenocortical enlargement, atrophy of the thymus and

lymphoid tissues and ulceration of the digestive tract, and distinguished three stages in the evolution of the state of stress: the stage of “alarm”, “adaptation” and “exhaustion”), Don Bradshaw attempts at an operational definition of stress, being concerned with the identification and measurement of its intensity: ... *the physiological resultant of demands that exceed an organism's regulatory capacities.* This resultant comes from a significant deviation of the internal environment from the state considered optimal, and which implies the surpassing of regulating responses, although they have been maximally activated.

Chapters 3 and 4 describe the most important methods used in ecophysiological research. These chapters are completed by 8 Appendices at the end of the book introducing the steps of an ecophysiological investigation. Out of the issues presented, we mention the ethical aspects that must be taken into consideration when experimenting with animals. In Chapter 6 one is acquainted with

morphological, physiological and behavioural adaptations to desert environment of frogs, lizards, tortoises, rodents, jackrabbits, island wallabies, birds and camels, while Chapter 7 presents examples of mammals which take up adaptive states which allow them to survive under conditions of food shortage of low temperature.

The last chapter offers information about the ecophysiology of albatrosses, penguins, seals and whales. It should be mentioned that the author of the book demonstrates not only a high professional competence, but also a great sensitivity towards the animal world and towards nature in general. The book is recommended for advanced undergraduate and post-graduate students, for researchers in ecology, biodiversity and conservation. It is equally useful for university professors teaching ecophysiology and even for high school teachers.

PANTE GHERGHEL